



25 **Abstract**

26 How green spaces in cities benefit urban residents depends critically on the interaction between  
27 biophysical and socio-economic factors. Urban ecosystem services are affected by both ecosystem  
28 characteristics and the social and economic attributes of city dwellers. Yet, there remains little synthesis  
29 of the interactions between ecosystem services, urban green spaces, and socio-economic factors.  
30 Articulating these linkages is key to their incorporation into ecosystem service planning and  
31 management in cities and to ensuring equitable outcomes for city inhabitants. We present a conceptual  
32 model of these linkages, describe three major interaction pathways, and explore how to operationalize  
33 the model. First, socio-economic factors shape the quantity and quality of green spaces and their ability  
34 to supply services by influencing management and planning decisions. Second, variation in socio-  
35 economic factors across a city alters people's desires and needs and thus demands for different  
36 ecosystem services. Third, socio-economic factors alter the type and amount of benefit for human  
37 wellbeing that a service provides. Integrating these concepts into green space policy, planning, and  
38 management would be a considerable improvement on 'standards-based' urban green space planning.  
39 We highlight the implications of this for facilitating tailored planning solutions to improve ecosystem  
40 service benefits across the socio-economic spectrum in cities.

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42 **Keywords**

43 Socio-economic factors; urban green spaces; urban ecosystem services; urban planning and  
44 management

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49 **1 Introduction**

50 Green spaces in urban areas, such as gardens, parks, street trees, and other 'natural' features, provide  
51 vital ecosystem services that contribute to the wellbeing and health of city residents (Elmqvist et al  
52 2013). This includes basic resources such as fresh water and food, as well as life-improving benefits such  
53 as opportunities for recreation, local climate regulation, and improvements in air quality (MA 2005;  
54 TEEB 2011). Given the projected dramatic increase in urbanization around the world (Seto et al. 2012),  
55 managing and optimizing urban ecosystem services is critical for social and ecological sustainability.  
56 Incorporating specific goals for managing and improving ecosystem services into urban planning and  
57 management has therefore been strongly endorsed (Bolund & Hunhammar 1999; Niemelä et al. 2010;  
58 Gómez-Baggethun & Barton 2013) and is increasingly explored in theory and practice (Tratalos et al.  
59 2007; Cowling et al. 2008; TEEB 2011; Elmqvist et al. 2013; Lovell & Taylor 2013). However, empirical  
60 research on urban ecosystem services has generally neglected clear, contextual links between  
61 ecosystems and the benefits people derive from them (Luederitz et al. 2015).

62 In seeking to address this research gap, some scholars have highlighted the importance of the  
63 socio-economic circumstances of urban residents for determining benefits received from urban green  
64 space (e.g. Lin et al. 2014, Shanahan et al. 2014). However, why, when, and how socio-economic factors  
65 mediate ecosystem service has been poorly synthesized to date (Carpenter et al. 2009). The paucity of  
66 usable models and tools presents an even more immediate challenge for real-world application to guide  
67 the inclusion of these considerations into urban planning and management. In this paper, we use the  
68 ecosystem service supply chain framework to synthesize how socio-economic factors influence those  
69 services for people living in cities, crafting a conceptual model as a decision aid. We then identify how  
70 this can be used by planners and managers to improve the provision of ecosystem services in cities.

71 The supply of and demand for ecosystem services is not homogeneous across any individual city.  
72 Importantly, ecosystem service demand is determined by the needs and desires of people and is

73 influenced by socio-economic factors such as income, wealth, education, and ethnicity (MEA 2005;  
74 Rounsevell et al. 2010; Ernston 2013). Socio-economic factors can also influence green space  
75 management and planning decisions, leading to uneven supply of green spaces across cities (Pham et al  
76 2012). Thus, spatiotemporal variation in socio-economic factors within cities can lead to significant  
77 variability in the supply and demand of ecosystem services derived from green spaces (McDonald 2009;  
78 Escobedo et al. 2011). This means that the relationships between socio-economic factors and ecosystem  
79 services should be a key planning and management consideration (Cowling et al. 2008; Lyytimaki &  
80 Sipila 2009; Gómez-Baggethun & Barton 2013), despite rarely being addressed in urban planning policy  
81 or scholarship.

82           Three key insights about the role of socio-economics in urban ecosystem services are currently  
83 evident from the literature and all hinge on 'differences': (1) green spaces are perceived and used  
84 differently by different demographic groups (e.g., Madge 1997; Tinsley et al. 2010), (2) there are often  
85 inequalities in green space provision along socio-economic gradients (e.g., Pedlowski et al. 2002; Pickett  
86 et al. 2008), and (3) the types and importance of ecosystem services to urban residents can differ along  
87 socio-economic gradients (e.g., Tratalos et al. 2007; Lubbe et al. 2010; Cilliers et al. 2013). Importantly,  
88 recent research has started to reveal the potential mechanisms by which socio-economic factors can  
89 influence ecosystem service benefits. For example, Shanahan et al. (2015) showed that higher formal  
90 education levels and greater neighbourhood socio-economic advantage are associated with the use of  
91 local parks that incorporate native remnant ecosystems. Additionally, Peterson et al. (2008) showed that  
92 residents choosing to live in more natural areas were older, better educated, and more environmentally-  
93 oriented than those choosing residential areas with less green space.

