

1 **Zn/Co-ZIF reinforced Sugarcane Bagasse Aerogel for Highly**
2 **Efficient Catalytic Activation of Peroxymonosulfate**

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12

13 **Abstract:**

14 Sugarcane bagasse is the main solid waste of sugar extraction industry. However,
15 there are less eco-friendly treatment methods for this kind of waste material. Most of
16 them have been burned and this is a challenge for effective utilization of sugarcane.
17 Herein, Zn/Co-ZIF nanoparticles are loaded onto the aerogel obtained from bagasse
18 cellulose by doping method to form pomegranate-like structure products, i.e., Zn/Co-
19 ZIF@GEL. By simulating the pomegranate-biomimetic structure, the leaching of cobalt
20 ions is suppressed and enhanced its catalytic activity. Also, Zn/Co-ZIF@GEL behaves
21 outstanding peroxymonosulfate (PMS) activation reactivity to degrade Rhodamine B
22 (RhB) with achieving 100% removal rate in 30 min with enhanced water stability. The
23 outer wrapping of nanoparticles by aerogel exhibits excellent reusability and the
24 removal rate remains above 92% after 4 cycles. X-ray photoelectron spectroscopy (XPS)
25 verifies the electron transfer between the heterostructures of Zn and Co could ensure
26 the catalyst continues to complete the redox cycle with synergistic effect. Electron
27 paramagnetic resonance (EPR) further investigates the non-radical singlet oxygen is the
28 predominant degradation pathway of RhB. This work provides a new strategy for
29 improved bagasse derivate ZIF catalytic application.

30

31 **Keywords:**

32 Bagasse cellulose aerogel, Zeolitic imidazolate frameworks, Zn doped,
33 Peroxymonosulfate (PMS) Oxidation, Water stability

34

35 **1.Introduction**

36 The conversion of biomass materials into green materials is an environmental-
37 friendly option, showing especially candidate prospects for the development of circular
38 bioeconomy [1, 2]. Sugarcane is a widely planted crop and mostly used in sugar
39 industry. The global production of its by-product, bagasse, is as high as 2.79 metric tons
40 in recent year [3]. Owing to technical limitations, labor and transportation costs,
41 bagasse is mostly landfilled or incinerated in situ, which causes great pollution to the
42 environment. However, bagasse is rich in natural resources, renewable, biodegradable
43 and non-polluting to the environment, due to which it is a green by-product of
44 sustainable development. Over the past years, bagasse has been used in films and
45 coatings, [1] glass ceramic materials [4], geopolymers [5], ceramic raw materials [6],
46 synthetic zeolites [7] and so on [8, 9]. Unfortunately, owing to the limitation of
47 technology and cost constraints, the above-mentioned applications of bagasse have not
48 been widely used. In recent years, metal-organic framework/cellulose composite
49 materials have attracted more attention. Zhu et al. [10] modified cellulose with
50 aldehyde-hydrazine and loaded with four different MOFs, including HKUST-1, ZIF-8,
51 ZIF-67 and MIL-100, which could effectively remove Rhodamine B, Cr (VI) ions and
52 benzotriazole. Yang et al. [11] used three-dimensional sodium carboxymethyl cellulose
53 aerogel as the substrate of Ni/Co-MOF, and the composite aerogel could effectively
54 remove tetracycline hydrochloride in 5 min with the removal efficiency of 80%.
55 Because of the super toughness of cellulose, it can be used as a mechanical support to
56 prepare composite materials with flexibility, formability and graded porosity [12, 13].
57 As a kind of abundant cellulose material which can be extracted from plants, bagasse
58 cellulose has great potential in practical applications. So that making full use of bagasse
59 resources will have a long-term and profound impact on waste treatment and
60 environmental protection.

61 At present, the problem of water pollution has become a worldwide environmental
62 governance problem. Organic pollutants such as drugs [14], personal care products [15],
63 dyes [16], etc., are difficult to treat with conventional water treatment processes for

64 their complex chemical structures, causing serious water environmental pollution
65 problems.. In addition, advanced oxidation process (AOPs) is a promising, efficient and
66 high-speed water treatment method that thoroughly oxidizes recalcitrant organic
67 pollutants in wastewater [17, 18]. Li et al. successfully constructed three isorecticular
68 porous 1D rod-based Ti-MOFs denoted as ZSTU-1, ZSTU-2 and ZSTU-3 for highly
69 effective visible-light photocatalysts for H₂ [19]. Zhou et al. reported on the use of light-
70 driven perylene diimide-based catalysts for the degradation of trace organic pollutants
71 present in the aqueous environment [20]. He et al. preparation Mn and O co-doped
72 polymeric carbon nitride with electronic structure modulate by Mn and O co-dopants
73 for enrofloxacin removal [21]. The AOPs based on peroxymonosulfate (PMS) has
74 shown great potential in the removal of organic pollutants. As a widely used oxidant,
75 PMS can be activated by a variety of transition metal ions (such as Mn²⁺, Co²⁺, Ni²⁺,
76 Fe²⁺, etc.), among which the Co²⁺/PMS system has the highest efficiency [22, 23].
77 However, the transition metal Co²⁺ is easy to dissolve in the reaction system, causing
78 the problem of secondary pollution [24].

79 Metal-organic frameworks (MOFs), also known as porous coordination polymers.
80 Due to the unique porous structure, ultra-high specific surface area and strong
81 adsorption capacity, MOFs attract wide attention for its potential applications in the
82 fields of separation, chemical sensing, ion exchange, energy storage and transformation,
83 drug delivery and catalysis [25-32]. Zeolitic imidazolate frameworks (ZIFs), the
84 subclass of MOFs, are as porous crystalline solid composed of tetrahedral divalent
85 metal ions (Zn²⁺, Co²⁺) connect with imidazole derivatives by coordination bonds [33,
86 34]. Since cobalt has been proven to be the most common and effective catalyst for
87 PMS, cobalt-based MOFs and the derivative materials, e.g., ZIF-67, attract much
88 attention in water treatment applications[35, 36]. Lin et al. [23] used ZIF-67 as a
89 heterogeneous catalyst to activate PMS to degrade Rhodamine B, and the degradation
90 equilibrium could be achieved within one hour at room temperature. Zhang et al. [37]
91 prepared yolk-shell Co₃O₄/C@SiO₂ nanoreactors and ZIF-67 was used as
92 the precursors for carbon and cobalt species, it could effectively degraded 90.1% of
93 Bisphenol A (BPA) in the presence of PMS. However, because of its powder

94 characteristics, MOFs are difficult to separate from water and recover from a large
95 number of reaction solutions, which hinders their application in water pollution
96 treatment [38]. Therefore, it is essential to reinforce the stability and recoverability of
97 MOFs in water. MOFs-based hydrogels and aerogels provide processability, better
98 stability and easier handling. Ren et al. used in situ synthesis to load ZIF-8 and ZIF-12
99 onto cellulose aerogels and successfully activated PMS [39]. We propose that such a
100 green strategy to prepare aerogels from biomass waste for loading MOFs particles is
101 highly desirable, which would further boost the stability of MOFs and even provide
102 opportunities to expand the potential applications of biomass materials.

103 In this experiment, we sought to construct cellulose aerogel was made from
104 bagasse as raw material, and Zn/Co-ZIF@GEL composite aerogel was prepared by
105 doping method which was used as heterogeneous catalyst to catalyze potassium
106 peroxymonosulfate oxidant to degrade RhB. Catalyst performance was evaluated
107 under different reaction conditions (Zn/Co ratios, catalyst dosage, initial pH, PMS
108 dosage and initial concentration of RhB). On one hand, cellulose aerogels served as the
109 skeleton of Zn/Co-ZIF to wrap nanoparticles that could improve the stability in solution,
110 and simplify the catalyst recovery method without affecting the degradation
111 performance. On the other hand, the doping of Zn^{2+} significantly reduced the
112 dissolution amount of transition metal cobalt which could solve the problem of
113 secondary pollution and avoid the reduce of recycle performance. In addition, the
114 catalytic degradation mechanism and pathways of RhB were investigated by XPS,
115 radical-quenching tests and EPR experiments. In this paper, agricultural waste bagasse
116 was transformed into a carrier for water pollution treatment and synthetic composite
117 aerogel was a kind of green and stable long-term ideal material for removing organic
118 pollutants from wastewater.

119

2. Experimental

2.1. Chemical Reagents

Cobalt nitrate hexahydrate ($\text{Co}(\text{NO}_3)_2 \cdot 6\text{H}_2\text{O}$, 99%), zinc nitrate hexahydrate ($\text{Zn}(\text{NO}_3)_2 \cdot 6\text{H}_2\text{O}$, 99%), 2-methylimidazole (2-MeIM), anhydrous methanol (MeOH), sodium hydroxide (NaOH), urea, N,N'-methylenebisacrylamide (MBA), *tert*-butanol (TBA), sodium thiosulfate ($\text{Na}_2\text{S}_2\text{O}_3 \cdot 5\text{H}_2\text{O}$), were purchased from Sinopharm Chemical Reagent Co., Ltd.; Rhodamine B (RhB), 2,2,6,6-Tetramethyl-4-piperidone (TEMP), 5,5-dimethyl-1-pyrroline N-oxide (DMPO), and potassium peroxymonosulfate (PMS, $\text{KHSO}_5 \cdot 0.5\text{KHSO}_4 \cdot 0.5\text{K}_2\text{SO}_4$) were obtained from Macklin. All the materials used in this study were of analytical grade and used as received without further purification. Deionized water was applied in all the experiments.