94           With such evidence accumulating, there is an urgent need to bring these threads together to  
95 improve the conceptual understanding of how socio-economic factors influence ecosystem services in  
96 cities that can then be operationalized for urban planning. Such a model could then directly improve

97 ecosystem service management by delineating and linking ecosystems service components such that  
98 urban policy-makers, planners, and managers can more clearly consider critical contextual factors in  
99 their focal areas (Cowling et al. 2008; Luederitz et al. 2015). Without this, there is the risk that planning  
100 initiatives to improve the quantity or quality of green space across cities will result in fewer or less  
101 equitable benefits for city inhabitants. We note here that, while some decision-making factors for  
102 private spaces differ from those for public spaces, planners and managers must influence both for  
103 equitable ecosystem service provision (Aronson et al. 2017). Many cities have simple prescriptive targets  
104 for green space quantity and spacing that are intended to provide equal access (Heynen et al. 2006), but  
105 these well-meaning targets may need to be reconsidered in the context of varying socio-economic  
106 contexts from city to city and within any given city.

107         Here, we first identify and conceptualize how socio-economic factors influence the supply,  
108 demand, and benefit of ecosystem services to people in cities. By framing this around the ecosystem  
109 service supply chain framework (also known as the 'ecosystem service cascade'), we distinguish  
110 between the biophysical supply of a service, the demand for it, and the benefit it gives people (Potschin  
111 & Haines-Young 2011). In turn, we focus on how socio-economic factors influence the links in the supply  
112 chain and illustrate this via three urban ecosystem service/disservice examples: moderation of  
113 temperature extremes, urban gardening, and fear and stress reactions. We then outline ways forward  
114 for planners and managers to apply this understanding by providing specific suggestions about how to  
115 use these concepts and the model to deliver better urban ecosystem service outcomes.

116

## 117 **2 Linking socio-economic factors to ecosystem services**

118 Our conceptual model distinguishes between the biophysical *supply* of an ecosystem service, the  
119 *demand* for it by people, and the *benefit* that people receive from a service that contributes to their  
120 well-being (Potschin & Haines-Young 2011; Tallis et al. 2012, TEEB 2010; Fig. 1). Urban ecosystems

121 provide biodiversity and ecosystem processes that can potentially provide ecosystem services to people  
122 (i.e. ecosystem service supply). Socio-economic factors in cities affect ecosystem services through two  
123 distinct and interrelated direct pathways: (1) by influencing the management of urban green space and  
124 in turn ecosystem service supply, and (2) by altering human needs and activities and therefore people's  
125 demand for specific ecosystem services. For certain services, there is an (3) indirect pathway whereby a  
126 resident's socio-economic status can influence how the provision of an ecosystem service affects their  
127 wellbeing (i.e., their physical or psychological health). Along each of these pathways, ecosystem services  
128 can also feed-back to influence socio-economics (e.g., Wolch et al. 2014) although we do not focus on  
129 that bidirectionality here. Our model emphasizes the need to understand these multiple pathways  
130 through which socio-economic variables influence both the biophysical and social aspects of urban  
131 ecosystem service provision (Bagstad et al. 2013).

132

### 133 *2.1 Socio-economic factors influence the supply of services*

134 Changes to the amount and characteristics of urban green space affect the presence and abundance of  
135 species, the structure of vegetation, the ability of urban residents to access green space, and,  
136 subsequently, the ability of urban green spaces to actually supply ecosystem services (Gaston et al.  
137 2013, Caynes et al. 2016). Socio-economic factors influence the ecosystem services supplied by green  
138 spaces by altering how much green space is present in cities and how it is managed (Figure 1). For  
139 example, city regulations, zoning laws, and management of both public and private green spaces often  
140 heavily influence the presence, composition, and structure of urban vegetation which can regulate  
141 temperature if managed toward that goal, and those policies and management approaches are often, in  
142 turn, influenced by socio-economics (Case Example 1).

143

144

145 **Case Example 1: Supply of regulatory services and urban vegetation**

146 The frequency of extreme temperature events has increased over time, a trend expected to increase in  
147 coming decades (Morak et al. 2013). Episodes of extreme temperatures are responsible for increased  
148 mortality in urban populations (Patz et al. 2005, Hondula and Barnett 2014) and are the second leading  
149 cause of climate-related deaths in the USA (Knowlton et al. 2011).

150 Urban green spaces and planted trees can ameliorate extreme temperatures as they reflect  
151 light, shade buildings, and lead to localized cooling through evapotranspiration (Loughner et al. 2012).  
152 For example, in the US coastal cities of Washington, D.C. and Baltimore, surface temperatures were 4°C  
153 cooler in streets in areas with vegetation while roads and buildings were 10-15°C cooler, and detailed  
154 climate modelling indicated that the presence of urban trees increased the velocity of cooling sea  
155 breezes into the cities (Loughner et al. 2012). In Phoenix, Arizona, high rates of fatalities were recorded  
156 among the homeless population within the central city area and industrial corridors where surface  
157 temperatures ran high, little vegetation cover existed, and air-conditioned shelters and medical services  
158 were less available (Jenerette et al. 2011, Harlan et al. 2013). Therefore, investment in high quality,  
159 heat-reducing green space for poorer neighborhoods is recommended as a means of reducing social  
160 inequity (Jenerette et al. 2011).

161 Policy initiatives can markedly influence the incentives and ability of a city and its planners and  
162 managers to address the needs of urban residents who have a strong need for a greater supply of  
163 temperature regulation from green vegetation (see Supplementary Materials). With programs that are  
164 context-specific and responsive to the different geographies of need in the city, city governments would  
165 be well positioned to increase that supply of regulatory services in areas where they are most needed.

166  
167 Neighborhoods with greater socio-economic advantage commonly have more public parkland  
168 and even private lawn space than their disadvantaged counterparts (Boone et al. 2009; Dai 2011). Such

169 differences often arise due to unequal power relationships between residents and local governments.  
170 More advantaged neighbourhoods often have greater leverage and can more effectively lobby city  
171 governments (Heynen et al. 2006; Pedlowski et al. 2002; Lovell & Taylor 2013). In Baltimore, Maryland,  
172 historic societal inequalities, such as segregation ordinances, are important determinants of current  
173 inequalities in access to green space (Boone et al. 2009). In turn, lower levels of accessibility and  
174 increased distances between people's homes and green spaces often mean lower levels of green space  
175 available for recreation (Coombes et al. 2010). However, tailored green space policies may shift this  
176 recurring pattern as seen in Bristol, England where public parkland is now equally or even over-provided  
177 in poorer neighborhoods (Jones et al. 2009).

178         The structure and function of urban green spaces, usually due to management decisions, can  
179 also vary according to the socio-economic conditions of the neighbourhood in which they are sited  
180 (Aronson et al. 2017). Those with greater socio-economic disadvantage often have lower vegetation  
181 cover (Iverson and Cook 2000; Pham et al. 2012; Talarchek 1990; Shanahan et al. 2014), fewer trees in  
182 public locations (Landry and Chakraborty 2009; Kuruneri-Chitepo & Shackleton 2011), and lower species  
183 richness (Clarke et al. 2013; van Heezik et al. 2013). A range of socio-economic reasons contribute to  
184 these patterns. For example, more advantaged populations can often afford larger properties in older  
185 neighbourhoods, which are associated with greater availability of space and time for vegetation  
186 establishment (Kirkpatrick et al. 2007; Lowry et al. 2012). Similarly, an individual's income and  
187 knowledge of the benefits that urban green space provides may influence the extent to which they  
188 create or maintain green space within their yard or communal space (Heynen et al. 2006; Andersson et  
189 al. 2007; Kirkpatrick et al. 2007).

190         Ethnicity and the subsequent norms thereof can also play a large part in modulating the  
191 characteristics of urban green spaces. In South Africa, residents of Botswanan descent clear their yards  
192 of all vegetation because of group norms about tidiness (Lubbe et al. 2010). Additionally, a number of



193 studies have found that culture, demographics, housing type, and ownership can influence private or  
194 community-land land management (e.g., Talarchek 1990; Troy et al. 2007). How urban space is  
195 managed, e.g., the type of plants chosen or the hours spent on maintenance, can result in striking  
196 differences in grass versus tree cover and in amount of greenery overall.

197

## 198 *2.2 Socio-economic factors influence demand for services*

199 The link between socio-economic factors and demand for services has, to date, received little attention  
200 (Burkhard et al. 2012). People have numerous needs, including basic material for a good quality of life,  
201 access to clean air and water, security from disasters, and good social relations (MA 2005). Maslow  
202 (1943) proposed a hierarchy of needs to define universal human needs and this framework has been  
203 widely adopted in psychology, sociology and management (Figure 2). It categorizes need according to  
204 five levels, physiological, safety, love/belong, esteem, and self-actualization, where those at the bottom  
205 (e.g., physiological, safety) are more 'fundamental' than those at higher levels (e.g., esteem, self-  
206 actualization). While the ranking of human needs in this way has been criticized (Wahba & Bridwell  
207 1976), we argue that such categorization, although not necessarily a strict hierarchy *per se*, is useful  
208 when considering how socioeconomic factors influence these different types of needs and,  
209 subsequently, how this might change demands for different ecosystem services. For example, as people  
210 increase in socio-economic advantage (e.g., increased income or higher levels of education), their  
211 demand for ecosystem services related to esteem and self-actualization (e.g., recreational or cultural  
212 services) may increase relative to those for services related to physiological health (i.e., food supply)  
213 that can be provided by remote locations outside the city or those services related to safety (e.g., flood  
214 or climate regulation) that can readily be met by technological means. This shift is exemplified in South  
215 Africa, where poor urban residents use their garden space for supplementary food production, whereas  
216 wealthier residents use gardens for relaxation and aesthetic services (Cilliers et al. 2013).

217            Socio-economic factors influence human behaviors that alter access to ecosystem services  
218 (Figure 1). Public parks are regularly cited as critical green space in urban landscapes; however, people  
219 must visit parks in order to receive certain ecosystem service benefits. Urban green space visitation  
220 rates are strongly influenced by crime rates, perceptions of safety, age, gender, cultural background, and  
221 socio-economic status (McCormack et al. 2010, Cohen et al. 2013, Reis et al. 2012, Peschardt et al. 2012,  
222 Lin et al. 2014, Shanahan et al. 2015). Visitation rates often reflect the outcome of supply, demand, and  
223 provision of ecosystem services but may directly indicate demand if supply and provision are controlled  
224 for or held constant. For example, Jones et al. (2009) found that over 40% of people in the most  
225 advantaged socio-economic group visited parks in Bristol, UK, compared to only 27% in the least  
226 advantaged group despite greater accessibility for this latter group. This disparity between socio-  
227 economic groups was driven by differing perceptions of reduced accessibility and compromised safety  
228 (Jones et al. 2009). Similarly, in an Australian city, Leslie et al. (2010) found that perceptions of safety  
229 and opportunities for socialization in green spaces resulted in more frequent park visitation and greater  
230 participation in walking activities for higher-status individuals. Perceptions that parks are unsafe are  
231 consistently more pronounced in disadvantaged areas and for specific ethnic groups (e.g., Lyytimäki &  
232 Sipilä 2009, McCormack et al. 2010) and could substantially diminish ecosystem service demand and  
233 thus any eventual benefits (further explored in Case Example 2).