2.2. Synthesis of Catalysts

2.2.1. Extraction of Cellulose from Sugarcane Bagasse

Cellulose was obtained from sugarcane bagasse through a facile alkaline extraction process. Firstly, the bagasse was grounded into small size, then washed with deionized water and ethanol for several times, collected by filtration and dried at $-12\text{ }^\circ\text{C}$ overnight to obtain the bagasse powder. Secondly, the bagasse powder was dewaxed in the mixture of toluene-ethanol (2:1 v/v) using a Soxhlet extractor for 6 hours. Thirdly, the pretreated bagasse (20 g) was soaked in NaOH solution (600 mL, 4 wt%) and stirred at $80\text{ }^\circ\text{C}$ for 4 hours. Then, the resulting sugarcane cellulose was obtained after washed with deionized water and ethanol for several times to remove the residual impurities. Subsequently, a certain amount of cellulose was placed into NaClO_2 solution (4 wt%, pH = 4, adjusted with glacial acetic acid), and stirred at $80\text{ }^\circ\text{C}$ for 4 hours until white cellulose suspension was obtained, which was washed with deionized water to neutral. Finally, after dried in vacuum at $60\text{ }^\circ\text{C}$ for 12 hours, the purified sugarcane cellulose was obtained.

2.2.2 Synthesis of Cellulose Aerogel

First, the NaOH/Urea/ H_2O cellulose mixture was prepared with a mass ratio of 7:12: 81, stored them in the refrigerator, and frozen at $-12\text{ }^\circ\text{C}$ for 12 hours. The 4 g of

149 bagasse fibers were added to 96 g configured cellulose dissolving solution, stirred, and
150 when the solution became clearer and more transparent, to which was added 0.8 g MBA.
151 The solution was then poured into a well plate to stand until the liquid phase was
152 converted to gel state. The cellulose hydrogel was peeled from the well plate and
153 washed repeatedly with deionized water. Finally, freeze-dried to obtain cellulose
154 aerogel. By adjusting the amount of bagasse cellulose added, different concentrations
155 of bagasse cellulose aerogels (2.5 wt%, 3 wt%, 4 wt%) were obtained.

156 *2.2.3 Synthesis of Zn doped Co-ZIF (Zn/Co-ZIF)*

157 The 3 mmol of cobalt nitrate hexahydrate and zinc nitrate hexahydrate was
158 dissolved in 90 ml of methanol, 48 mmol of 2-methylimidazole were dissolved in 90
159 ml of methanol. After stirring for 5 minutes at room temperature, the methanol solution
160 of mixed metal ions was poured into the 2-methylimidazole solution, and the mixture
161 was stirred and then left to stand. The obtained purple suspension was washed with
162 methanol several times, and the sample was collected by centrifugation. By adjusting
163 the ratio of transition metal cobalt and zinc (Co: Zn = 5: 5, 6: 4, 7: 3, 8: 2, 9: 1) to
164 prepare ZIFs materials with different doping ratios.

165 *2.2.4. Synthesis of Zn doped Co-ZIF@Aerogel (Zn/Co-ZIF@GEL)*

166 First, 4 g of purified bagasse cellulose was added to 96 g of cellulose solution and
167 placed in a mechanical agitator for intense stirring. 1 g of Zn/Co-ZIF was added to the
168 solution and stirred evenly. 0.3 g of MBA was added and continued to be stirred and
169 ultrasonic. The solution is then poured into the orifice plate to stand and washed with
170 deionized water. Finally, the Zn/Co-ZIF@GEL composite aerogel was obtained by
171 freeze-drying. The schematic of the Pomegranate-like Zn/Co-ZIF@GEL synthesis
172 process is illustrated in Fig. 1.

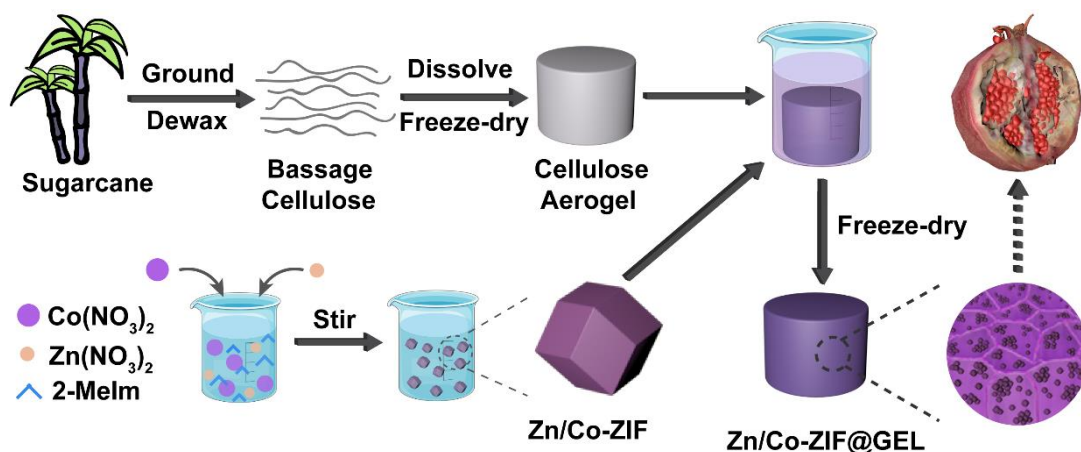


Fig. 1. The schematic of the Zn/Co-ZIF@GEL preparation process.

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176 2.3. Material Characterizations

177 The composition and water stability of the cellulose were analyzed by X-ray
 178 diffraction (XRD, Bruker D8, Germany) with Cu K α radiation (100 mA, 40 kV) at a
 179 scanning rate of 5° min⁻¹. The surface morphology of MOFs on aerogels was studied
 180 by field emission scanning electron microscopy (FESEM, Hitachi S-4800, Japan) and
 181 high-resolution transmission electron microscopy (HRTEM, JEM-2010, Japan). The
 182 mixed cellulose aerogels were characterized by thermogravimetric analysis (TGA, TG
 183 209F1, Germany). The heating rate was 20°C min⁻¹ and the temperature was 30 ~ 800°C
 184 in nitrogen atmosphere. Free radicals in the system were detected by electron
 185 paramagnetic resonance (EPR, Bruker A300, Germany) with a modulation amplitude
 186 of 3G, microwave power of 20.00 mW, microwave frequency of 9.79 GHz, and
 187 modulation frequency of 100 kHz. The elemental composition and valence states of the
 188 material were analyzed by X-ray photoelectron spectroscopy (XPS, Axis Ultra DLD,
 189 England). Al K α was used as the radiation source. The measurement range was from 0
 190 eV to 1380 eV. After the degradation experiment, the cobalt element was analyzed by
 191 inductively coupled plasma atomic emission spectrometry (ICP-MS, ICPS-7510,
 192 Japan).

193 2.4. Catalytic Activity Measurements

194 The degradation experiments were carried out in a 250 ml beaker at 25 °C. The
 195 RhB stock solution was dissolved in deionized water and stirred for 12 hours to prepare
 196 a RhB solution with a concentration of 50 mg/L, with an initial pH of 6.8. In typical

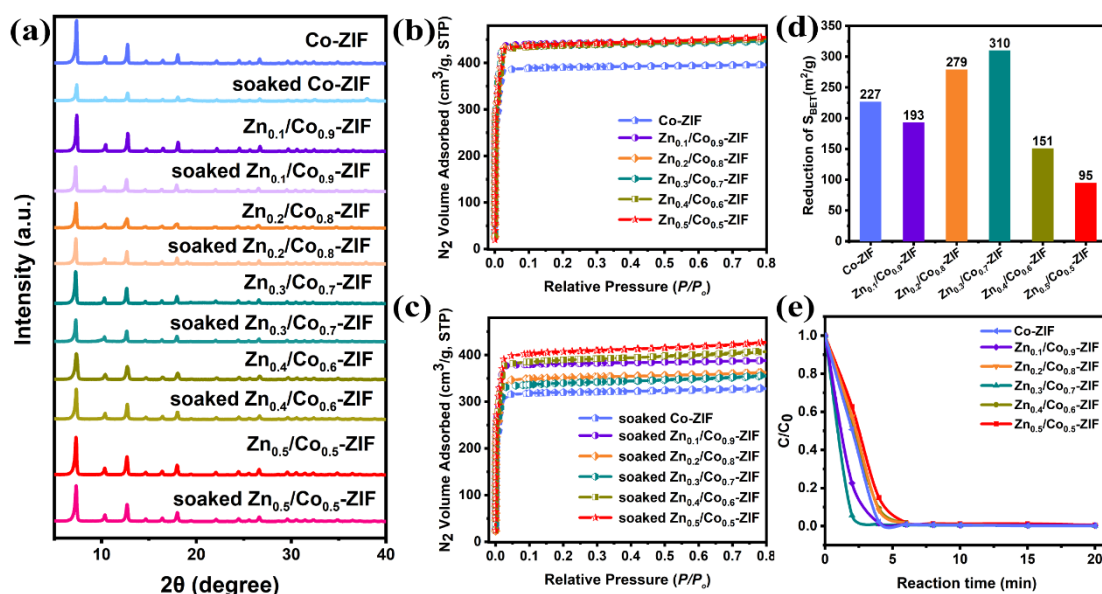
197 procedure, a certain amount of catalyst was added to the RhB solution and stirred for 5
198 minutes to completely disperse the catalyst. Subsequently, a predetermined dose of
199 PMS was added. Drew out the reaction solution (1 mL) and injected it into the Na₂S₂O₃
200 solution (50 μL) at regular intervals to scavenge free radicals and prevent further
201 reactions. The catalytic efficiency is Ct/Co (Ct: concentration of the reaction RhB, Co:
202 RhB initial concentration), and the concentration of RhB was measured at 554 nm using
203 UV spectrophotometer. Measured by the composite aerogels and bagasse cellulose
204 aerogel density, calculated as the following equation aerogel composite load:

$$\text{Loading rate} = (\rho_c - \rho_a) / \rho_c$$

205
206 ρ_c represents the density of the mixed cellulose aerogel, and ρ_a represents the
207 density of the cellulose aerogel. The calculated load rate of the mixed aerogel was $48 \pm$
208 1.8%, and the calculated load rate of the composite aerogel was 31 ± 1.2 wt%.

209 **3. Results and Discussion**

210 *3.1. Function of Zn doping in enhancing hydrostability of Co-ZIF*



211
 212 Fig. 2. (a) XRD patterns of Zn_x/Co_{1-x}-ZIF before and after soaking in water for 12 h, (b)(c) N₂
 213 adsorption-desorption isotherm curves, (d) BET surface area data of Co-ZIF and Zn_x/Co_{1-x}-ZIF, (e)
 214 degradation efficiency diagram of RhB by Zn/Co-ZIF with different Zn/Co ratios.