234            In the USA and parts of Europe, ethnicity explains some major differences in the use and  
235 preferences for outdoor recreation of non-white immigrants or non-white established populations  
236 compared to established white populations (Madge 1997; Johnson & Bowker 1999; Gobster 2002;  
237 Tinsley et al. 2010; Gentin 2011). These ethnic differences can also play out at a country-wide level.  
238 Özgüner (2011) highlights that Turkish visitors use parks more for passive recreation (e.g., picnicking)  
239 than visitors from Western countries, perhaps as a reflection of the more collective Turkish lifestyle.

240 Even across a city where parks are managed in similar ways and their distribution is equitable, they may  
241 provide very different benefits if demand for their services varies with socio-economic conditions.

242

243 **Case Example 2: Disservices that diminish park visitation demand**

244 While maximizing trees and shrubs in urban parks can appear to be a good idea, benefiting climate  
245 regulation, air purification, noise reduction, recreation, and aesthetics (Escobedo et al. 2011; Dobbs et  
246 al. 2014), for some urban residents that type of park design can have significant trade-off's (see  
247 Supplementary Materials). In fact, higher levels of woody vegetation may lead to heightened fear and  
248 stress as well as other disservices such as increased allergens and potential for infrastructure damage  
249 (Lyytimaki & Sipila 2009; Escobedo et al. 2011; Dobbs et al. 2014). In Leicester, Britain, Madge (1997)  
250 found that fear was a strong deterrent against park usage and demand for parks by women, the elderly,  
251 and Asian and African-Caribbean demographic groups, stemming from concerns about sexual violence,  
252 theft, and racial discrimination respectively. Vegetation cover can contribute to a perception that  
253 vegetation can conceal criminals and limit the vision of potential victims and surveillance (Kaplan et al.  
254 1998; Reis et al. 2012).

255         Responsive city and neighborhood policies and management practices can alter these  
256 disservices, which may be especially important for vulnerable demographics. In Zimbabwe, lighting was  
257 more important than vegetation in determining crime in poorer neighborhoods (Nyabvedzi & Chirisa  
258 2012). Obviously well-maintained vegetation can deter criminal activity due to the indication of higher  
259 levels of authority and surveillance (Wolfe & Mennis 2012). Thus, demand for green space services can  
260 be enhanced through top-down regulation that aims to increase the perception of safety in  
261 neighborhoods with higher crime rates. This could take the form of outreach programs as well as  
262 specific park design considerations that alter the look and feel of parks in areas where perceptions or  
263 realities linked to socio-economic conditions might diminish apparent demand for green areas.

264 Increased community involvement in parks and greater ‘informal surveillance’ along with the presence  
265 of authority figures may also alleviate perceptions of fear and stress disservices (Madge 1997).

266

267 Maslow’s categories of human needs also vary with social factors in their potential to be met via  
268 technology and built infrastructure instead of from ecosystem services provided by urban greenspace  
269 and natural features. Those related to physical wellbeing and safety can be most easily substituted with  
270 increases in material wealth. Water and waste treatment needs can be met through water supply and  
271 sewer systems; flood regulation by the construction of dams, canals, and levees; climate regulation from  
272 air-conditioned buildings, and food through the import of agricultural products from more distant  
273 locations. Wealthier or more educated cities and countries may be better able to substitute or use  
274 technological solutions for water provision or flood mitigation (Luck et al 2009), reducing demand for  
275 these services from natural ecosystems. Poorer inhabitants of cities may rely more upon the cooling  
276 effect of nearby vegetation during heatwaves, while wealthier residents rely on more expensive air  
277 conditioning (Cavan et al. 2014). The MillionTreesNYC campaign recognizes that socio-economic status  
278 influences demand for temperature regulation from trees and places substantial focus on planning in  
279 “low-income and poor-health” neighborhoods (McPherson et al. 2011). Thus, substitution may reduce  
280 the demand for urban green space to provide certain ecosystem services but only if socio-economic  
281 conditions allow for adequate substitution. In contrast, substitution of services related to self-  
282 actualization or esteem (e.g., cultural services) may be more difficult. Therefore, demand for ecosystem  
283 services related to these particular needs may be insensitive to changes in socioeconomic factors. The  
284 impact of socio-economic factors on demand for ecosystem services may be especially complex if there  
285 is a negative relationship between true need and apparent demand. As described above, those who may  
286 benefit most from green space may not necessarily express (or have the power to express) demand for

287 that space or associated services. This potential tension and its effect on ecosystem services should be  
288 explicitly considered in green space planning and management.

289

### 290 *2.3 Socio-economic factors moderate benefits of services*

291 Socio-economic factors can also influence the actual benefit that people receive from the use of an  
292 ecosystem service, even as the level of service supply or demand stay constant between groups of  
293 people (Figure 1). A service can be fully supplied and there can be demand for it, but the benefit it  
294 provides (e.g., how it contributes to human wellbeing) can vary depending on socio-economic factors  
295 (de Groot et al. 2010; Potschin & Haines-Young 2011). For example, urban gardens can be equitably  
296 distributed and even similarly structured (supplied) and equally used (demanded) by differing groups of  
297 people but the benefit they derive from them may differ depending on whether they gain primarily a  
298 provisional service benefit, such as food, or primarily a cultural service benefit, such as sense of place  
299 (Case Example 3). Those differences in how the same urban green space can benefit an individual or  
300 community can be driven by socio-economic status. Of all the connections between ecosystem services  
301 and socio-economic factors, the link between socio-economics and benefits is the least studied and  
302 most poorly understood or appreciated.

303

#### 304 **Case Example 3: Benefits of provisioning & cultural services and urban gardens**

305 Urban gardens are often associated with the cultural values and liveability of cities, providing a range of  
306 ecosystem services (Barthel and Isendahl 2013). In South Africa, the importance of food provision from  
307 gardens relates to socio-economic gradients in that species that are useful as food are more frequent in  
308 the gardens of poorer residents who use gardens as a source of additional income or supplemental food  
309 (Lubbe et al. 2010; Cilliers et al. 2013). The same gardens that provide food may also form a crucial part  
310 of a community's sense of place and control, services that marginalized populations may find especially

311 difficult to procure (Anguelovski 2013). Thus, the realization of different ecosystem service benefits may  
312 vary along with changes in socio-economic status (see Supplementary Materials).

313           When focusing on enhancing benefits from ecosystem services, city planners and managers  
314 would likely adjust policies and management schemes, rather than generating new ones. Urban  
315 managers could influence the strength and type of benefits through outreach efforts focused on  
316 increasing awareness around different functions of urban gardens, including holding gardening classes,  
317 and also by offering incentives that encourage and enable disparate urban dwellers to participate in  
318 gardening that is tailored to their needs (e.g., food versus aesthetics). Alternatively, planners and  
319 managers could focus their efforts in direct response to the existing type and level of demand and  
320 develop garden-friendly incentives and programs in areas of highest demand where those efforts would  
321 have the most rapid uptake and impact.

322

323           Perhaps the best example of this link between socio-economics and ecosystem service benefits  
324 relates to food security, which depends on food availability, access, utilization, and stability (FAO 2006).  
325 In South Africa, urban residents make socioeconomically-dependent planting choices in their urban  
326 gardens with implications for eventual food security benefits (Lubbe et al. 2010; Cilliers et al. 2013).  
327 Lubbe et al. (2010) found that South Africans with lower socio-economic status planted more utilitarian  
328 plants such as fruit trees despite their higher expense and long-term commitment needed for their  
329 culture because of job and market insecurities. However, while urbanites may not be barred  
330 economically or culturally from investing in natural resources such as fruit trees (i.e. increasing the  
331 supply to match demand), their ability to actually benefit from such investments can be hindered by  
332 other socio-political limitations like tenure security (e.g., Otsuka et al. 2001). Thus, despite investments  
333 in supply of certain ecosystem services and apparent demand, we speculate that the end benefit of the  
334 service may not be realized due to socio-economic factors. There may also be different levels of benefit

335 that differing demographics may receive from ecosystem service provision. For example, the health and  
336 wellbeing benefits that can be gained from recreation in green space could be much higher for  
337 disadvantaged communities simply because their base-line wellbeing is lower and ultimately these  
338 people can have more to gain. There is support for this concept in that the health benefits of  
339 neighborhood green space tend to be much more evident for lower income communities (Mitchell &  
340 Popham 2008). The link between service provision and actual benefit is a nuanced one. Many of the  
341 same strategies that managers or city government officials can take to enable or incentivize benefits of  
342 ecosystem services will be closely related to, or even the same as, those used to alter people's demand.  
343 Yet consideration of the transformation of service provision to actual benefit will improve the chances  
344 that ecosystem services will benefit target audiences and thus feedback to influence the demand for  
345 such services.

346

### 347 **3 Implications for city planners & land managers**

348 *We detailed the conceptual model to demonstrate its utility in organizing thinking and examined case*  
349 *examples to demonstrate its ability to operationalize current frameworks and corresponding theory and*  
350 *evidence. Practical implications of the use of this model are detailed below along with complementary*  
351 *methods and tools.*

#### 352 *3.1 Improvement in 'standards-based' urban green space planning*

353 Urban green space planning is commonly based on targets that describe a minimum area of green space  
354 per person or household and proximity to residential areas (Heynen et al. 2006). For example,  
355 accessibility standards for the United Kingdom are based on targets for the area of green space that  
356 should be within certain distances of people's homes (Natural England 2010), and the UN Habitat State  
357 of the World's Cities report suggests that a minimum of 8 m<sup>2</sup> of green space per person is required (UN-  
358 Habitat 2013). These approaches provide important guidelines that, if implemented, can assist in

359 creating equity in the amount of green space available across socio-economic gradients (Shanahan et al.  
360 2014). Yet, even if supply is uniform across a city, demand almost certainly will not be due to the  
361 different ways socio-economic factors influence supply versus demand versus benefits (Fig 1). The  
362 implications are that targeted green space provision, based on the spatial distribution of demand and  
363 potential benefits relative to socio-economic factors, can result in more equitable distribution of  
364 ecosystem service benefits. As such, a one-size-fits all approach to green space planning and  
365 management will not ensure that ecosystem service benefits are equally realized (Escobedo et al. 2011).