215
 216 Considering the Co-ZIF was moisture-sensitive and suffered from structure collapse
 217 due to the attack of water molecules, Zn doping was proposed to enhance the structural
 218 stability of Co-ZIF after soaking in water [40]. The crystal and pore structure were two
 219 important structural features among ZIFs and were separately confirmed by XRD and
 220 N₂ adsorption-desorption analyses. Fig. 2a shows the XRD patterns of Co-ZIF and
 221 Zn_x/Co_{1-x}-ZIF composites before and after soaking in water. Zn_x/Co_{1-x}-ZIF composites
 222 with different Zn doping ratios had the same SOD topology structure as the pure Co-
 223 ZIF, indicating that the Zn(II) ions have been successfully doped into the framework
 224 and a portion of Co(II) ions has been well substituted in the [Co₄O]⁶⁺ clusters [41].
 225 According to other reports, all the different Zn/Co ratio of Zn/Co-ZIF were isostructural
 226 to Co-ZIF [42], and the result was consistent with reported works. The soaked Co-ZIF
 227 showed much weaker peaks than pure Co-ZIF, suggesting that there was some
 228 decomposition happened in the framework after soaking. Since Co-N bonds are more

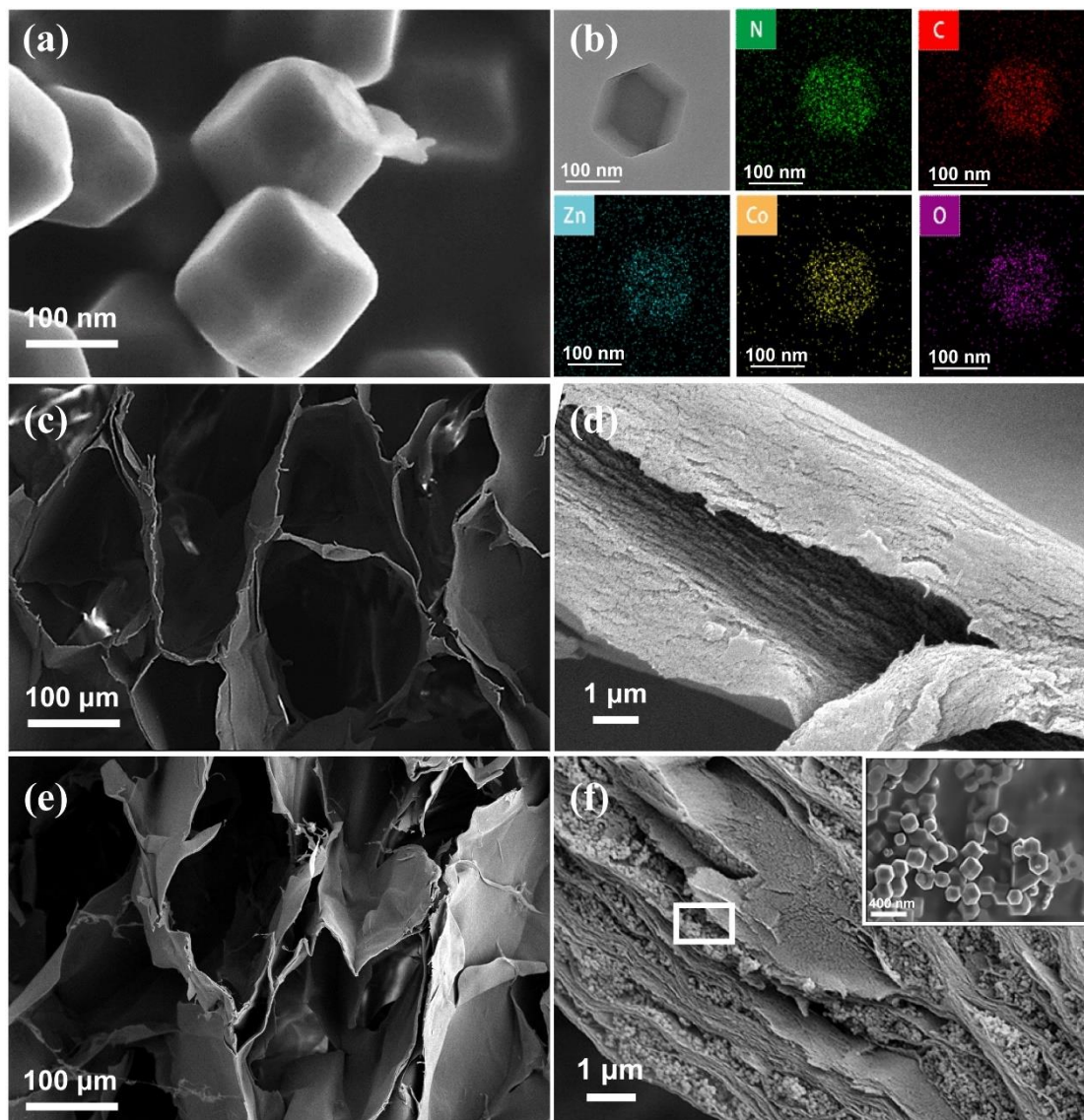
229 vulnerable to hydrolysis than Zn-N bonds [43]. After doping of Zn, the change in the
230 peaks of Zn_x/Co_{1-x} -ZIF before and after soaking became smaller than that of Co-ZIF,
231 which means that doping of Zn in the lattice results in different levels of enhancement
232 in the hydrostability of Co-ZIF. Noticeably, the crystallinity and hydrostability of
233 $Zn_{0.5}/Co_{0.5}$ -ZIF was superior to the other Zn doped Co-ZIF.

234 N_2 adsorption-desorption measurements were performed to further explore the
235 pore hydrostability of ZIFs, the surface area and pore-size distribution were analyzed
236 by Brunauer-Emmett-Teller (BET) measurements. Fig. 2b and c shows the N_2
237 adsorption-desorption isotherms of pure Co-ZIF and five different doping ratios
238 Zn_x/Co_{1-x} -ZIF before and after soaking. As seen in Fig. 2b, all Zn_x/Co_{1-x} -ZIF exhibited
239 similar microporous behavior (Type I isotherms) and almost identical equilibrium N_2
240 adsorbed volume, which was much higher than that of Co-ZIF. This indicated that the
241 Zn-doping was favorable for improved pore volume in Co-ZIF [32]. After soaking in
242 water, as displayed in Fig. 2c, all samples suffered from different extent of reduction of
243 pore volume after soaking treatment for 12 h. The reduction surface area of Co-ZIF,
244 $Zn_{0.1}/Co_{0.9}$ -ZIF, $Zn_{0.2}/Co_{0.8}$ -ZIF, $Zn_{0.3}/Co_{0.7}$ -ZIF, $Zn_{0.4}/Co_{0.6}$ -ZIF and $Zn_{0.5}/Co_{0.5}$ -ZIF
245 were calculated to be 227 (22.4), 193 (15.9), 279 (24.9), 310 (28.6), 151 (12.2), and 95
246 m^2/g (7.3%), respectively. The doping of Zn had significant influence in the
247 enhancement of pore hydrostability for Co-ZIFs except $Zn_{0.2}/Co_{0.8}$ -ZIF and $Zn_{0.3}/Co_{0.7}$ -
248 ZIF. Noticeably, the $Zn_{0.5}/Co_{0.5}$ -ZIF showed the lowest reduction of surface area,
249 suggesting its highest hydrostability to other prepared Zn doped ZIF, which was in good
250 agreement with XRD results. The enhanced hydrostability of Co-ZIFs via Zn doping
251 could be attributed to the enhanced inertness of Zn compared with Co and more stable
252 of $[Zn_x(Co_{4-x})O]^{6+}$ clusters than $[Co_4O]^{6+}$ clusters, which could decrease the hydrolysis
253 vulnerability [29, 44]. Further confirmed that Zn(II) ions have been successfully doped
254 into the framework and a portion of Co(II) ions has been well substituted in the $[Co_4O]^{6+}$
255 clusters. However, higher doping ratio of Zn may have adverse effect on the catalytic
256 performance of Co-ZIFs due to the inertness of Zn. Among all the samples, considering
257 the $Zn_{0.5}/Co_{0.5}$ -ZIF has the highest hydrostability and comparable catalytic activity and

258 kinetic to Co-ZIF (Fig. 2e), 50% was selected as the optimized doping ratio of Zn for
259 achieving high hydrostability of Co-ZIF based catalysts in the following work.

260

261 3.2. Structural characterization of $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL



262

263 Fig. 3. SEM images of (a) $Zn_{0.5}/Co_{0.5}$ -ZIF, (b) HAADF-STEM image and element mapping of
264 $Zn_{0.5}/Co_{0.5}$ -ZIF, (c)(d) bagasse cellulose aerogels, (e)(f) $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL, the upper right
265 corner is the SEM image of Zn/Co-ZIF loaded on aerogel.

266

267 The morphologies and microstructures of GEL, Zn/Co-ZIF and Zn/Co-ZIF@GEL
268 were characterized using both SEM and HRTEM analyses (Fig. 3). Fig. 3a is the SEM
269 image of $Zn_{0.5}/Co_{0.5}$ -ZIF. It was a typical rhombohedral dodecahedron crystal

270 morphology and the grain size was evenly distributed between 250 and 300 nm. This
271 was similar to the image of Co-ZIF (Fig. S1), indicating that the doping of Zn rarely
272 damages the morphology of Co-ZIF. To further research the elemental composition of
273 $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL, HAADF-STEM elemental mapping was carried out and the
274 results are shown in Fig. 3b. The distribution of N, C, Zn, Co and O were homogeneous
275 throughout the composite aerogel, reconfirming the successful doping of Zn and
276 assembling of $Zn_{0.5}/Co_{0.5}$ -ZIF heterostructure.