366

### 367 *3.2 Understanding relationships between socio-economic factors and ecosystem services*

#### 368 3.2.1 Local assessment of ecosystem service supply and demand

369 Simply identifying where socio-economically advantaged and disadvantaged groups live within cities will  
370 likely provide some information to guide efforts directed at enhancing green space supply and demand.  
371 However, the most useful information will come from community surveys, focus groups and interviews  
372 that examine residents' perceptions and usage and experience of green spaces. This will be particularly  
373 useful for developing strategies tailored to the specific concerns or barriers associated with any one  
374 community. Community surveys can help managers gauge high and low demand so that they can  
375 prioritize management of particular ecosystem services relevant to the neighborhoods of that area  
376 (TEEB 2010). For example, in communities where personal safety is considered an important barrier to  
377 green space use, social strategies that include increased policing (Wilbur et al. 2002) or planning  
378 strategies that enhance the design of green spaces to increase visibility and perceptions of safety  
379 (Schroeder and Anderson 1984) may be appropriate. These strategies speak to the interplay between  
380 management of green space and human needs and activities as mediated by considerations such as  
381 access, incentives and outreach, as well as policy goals (Figure 1).



382 Understanding community values will complement current understanding of perceptions and  
383 usage of urban green spaces. Management of green spaces, particularly around ecosystem services, is a  
384 process of articulating values, both of management and of stakeholders, and responding to those values  
385 (Ernstson & Sorlin 2013; Ives & Kendal 2014). Various mapping tools can be used to elicit the values of  
386 stakeholders spatially, such as Public Participation GIS, which may be particularly useful to green space  
387 managers (Ives et al. 2017). Using data from community surveys or methods like Public Participation GIS,  
388 managers can map out and qualitatively model the flow of prioritized services (e.g., Brown et al. 2014).  
389 To enhance green space planning and policy, the available information on community-specific socio-  
390 economic factors that prevent the use of green space could be used to identify particular areas or  
391 groups of need.

### 392 3.2.2 Quantitative analysis to understand drivers of green space benefits

393 The above methods will allow a basic characterization of our conceptual model's components whereas  
394 quantitatively-based modeling approaches are one suite of tools that could provide understanding of  
395 the interactions between supply and demand and predict ecosystem service outcomes. Knowledge of  
396 the strength and form of these interactions should better enable planners and managers to anticipate  
397 how altering characteristics of one component of the model may affect ecosystem service provision (Fig  
398 1). The dynamics of socio-ecological systems often also have strong feedbacks between the social and  
399 ecological components (*en sensu* McPhearson et al. 2016). In particular, these feedbacks can drive the  
400 land management decisions made by municipalities and individuals in urban areas that may either  
401 negatively or positively influence urban ecosystems (Alberti et al. 2003). This more predictive  
402 understanding would be helpful in cases when new management strategies are being tested or where  
403 the demographics or wealth of a neighborhood around or containing green spaces are changing.  
404 Qualitative, participatory methods that include economic valuation are likely to be more appropriate if

405 the objective is to explore the deeper meanings, values and interactions urban residents have with their  
406 local environment.

407           There is a need to develop more effective modeling techniques to enable landscape  
408 practitioners to apply evidence of the links between ecosystem service components and socio-  
409 economics in real-world contexts. Whichever modeling approach is used, there are three key  
410 components of the process: (1) gather data on critical or likely socio-economic factors that influence  
411 supply, demand, and benefits of ecosystem services (e.g., common factors detailed above as influential  
412 to services), (2) relate these to the physical/environmental variables that influence them (e.g., green  
413 space provision, condition, arrangement), and (3) model the impact of specific planning or management  
414 interventions that can affect outcomes (e.g., management actions, behavioral incentives, access  
415 improvement) (see Cowling et al. 2008). One of the biggest challenges of such quantitative modeling is  
416 the integration of social and environmental factors, which are measured using different techniques,  
417 scales, and units. In particular, many socio-economic variables are non-spatial, while the green spaces  
418 being managed are spatially located. In recent years, much work has been done on spatially mapping  
419 ecosystem service flow, supply and demand (van Jaarsveld et al. 2005; Burkhard et al. 2012, Garcio-  
420 Nieto et al. 2013, Dobbs et al. 2014). Yet future work must move past spatial representation of existing,  
421 static relationships to prediction and extrapolation across space and time. Examples of emerging  
422 approaches that can help in this strategy include applying techniques developed for species distribution  
423 modeling to associations between social values and environmental conditions (e.g. the Social Values for  
424 Ecosystem Services tool; Sherrouse et al., 2011) and spatially-referenced agent-based modeling (e.g.,  
425 Matthews et al. 2007).

426

427 *3.3 Implementation changes in planning, policy, and practice to enable ecosystem service benefits*

428 A variety of innovative solutions for planners and managers can enable greater realization of ecosystem  
429 service benefits to a broader range of socio-economic groups. We note that a few success stories exist  
430 where policy makers and urban planners and managers successfully incorporated socioeconomic factors  
431 into ecosystem service work, such as the Milwaukee River Greenway run by a private and public  
432 community coalition (Aronson et al. 2017) and the Corridors of Freedom initiative in South Africa which  
433 is intended to connect socio-economically segregated communities via green infrastructure (The  
434 Guardian 2015). A few more posited interventions have already been mentioned here regarding specific  
435 services, such as planting more shade trees in neighborhoods that have less access to air-conditioning  
436 (Case Example 1).

437 For ecosystem service benefits such as recreation or food-provision, planners and managers can  
438 enact strategies to alter the supply of services and help enable positive behavioral or perception  
439 changes (see dashed lines between 'Management of Green Space' and 'Human Needs and Activities' in  
440 Figure 1). For example, when planning for new green spaces, underutilized urban areas can be  
441 incorporated such as vacant lots which may already be more prevalent in underserved communities.  
442 These types of new green spaces and others, like community gardens, can be co-managed with informal  
443 managers, dedicated citizens who can help foster community buy-in and build social capital (Andersson  
444 et al. 2007). Programs that lower the knowledge and resource barrier to private space gardening and  
445 greening (e.g., free tree seedlings or classes) might encourage community-level behavior shifts, though  
446 messaging must be carefully tailored to ensure equitable community buy-in (see Locke & Grove 2014  
447 and dashed lines in Fig 1). Community engagement programs and activities in parks as well as  
448 government commitment to increase safety and a sense of belonging can also help overcome socio-  
449 economic barriers to park use (Cohen et al. 2013). In order to work with demographic differences, park  
450 managers might do well to provide an array of facilities to attract a more diverse array of visitors  
451 (Burgess et al. 1988; Gobster 2002) and design public spaces that satisfy public preferences for

452 cleanliness and order, even in more natural settings (Burgess et al. 1988; Gobster & Westphal 2004, Ives  
453 & Kelly 2016). Managers can also use different marketing strategies, including social marketing  
454 strategies, about specific park amenities to attract underrepresented sectors of society (Johnson &  
455 Bowker 1999; Lovell & Taylor 2013; Ives & Kendal 2014).

456

#### 457 **4 Conclusions**

458 A number of ecosystem service frameworks have been put forward that consider socio-economic  
459 variables or influences (e.g., Carpenter et al. 2009; Daily et al. 2009; de Groot et al. 2010). However, the  
460 specific links between socio-economic variables and ecosystem service provision have rarely, if ever,  
461 been explicitly conceptualised for urban planning (Carpenter et al. 2009). Our conceptual model  
462 explicitly embeds these links within the ecosystem service supply chain framework. By doing so, it  
463 emphasizes the importance of socio-economic factors in managing urban ecosystem services and  
464 identifies potential pathways through which land managers and policy-makers might intervene to alter  
465 ecosystem service provision.

466         Socio-economic factors can have a profound influence on the demand and supply of urban  
467 ecosystem services, and they heavily mediate the benefits that city residents can receive from green  
468 spaces. Consequently, urban planning that incorporates these factors into the provision and design of  
469 green spaces has the potential to markedly enhance health and wellbeing through more effective  
470 delivery of ecosystem services. Our model allows the identification of specific socio-economic barriers to  
471 ecosystem service delivery and will potentially reveal what types of interventions are necessary and  
472 where. Ultimately, this approach could shift planning strategies towards ecosystem service provision  
473 that better meets the needs and desires of diverse urban residents.

474

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480

481

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**Box 1: Glossary**

*Ecosystem services*: the biophysical and social conditions and processes by which people, directly or indirectly, obtain benefits from ecosystems that sustain and fulfill human life (MA 2005).

*Ecosystem service supply*: the full potential of ecological functions or biophysical elements in an ecosystem to provide a given ecosystem service, without consideration of whether human recognize, use, or value that function or element (Tallis et al 2012, Villamagna et al 2013).

Ecosystem service benefit

*Ecosystem service demand*: the level of service benefit desired or required by people. Demand is influenced by human needs, values, institutions, built capital, and technology (Villamagna et al 2013).

*Ecosystem service provision*: the realisation or delivery of an ecosystem service resulting in actual benefit to people. Provision depends on both the supply of and demand for a service (Tallis et al 2012, Villamagna et al 2013).

*Urban green space*: all the natural, semi-natural and artificial networks of multifunctional ecological systems within, around and between urban areas, at all spatial scales (Tzoulas et al. 2007). This includes both public and private green space, including parks, private yards and gardens, street trees, green roofs, etc.

*Socio-economic factors*: the combination or interaction of social or economic characteristics related to an individual or group, including occupation, education, income, and place of residence.

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719 **TABLES**

720 **Table 1.** Ecosystem services considered to be especially relevant to urban residents, list adapted from  
 721 Chapter 11: Urban Ecosystem Services in Elmqvist et al. 2014 using the service categories from the  
 722 Millenium Ecosystem Assessment 2005.