277 As shown in Fig. 3c, bagasse cellulose aerogel has a three-dimensional network
278 structure with large pores (100-300 μm) and honeycomb structures. The solution can
279 pass through the pores in a large amount. Fig. 3d is a partial enlarged view of Fig. 3c.
280 The cross-section of the pore walls of bagasse cellulose aerogel were rough and full of
281 folds, but the pore walls were relatively smooth. It can be seen that $Zn_{0.5}/Co_{0.5}$ -
282 ZIF@GEL has the similar morphology and internal structure with GEL (Fig. 3e). The
283 porous structure can expand the diffusion surface area of RhB molecules in the solution
284 and enhance the residence time of RhB molecules when flowing in the channel. After
285 the bagasse cellulose aerogel was compounded with Zn/Co-ZIF, the basic morphology
286 structure of the aerogel remains almost the original state. Fig. 3f is a magnified image
287 of $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL. It can be observed that the thickness of the pore walls had
288 increased several times, and the inside of the cellulose pore walls were divided into
289 many layers, and dense Zn/Co-ZIF nanoparticles were inserted between the layers to
290 form a sandwich structure. It was indicated that Zn/Co-ZIF@GEL has a pomegranate-
291 like morphology with GEL as peel and Zn/Co-ZIF nanoparticles as pomegranate seeds,
292 so that nanoparticles could be coated with aerogel and maintain its stability in reaction
293 solution.

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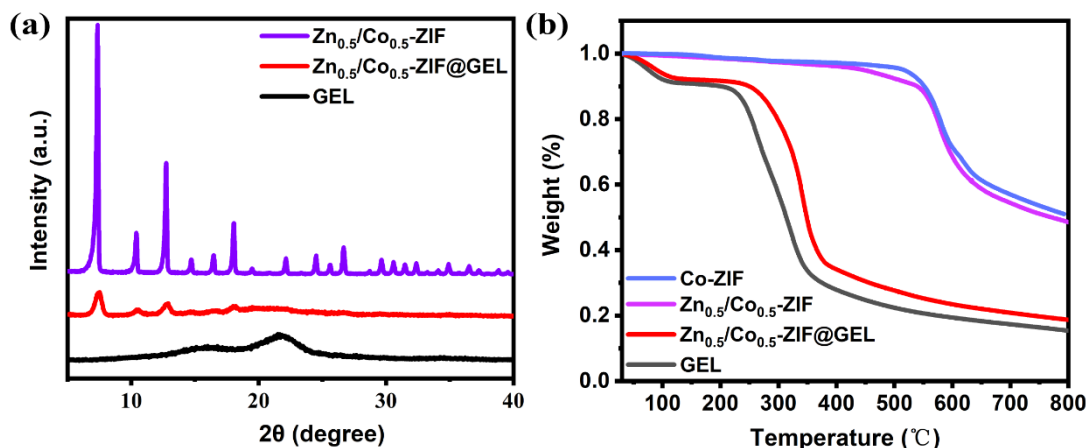
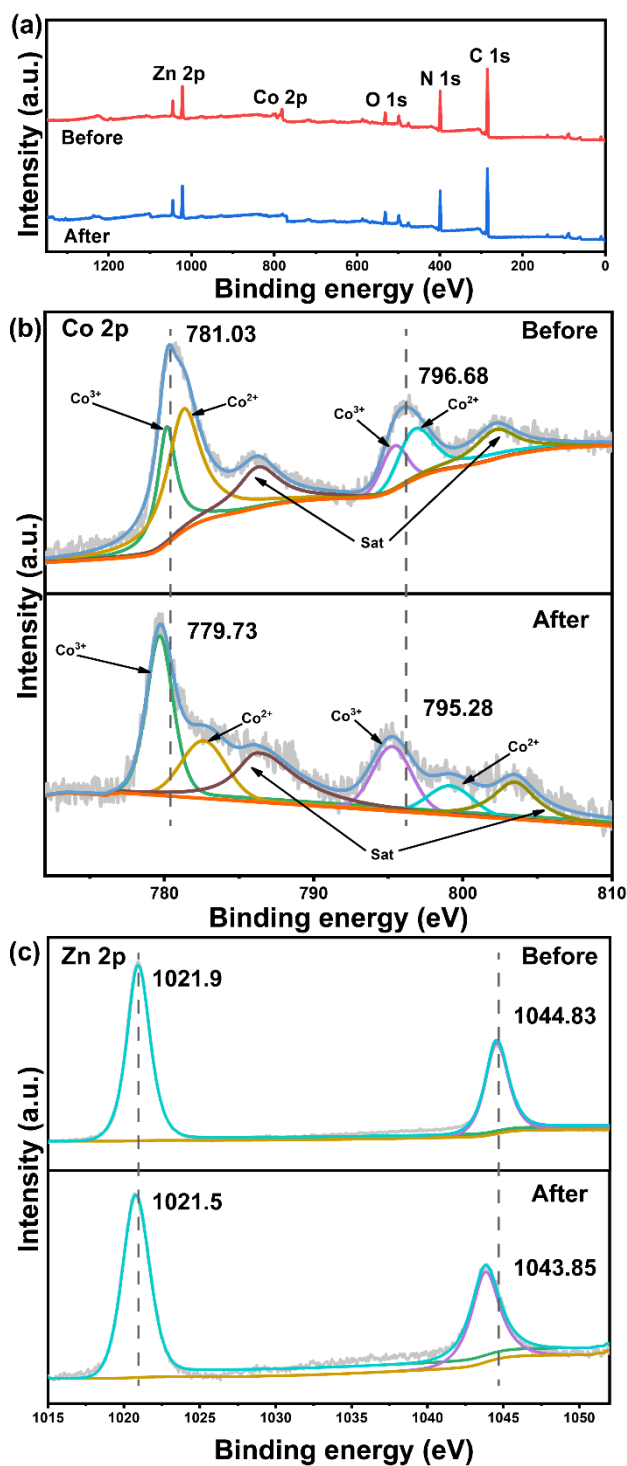


Fig. 4. (a) XRD image of Zn_{0.5}/Co_{0.5}-ZIF@GEL, (b) TG image of Zn_{0.5}/Co_{0.5}-ZIF@GEL.

The XRD pattern of GEL, Zn_{0.5}/Co_{0.5}-ZIF and Zn_{0.5}/Co_{0.5}-ZIF@GEL are shown in Fig. 4a. GEL had two weaker and wider diffraction peaks at 12.3 and 20.1°, and could be indexed to the cellulose crystal type II, which is different from the cellulose type I of parent bagasse (Fig. S2), suggesting the successful purification of cellulose for the GEL [45]. The sharp peaks of Zn_{0.5}/Co_{0.5}-ZIF were corresponding to the planes of (011), (002), (112) and (222), respectively, which was in accordance with that of the reported Co-ZIF [41]. The Zn_{0.5}/Co_{0.5}-ZIF@GEL simultaneously contained the main characteristic peaks of Zn/Co-ZIF and cellulose aerogel, which was further proved the successful combination of Zn_{0.5}/Co_{0.5}-ZIF and cellulose aerogel. The characteristic peak of Zn/Co-ZIF was inapparent, which may be because Zn/Co-ZIF was encapsulated in the composite aerogel.

The thermal stability of GEL, Zn_{0.5}/Co_{0.5}-ZIF and Zn_{0.5}/Co_{0.5}-ZIF@GEL was monitored by TGA, and the results are shown in Fig. 4b. The mass loss of GEL was significant at 250 ~ 380 °C, and the mass residual rate at 800 °C was 14.7%. For Zn_{0.5}/Co_{0.5}-ZIF, the mass residue rate reached 42.94% at 800 °C. The mass residue rate of Zn_{0.5}/Co_{0.5}-ZIF@GEL increased to 20% at 800 °C, indicating that Zn_{0.5}/Co_{0.5}-ZIF nanoparticles were successfully immobilized on the bagasse cellulose aerogel matrix, which improved the thermal stability of the composite.



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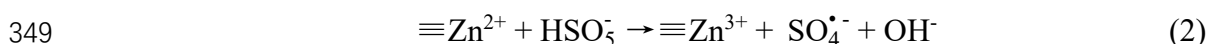
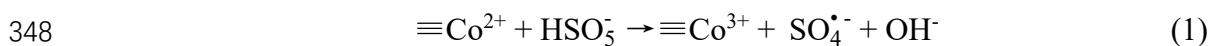
319 Fig. 5. XPS spectra of $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL (a) full-range scan of the samples, (b) Co 2p core
 320 level, (c) Zn 2p core level.

321

322 In order to further understand the pathway of $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL in the PMS
 323 activation process, X-ray photoelectron spectroscopy (XPS) was used to determine the

324 element valence states of the composite aerogel. The change of the chemical valence
 325 states of Co and Zn in the XPS reaction provides a possible catalytic way. As shown in
 326 Fig. 5a, the survey XPS spectrum displays characteristic peaks of Zn 2p, Co 2p, O 1s,
 327 N 1s, and C 1s, indicating the existence of the above elements. There was no obvious
 328 change on intensity and binding energy of each element in the spectra before and after
 329 use, indicating that Zn/Co-ZIF@GEL has good stability. Fig. 5b shows the XPS
 330 spectrum of Co 2p of fresh and used composite aerogels. It can be seen that the two
 331 main peaks of Co^{3+} 2p $3/2$ and Co^{3+} 2p $1/2$ located at 780.2 eV and 794.9 eV,
 332 respectively. The other two peaks, 781.9 eV and 796.3 eV, were derived from Co^{2+} 2p
 333 $3/2$ and Co^{2+} 2p $1/2$. And two satellite peaks were accompanied the main peaks at the
 334 same time. After the catalytic reaction, the proportion of Co^{2+} dropped from 43.6% to
 335 35.9%, and 23.9% of Co^{3+} was formed, indicating that Co^{2+} and Co^{3+} involved in the
 336 PMS activation. It showed that the electron transfer occurred when Co participated in
 337 the reaction process. As such, Co^{2+} provided electrons to PMS, then activated PMS to
 338 produce $\text{SO}_4^{\bullet-}$. At the same time, the valence of Co has also changed from +2 to +3. The
 339 reaction formula is as Eq. (1).

340 Fig. 5c displays the spectrum of Zn 2p, the peak value of Zn 2p before used was
 341 at 1021.9 eV. After used, it shifted slightly and moved to 1021.5 eV. And after the
 342 catalytic reaction occurred, the peak area was slightly reduced. It showed that part of
 343 the Zn^{2+} on the surface of the Zn/Co-ZIF@GEL catalyst has been transformed into Zn^{3+}
 344 during the reaction Eq. (2). However, XPS spectrum was not recognized Zn^{3+} . Since it
 345 was found that there was no thermodynamically stable Zn^{3+} compound in recent
 346 calculations, and Zn^{3+} must be transformed into a stable Zn^{2+} state immediately [46,
 347 47].



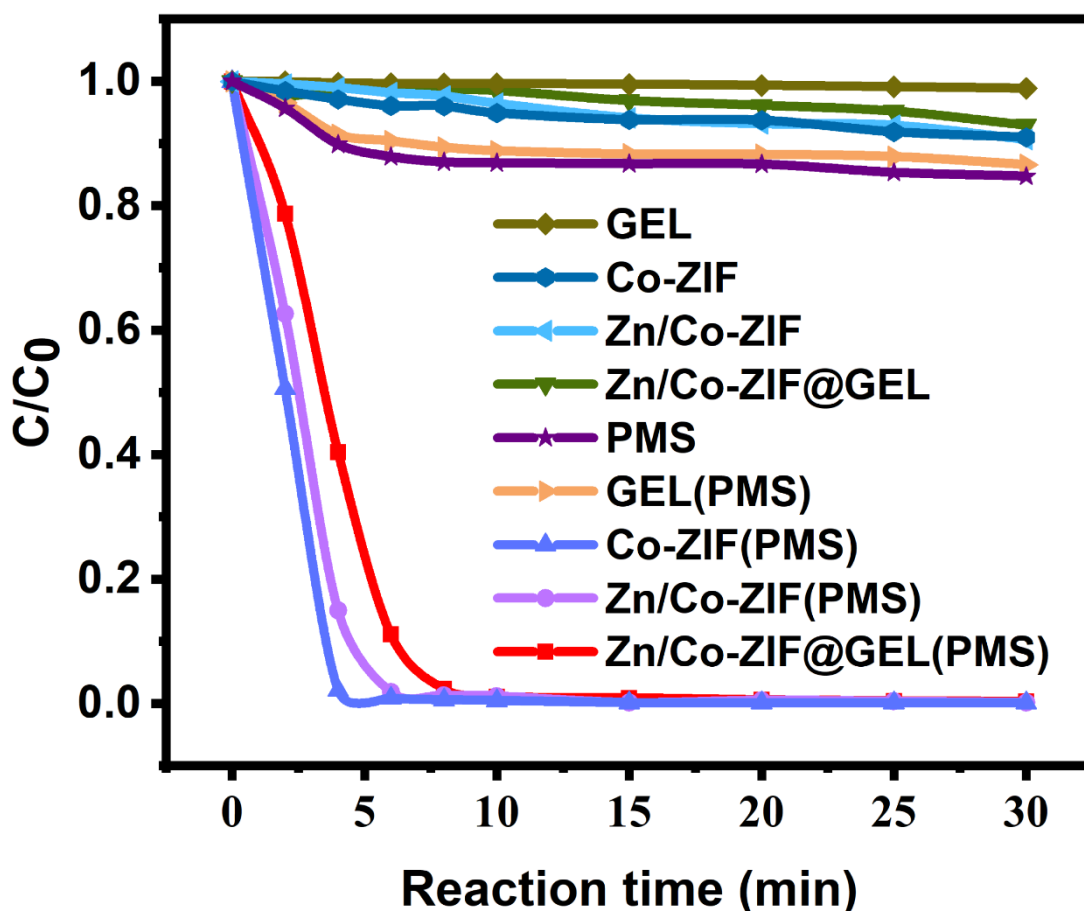
350 It can be inferred that PMS was activated by Co^{2+} and Zn^{2+} on the surface of
 351 Zn/Co-ZIF to produce $\text{SO}_4^{\bullet-}$. All these findings suggested that Co^{3+} and Zn^{3+} continued
 352 to complete the redox cycle, so that the Zn/Co-ZIF@GEL catalyst can work

353 continuously. And the redox cycles occurred on the surface of Zn/Co-ZIF@GEL made
 354 it maintain outstanding catalytic performance [42]. The above reaction formulas are as
 355 Eq. (3-4).



358

359 3.4. Performance of catalytic degradation by PMS activation



360

361 Fig. 6. Degradation performance of GEL, Co-ZIF, Zn_{0.5}/Co_{0.5}-ZIF and Zn_{0.5}/Co_{0.5}-ZIF@GEL for
 362 RhB in PMS system. (experiment conditions: [RhB] = 50 mg/L, [PMS] = 0.4mM, [Co-ZIF] =
 363 50mg/L, [Zn_{0.5}/Co_{0.5}-ZIF] = 50mg/L, [Zn_{0.5}/Co_{0.5}-ZIF@GEL] = 150 mg/L, pH = 6.8, T = 25 °C)

364

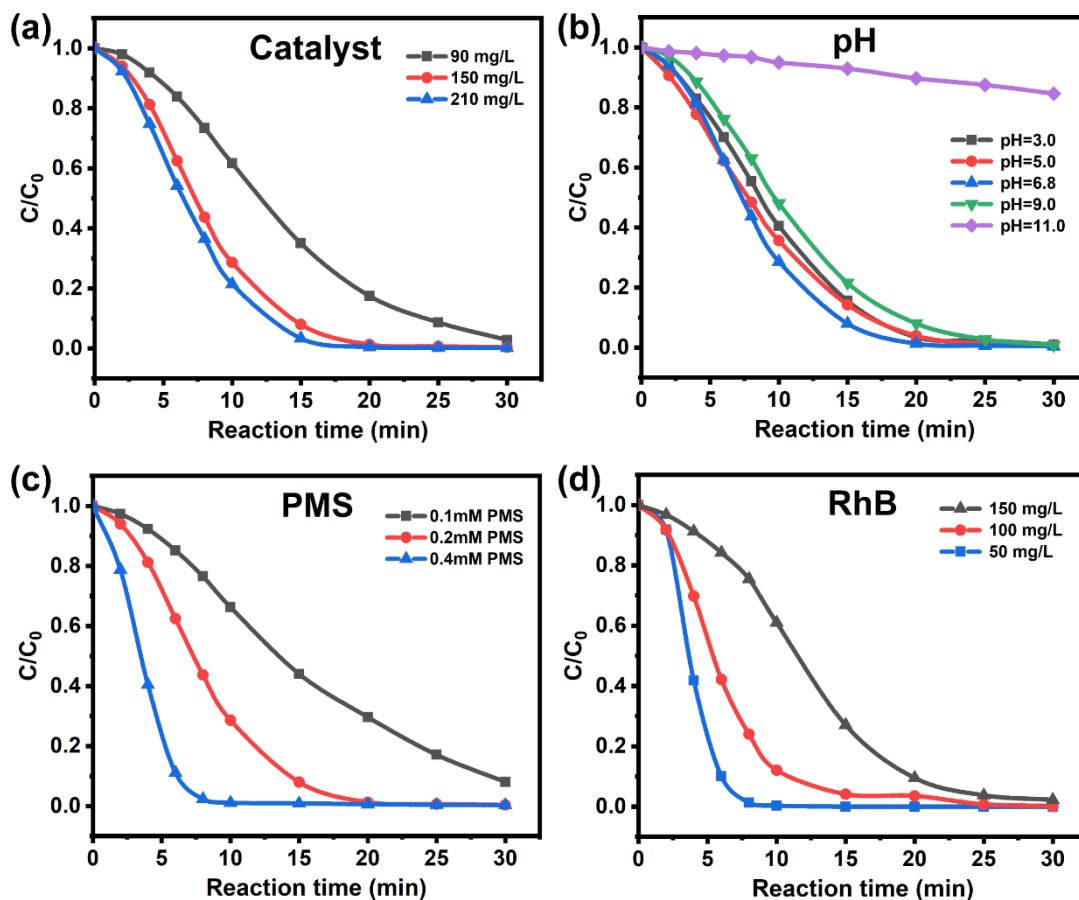
365 The catalytic performances of different catalysts were evaluated through PMS
 366 activated degradation of cationic dye RhB. Before catalytic reaction, the RhB
 367 adsorption performance of GEL, Co-ZIF, Zn_{0.5}/Co_{0.5}-ZIF and Zn_{0.5}/Co_{0.5}-ZIF@GEL

368 were investigated without adding PMS. As shown in Fig. 6, negligible removal
369 efficiency of RhB was observed in GEL, which was associated to the macroporous
370 feature and ultralow surface area of GEL (Fig. 3e). In comparison to GEL, Co-ZIF、
371 $Zn_{0.5}/Co_{0.5}$ -ZIF and $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL showed slightly higher removal amount of
372 RhB (10, 10 and 8%, respectively). However, this results still indicated the limitation
373 of adsorption removal of RhB by using GEL, Co-ZIF, $Zn_{0.5}/Co_{0.5}$ -ZIF and $Zn_{0.5}/Co_{0.5}$ -
374 ZIF@GEL. Thus, it is necessary to stimulate the catalytic reaction to achieve high
375 removal performance of RhB on as-synthesized catalysts.

376 PMS was applied to the systems and activated by the catalysts. Degradation
377 experiments of RhB were carried out with PMS alone, GEL with PMS, Co-ZIF with
378 PMS, $Zn_{0.5}/Co_{0.5}$ -ZIF with PMS and $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL with PMS. As showed in
379 Fig. 6, the degradation curves of PMS alone and GEL with PMS were almost identical,
380 which both have 15% degradation efficiency of RhB. This suggested the limited
381 oxidation activity of PMS without activation and inertness of GEL. Once the as-
382 prepared catalysts applied into the system with PMS, significant enhancement in
383 degradation efficiency of RhB were observed. All catalysts (Co-ZIF, $Zn_{0.5}/Co_{0.5}$ -ZIF
384 and $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL) with PMS showed similar ultimate degradation efficiency
385 of 100% within 10 min but different kinetic rates of reaction. As depicted in Fig. S3,
386 the kinetic rates of Co-ZIF, $Zn_{0.5}/Co_{0.5}$ -ZIF and $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL were calculated
387 to be 0.309, 0.066 and 0.059 min^{-1} . The Compared with the degradation rate of
388 $Zn_{0.5}/Co_{0.5}$ -ZIF and $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL were relative sluggish to Co-ZIF, which was
389 probably the inertness of doping metal of Zn.