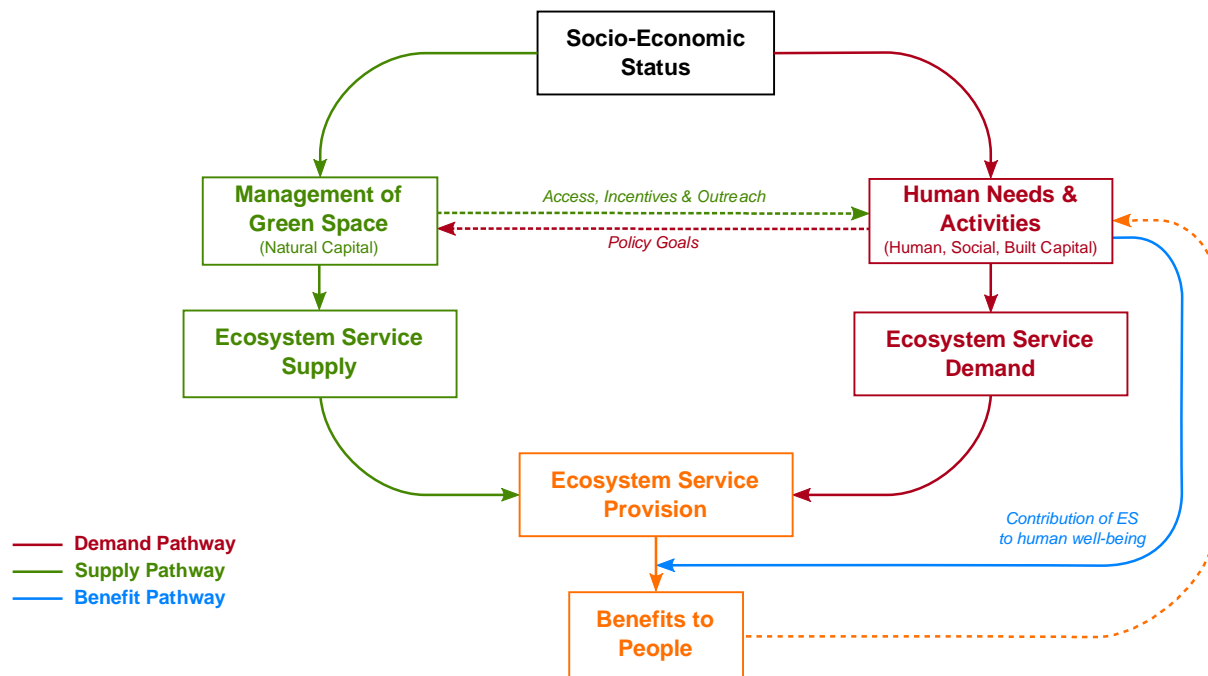
Categories	Services
Provisioning	Food supply
	Water supply
Regulating	Urban temperature regulation
	Noise reduction
	Air purification
	Moderation of climate extremes
	Runoff mitigation
	Waste treatment
	Pollination, pest regulation & seed dispersal
	Global climate regulation
Cultural	Recreation
	Aesthetic benefits
	Cognitive development
	Place values & social cohestion
Supporting	Habitat for biodiversity
Disservices	View blockage
	Allergies
	Accidents
	Fear & stress
	Damages on infrastructure
	Habitat competition with humans

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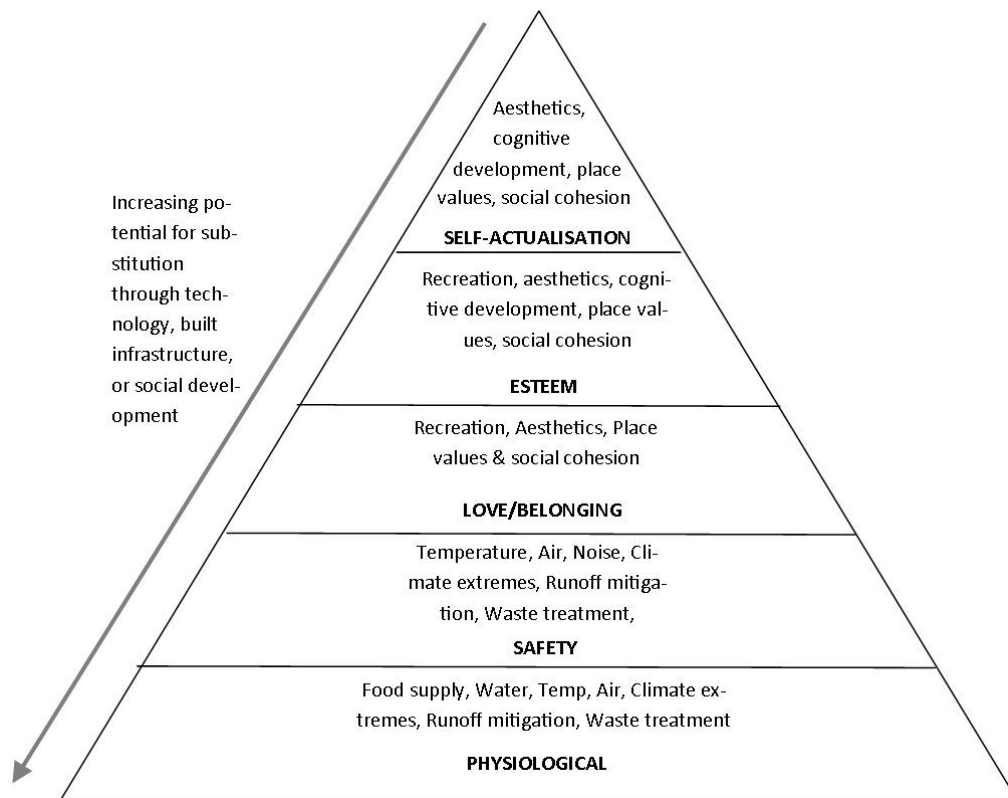
### Urban Socioecological System



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727 **Figure 1.** How socio-economic status affects the flow of ecosystem services in an urban socioecological  
 728 system. The differently colored components refer to the three main pathways by which socio-economics  
 729 can impact ecosystem service supply (1), demand (2), and benefit (3).

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731

732 **Figure 2.** Urban-relevant ecosystem services can be parsed out according to Maslow's hierarchy of  
 733 needs and the importance of ecosystems for delivering specific services may differ between differing  
 734 socio-economic sectors of a population. As the type of needs become more survival-related (more base-  
 735 level in the pyramid), there is increasing potential for substitution of ecosystem services for the same  
 736 type of services derived from technology, built infrastructure or social development.