390

391 3.5. Effect of catalyst dosage, initial pH, PMS concentration and initial RhB
 392 concentration on catalytic activity

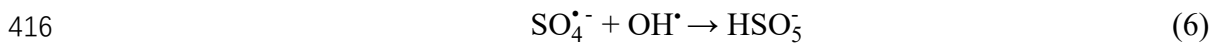
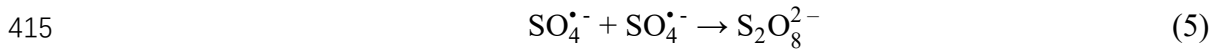


393
 394 Fig. 7. The effect of different conditions of RhB removal rate: (a) catalyst dosage, (b) initial
 395 solution pH, (c) PMS concentration, (d) initial concentration of RhB. (experiment conditions: [RhB]
 396 = 50 mg/L (for a, b, c), [PMS] = 0.2 mM (for a, b, d), [catalyst] = 150 mg/L (for b, c, d), pH = 6.8
 397 (for a, c, d), T=25°C).

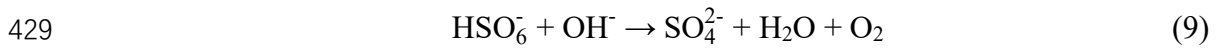
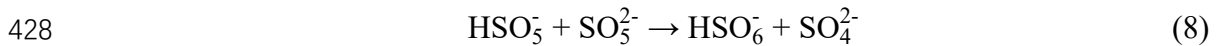
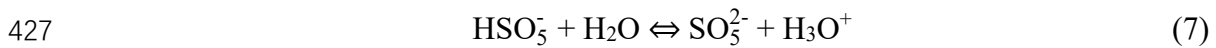
398

399 The effect of operation parameters (catalyst dosage, initial pH, PMS concentration
 400 and initial RhB concentration) on the degradation performance of RhB in of Zn_{0.5}/Co_{0.5}-
 401 ZIF@GEL/PMS was explored and shown in Fig. 7(a-d). The degradation results of RhB
 402 using different dosage of catalysts from 90-210 mg/mL were shown in Fig. 7a. The
 403 degradation percentage could be achieved to 98% within 30 min for 90 mg/L of
 404 Zn_{0.5}/Co_{0.5}-ZIF@GEL. As the catalyst dosage added, the significant increase in the
 405 degradation efficiency was observed. Compared with the case of catalyst dosage at 90
 406 mg/mL, higher removal percentage (100%) for RhB within shorter time (20 min) was
 407 observed in case of catalyst dosage at 150 and 210 mg/mL. This was attributed to the

408 increase of catalytic activity sites, which is favorable for promoting the activation of
 409 PMS and generating higher amounts of radicals. It was noted that the degradation rate
 410 increases rapidly when the catalyst dosage increased from 90 to 150 mg/mL, but when
 411 the catalyst dosage increased from 150 mg/mL to 210 mg/mL, the degradation rate was
 412 slow and almost approaching the equilibrium of the catalytic reaction. This indicated
 413 that the self-quenching of free radicals may occur at high concentration (Eq. (5) and
 414 (6)), thus limiting the increase of the reaction rate [48].



417 Initial pH is one of the critical parameters that could affect the catalytic
 418 performance. Fig. 7b depicted the results of RhB degradation in the catalytic systems
 419 with different initial pH ranging from 3-11. Obviously, in the wide pH range of 3.0 ~
 420 9.0, pH had no obvious effect on the degradation effect of RhB, indicated that the
 421 Zn/Co-ZIF@GEL/PMS system has good pH resistance. However, when the pH reached
 422 11, the performance decreased significantly, which may be due to the formation of Co
 423 $(\text{OH})_2$ with low reactivity on the surface of the catalyst and the conversion of a large
 424 amount of $\text{SO}_4^{\bullet-}$ to OH^- . Under alkaline conditions (initial pH = 11), PMS may undergo
 425 self-degradation without producing $\text{SO}_4^{\bullet-}$, and the following reaction occurs (Eq. (7)-
 426 (10)), thereby reducing the degradation ability of RhB.

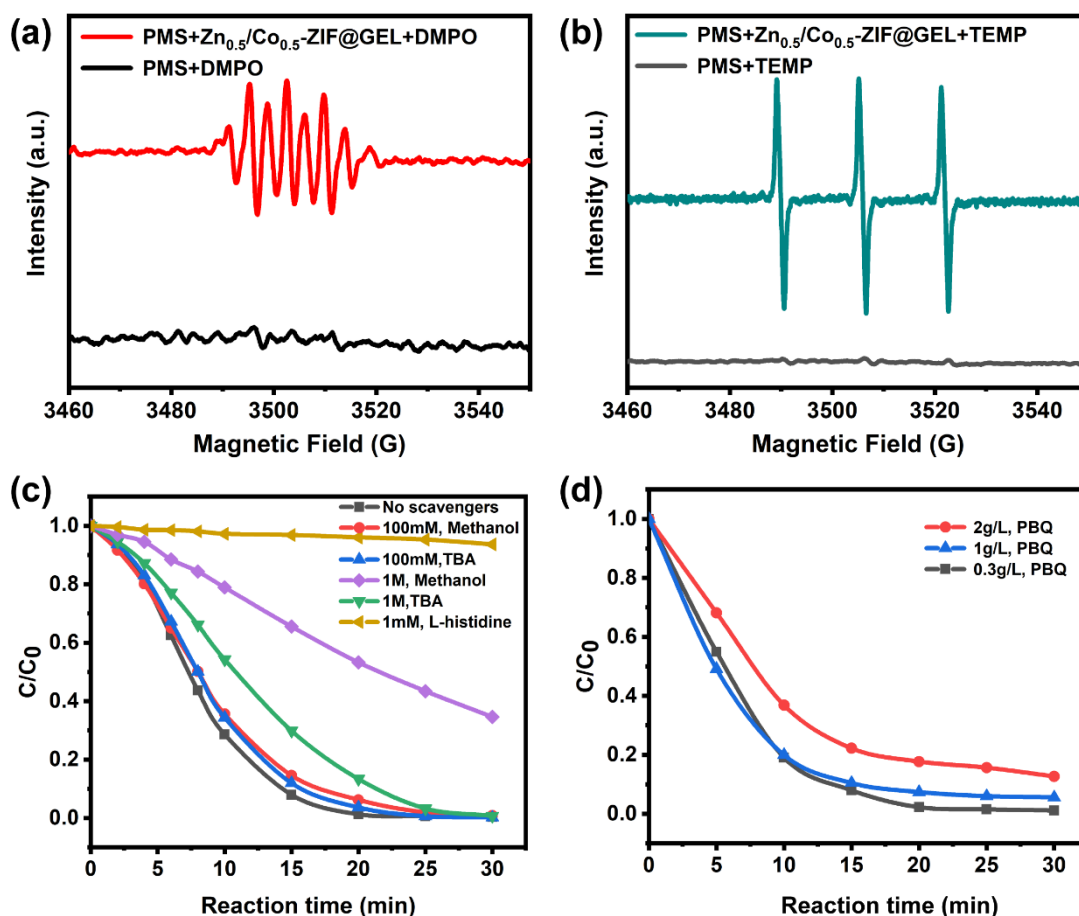


431 The effect of PMS concentration on RhB degradation was shown in Fig. 7c. PMS
 432 as the source of OH^- and $\text{SO}_4^{\bullet-}$ [49], its content has a critical effect on the reaction. The
 433 results showed that the PMS concentration from 0.1 mM to 0.4 mM has a significant
 434 effect on the catalytic performance. The catalytic performance was not satisfactory
 435 when the PMS concentration (0.1mM) was insufficient, which was caused by the rapid

436 consumption of PMS by $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL. ~~When the PMS amount has reached to~~
437 ~~0.4mM, the effect of degrading RhB was inhibited. The reason might be once the~~
438 ~~amount of PMS increase in the solution, the acidity and alkalinity of the solution was~~
439 ~~destroyed and caused free radical quenching, thereby the activity of catalyst was~~
440 ~~inhibited [50].~~ When the PMS concentration has reached to 0.4 mM, the effect of
441 degrading RhB has increased significantly. High concentrations of PMS generated more
442 free radicals and provided more active sites for the catalyst. At this point the reaction
443 system achieved the highest reaction rate and the complete degradation rate.

444 Fig. 7d shows the degradation efficiency of RhB at different initial concentrations,
445 and further investigates the catalytic activity of $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL. It clearly found
446 that the initial concentration of 50 mg/L and 100 mg/L initial concentration achieved
447 completely removal within 25 min. Moreover, RhB with high concentrations had a
448 slight decline removal in the same reaction time, 97.6% degradation effect was reached
449 at 150 mg/L in 30 min. This could be attributed to when the concentration of RhB goes
450 up, the concentration difference between RhB and radicals increased, resulting in the
451 slower reaction rate [51].

452



454

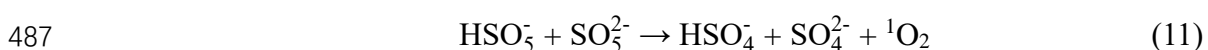
455 Fig. 8. (a)(b) EPR spectra of DMPO and TEMP as the trapping agent, respectively, (c)(d)
 456 Effects of different free radical scavengers on the degradation of RhB in Zn_{0.5}/Co_{0.5}-ZIF@GEL/PMS
 457 system. (experiment conditions: [RhB] = 50 mg/L, [PMS] = 0.2mM, [catalyst] = 150 mg/L, pH =
 458 6.8, T = 25 °C).

459

460 In the process of PMS activation by transition metals, SO₄^{•-} and OH[•] are
 461 simultaneously produced [22]. DMPO was used as the spin-trapping agent for SO₄^{•-} and
 462 OH[•], and EPR spectroscopy of the radical was identified. As shown in Fig. 8a, when
 463 PMS was participated in the reaction solution, only a weak signal appeared which was
 464 caused by the self-decomposition of PMS in the absence of catalyst. It was worth noting
 465 that there were no distinctive signals of DMPO-SO₄^{•-} and DMPO-OH[•] appear in the
 466 Zn/Co-ZIF@GEL system, but clear seven peaks signal ($\alpha_N = 7.3 \pm 0.1$ G and $\alpha_H = 3.9$
 467 ± 0.1 G) [48]. The phenomenon was due to the hyperfine splitting of DMPO oxidation

468 products, which were 5,5-dimethylpyrrolidone-2-(oxy)-(1) or 5,5-dimethyl-2-
469 pyrrolidone-1-oxyl (DMPO-X) [52, 53]. The appearance of DMPO-X meant that highly
470 active oxygen-containing molecules were produced during the activation reaction of
471 PMS, which leads to the rapid oxidation of DMPO [54]. The generated oxidized
472 molecules had strong oxidizing ability and were difficult to capture by adding DMPO.
473 On the one hand, the disproportionation reaction of superoxide radicals in water may
474 inhibit the reaction between superoxide radicals and DMPO. On the other hand, it may
475 be that the DMPO-O₂^{•-} signal was easily covered by the DMPO-X signal. And similar
476 phenomenon has been seen in previous studies[55, 56].

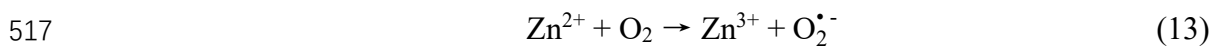
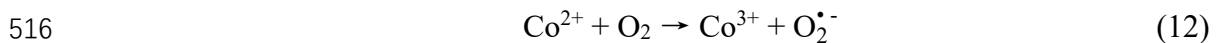
477 In order to identify other free radicals that may form, TEMP was used as the spin-
478 trapping agent for testing. As shown in Fig. 8b, it can be observed that PMS/TEMP can
479 form a weak triplet signal with an intensity ratio of 1:1:1, which may be caused by the
480 self-degradation of PMS to form ¹O₂ (Eq. (11)) [57]. However, in the Zn_{0.5}/Co_{0.5}-ZIF
481 @GEL/PMS system, the triplet high-intensity EPR spectral peak (αN = 16.9 g) was
482 observed, which was consistent with the TEMP-singlet oxygen ¹O₂ adduct (TEMP) [58],
483 which indicated that Zn/Co-ZIF can promote the self-degradation of PMS and produce
484 singlet oxygen. And it proved that ¹O₂ played indispensable role in degradation process
485 of Zn/Co-ZIF/PMS system. It can be recommended that the non-radical pathway (¹O₂)
486 and free radicals (SO₄^{•-}, OH[•] and O₂^{•-}) jointly act on the degradation of RhB.



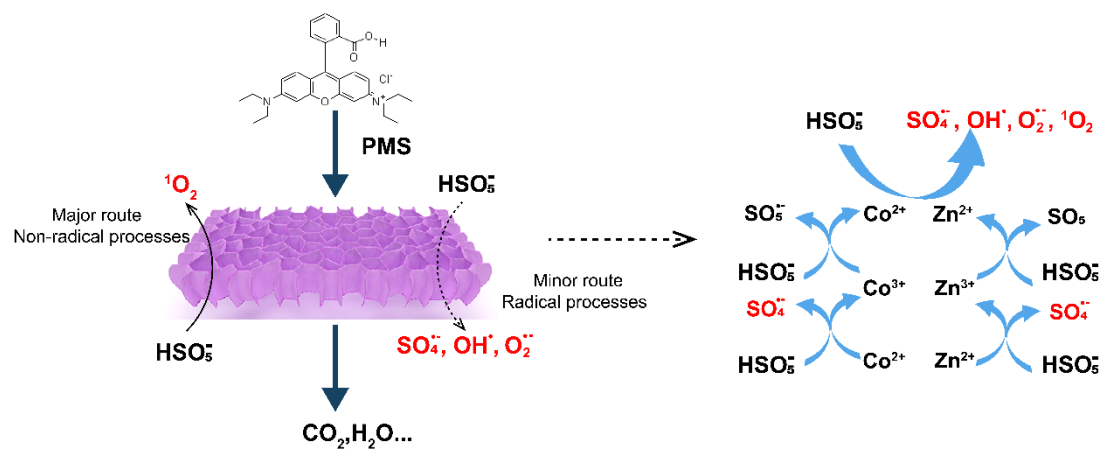
488 In order to further understand the degradation mechanism, the results of using
489 methanol, *tert*-butyl alcohol (TBA) and L-histidine as free radical scavengers are shown
490 in Fig. 8c. Methanol was regarded as a scavenger of sulfate radical (SO₄^{•-}) and hydroxyl
491 radical (OH[•]), TBA was a selective probe for OH[•], and benzoquinone (PBQ) was used
492 as a scavenger of O₂^{•-}, L-Histidine acted as a scavenger of singlet oxygen (¹O₂) [59].

493 When there was no scavenger, the removal efficiency of 98.7% can be achieved
494 after 20 min of reaction. When equal amounts (100 mM) of TBA and methanol were
495 added, the degradation of RhB was only slightly inhibited. After adding 1M TBA and
496 methanol, both of them showed more obvious inhibition effect, but the inhibition effect
497 of methanol was more significant than that of TBA, and the degradation effect could

498 reach 65.4%. The phenomenon can be attributed to the fact that both OH^\bullet and $\text{SO}_4^{\bullet-}$
 499 were produced and participate in the degradation of RhB, but $\text{SO}_4^{\bullet-}$ played a more
 500 important role than OH^\bullet . But this meant that there were some other common free
 501 radicals or non-radical reactive molecules involved in the reaction. The result of adding
 502 PBQ into the solution is shown in Fig. 8d. When different amount of PBQ was added
 503 (0.3 g/L, 1 g/L, and 2 g/L), the degradation effect of RHB decreased from 98.8% to
 504 94.4 %, 87.27 %, respectively. Therefore, we inferred that a small amount of $\text{O}_2^{\bullet-}$ was
 505 produced by the oxidation of Co^{2+} and Zn^{2+} by O_2 in the reaction process (Eq. (12) and
 506 (13)) [50], but $\text{O}_2^{\bullet-}$ had limited degradation of RhB. Since $^1\text{O}_2$ was regarded as a typical
 507 active oxidation molecule (ROS) in non-free radical processes [60], we speculated that
 508 $^1\text{O}_2$ may exist in the reaction system of Zn/Co-ZIF@GEL/PMS. Therefore, when 1 mM
 509 L-histidine was added into the reaction solution (Fig. 8c), the degradation of RhB by
 510 Zn/Co-ZIF@GEL/PMS system was significantly inhibited, and only 9.9% of RhB was
 511 removed. It indicated that the non-radical ($^1\text{O}_2$) degradation route was the main
 512 degradation route of RhB. In combination with the previous results, it can be
 513 determined that in the Zn/Co-ZIF@GEL/PMS system, the non-free radical pathway
 514 ($^1\text{O}_2$) and free radicals ($\text{SO}_4^{\bullet-}$, OH^\bullet and $\text{O}_2^{\bullet-}$) interacted together to achieve the
 515 degradation effect of RhB.



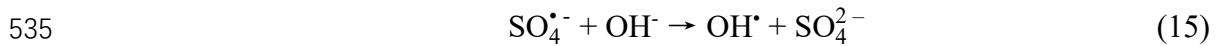
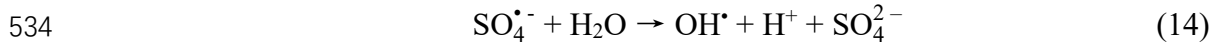
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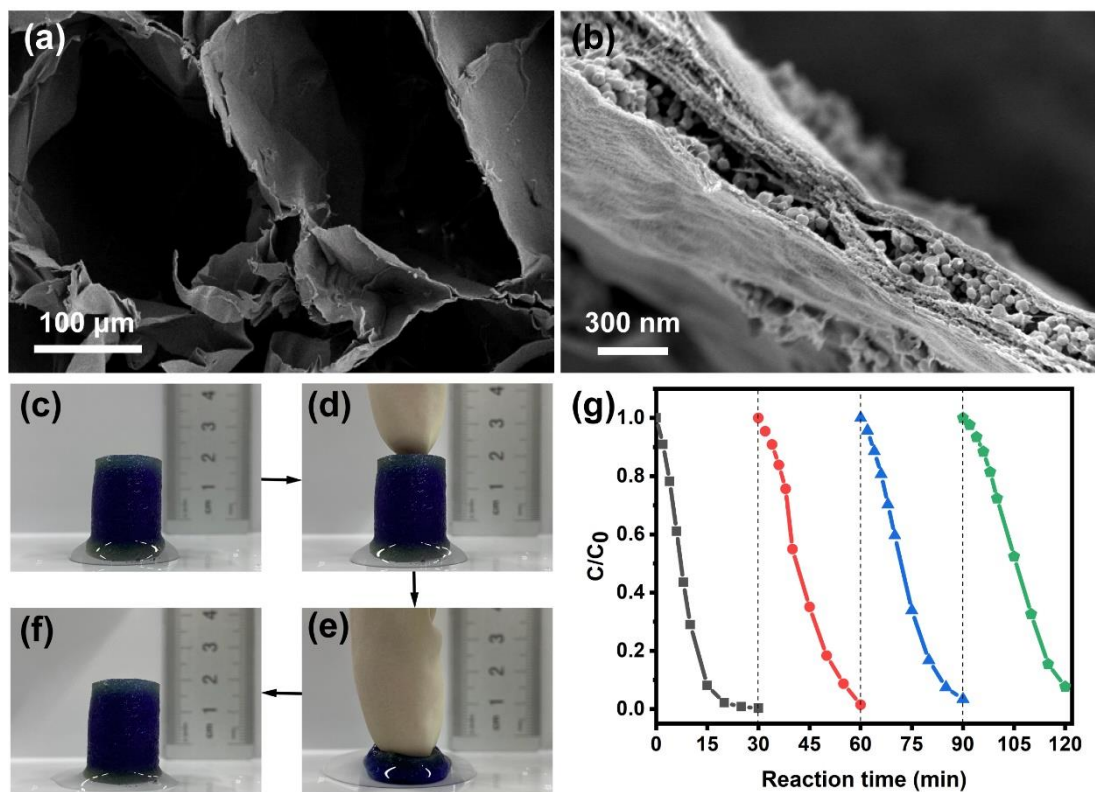
520 Fig. 9. The degradation mechanisms of RhB in Zn/Co-ZIF@GEL/PMS system.

521 In summary, based on the above results and previous studies, a possible
522 mechanism for the efficient activation of PMS by Zn/Co-ZIF@GEL was proposed (Fig.
523 9). First, the HSO_5^- in PMS could be activated by Co^{2+} and Zn^{2+} ions in Zn/Co-
524 ZIF@GEL, electron transfer occurred to produce $\text{SO}_4^{\bullet-}$, and the valence of Co and Zn
525 also converted +2 into +3 (Eq. (1) and (2)). Moreover, Co^{3+} and Zn^{3+} continue to react
526 with HSO_5^- to produce Co^{2+} and Zn^{2+} when the reaction goes on (Eq. (3) and (4)), so
527 that the effective redox cycle was maintained. Also, $\text{O}_2^{\bullet-}$ and OH^\bullet were generated during
528 this process, which could contribute to the degradation of RhB (Eq. (12-15)). In addition,
529 we have confirmed that the non-radical degradation mechanism of Zn-
530 Co/ZIF@GEL/PMS system was dominated by $^1\text{O}_2$ which produced by self-degradation
531 of PMS (Eq. 11). Besides, aerogel had a large specific surface area and high porosity,
532 which was conducive to the retention of RhB molecules, so as to make the degradation
533 of pollutants more efficient.



536

537 3.7. Structural and composition stability of $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL under PMS
538 activation catalysis



539
540 Fig. 10. (a)(b) SEM image of $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL after 4 cycles, (c-f) The process of
541 restoring the water-containing composite aerogel to its original state after being compressed, (g)
542 Recycling ability for RhB removal by $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL.

543
544 $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL had demonstrated its high efficiency as a heterogeneous
545 catalyst to activate PMS, but for the long-term use of catalyst, the evaluation of its
546 recyclability was necessary. In order to investigate the reusability and stability of
547 $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL in PMS activation, four consecutive parallel degradation
548 experiments of RhB were carried out under the same reaction conditions. Fig. 10a and
549 b are SEM images after four cycles of Zn/Co -ZIF@GEL. It can be seen that bagasse
550 cellulose composite aerogel still maintained the porous structure and the aerogel
551 skeleton was not damaged. With the increase of the number of cycles, the
552 microstructure of the Zn/Co -ZIF nanoparticles contained in the pore wall of cellulose
553 was not destroyed, and a large amount of Zn/Co -ZIF was still retained in the pore wall,
554 which could be reused continuously. Zn/Co -ZIF nanoparticles were encapsulated in

555 aerogel matrix and this type of encapsulation structure will minimize the shedding of
556 nanoparticles during use. The elemental mapping with before and after using Zn/Co-
557 ZIF@GEL demonstrated that Co and Zn were uniformly distributed in the composite
558 aerogel before and after use (Fig. S4).

559 The influence of PMS concentration on the dissolution of cobalt ions in Co-ZIF,
560 Zn_{0.5}/Co_{0.5}-ZIF and Zn_{0.5}/Co_{0.5}-ZIF@GEL is shown in Fig. S5. At different additions of
561 PMS concentrations, the cobalt ions dissolution of Zn/Co-ZIF@GEL was much lower
562 than that of Co-ZIF and Zn/Co-ZIF. The content of Co loaded on the composite aerogel
563 was approximately 4.6 wt%. Compared to the dissolved content of Co, ~~This indicated~~
564 ~~that~~ bagasse cellulose aerogel had a good protective effect on cobalt ions in the reaction
565 system. In addition, the dissolution amounts of cobalt ions of Co-ZIF in different PMS
566 concentrations was about three times that of Zn/Co-ZIF. Implying that the addition of
567 Zn²⁺ can effectively improve the chemical stability of ZIFs and reduce the dissolution
568 amount of Co²⁺, so as to ensure the long-term stable use of Zn/Co-ZIF@GEL catalyst.
569 In addition, when the PMS concentration was 0.4 mM, the cobalt ions were all dissolved
570 more than 1 mg/L. This was because with the increase of PMS concentration, the pH in
571 the reaction system continued to decrease, led to a decrease in the stability of Zn/Co-
572 ZIF. At the same time, the free radicals produced by catalyzing PMS will also cause a
573 certain amount of damage to Zn/Co-ZIF and Zn/Co-ZIF@GEL, leading to an increase
574 in the dissolution of cobalt ions.

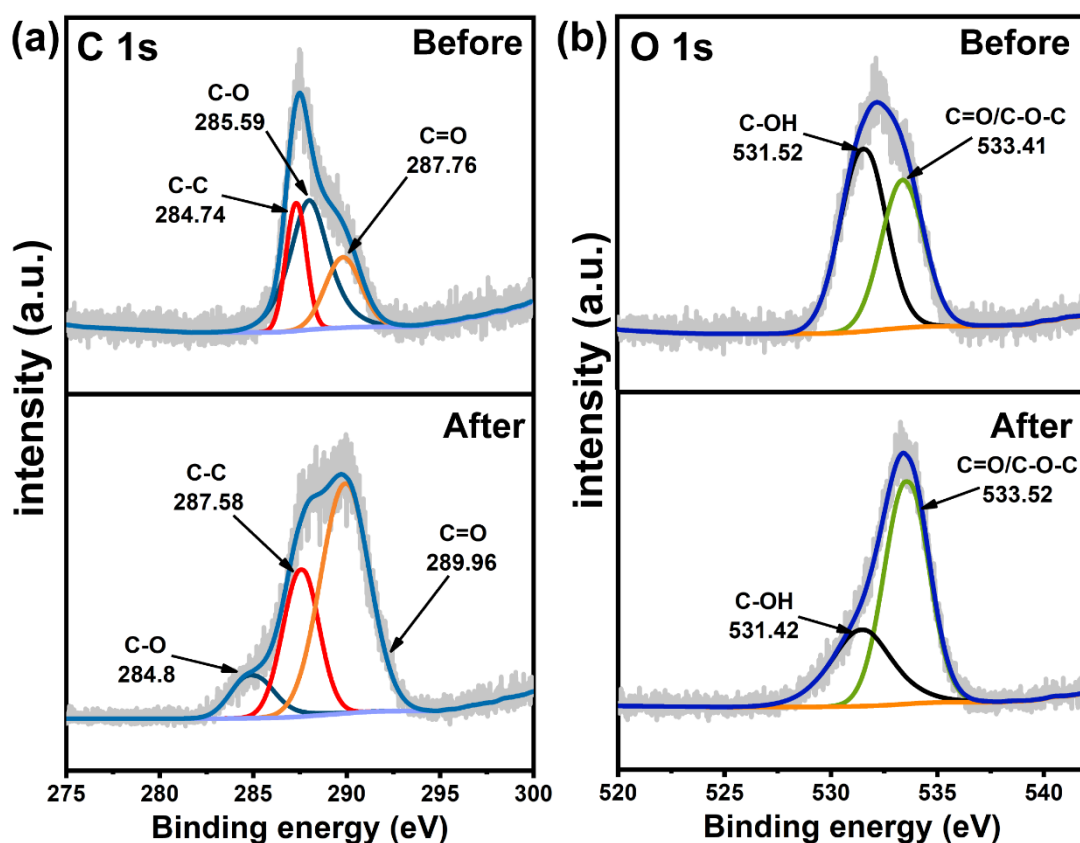
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576 3.8. Reusability of Zn/Co-ZIF@GEL

577 Fig. 10(c-f) are the process of the water-containing composite aerogel returning to
578 its original state after being completely compressed. First, pressure applied to the water-
579 containing composite aerogel to make it completely compressed, and the water
580 adsorbed in the composite aerogel was extruded. The pressure then removed and the
581 composite aerogel began to absorb water and fully returned to its original shape.
582 ~~However,~~ It indicated that the composite aerogel can be easily removed from water and
583 reused without changing the aerogel morphology, which indicated that the composite
584 aerogel has good reusability.

585 Catalyst with good stability and reusability were strongly expected in practical
 586 application. Therefore, catalytic recycling measurements of $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL
 587 were conducted and the results were showed in Fig. 10g. After four-cycle test, the
 588 catalytic degradation efficiency for RhB in $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL/PMS system slightly
 589 decreased but maintained high reusability. The slight decrease of catalytic efficiency
 590 might be due to the influence of the change of active site and surface coverage on the
 591 intermediates produced during the reaction, but the $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL after the
 592 reaction still has a high electron transfer efficiency so that has a good reusability [61].
 593

594 *3.9. The role of aerogel in reaction system*



595
 596 Fig. 11. XPS spectra of $Zn_{0.5}/Co_{0.5}$ -ZIF@GEL (a) C 1s core level, (b) O 1s core level.
 597

598 In order to detect the role of aerogel in the activation of PMS, XPS was further
 599 used to study the reaction process of composite aerogel. Fig. 11 shows the core level
 600 XPS spectra for C 1s and O 1s. In the C1s spectrum of fresh Zn/Co -ZIF@GEL (Fig.
 601 11a), the peaks at 284.74 eV, 285.59 eV and 287.76 eV belonged to C-C, C-O and C=O,

602 respectively, which were mainly derived from cellulose. Compared with fresh C 1s, the
603 proportion of C=O increased from 21.9% to 46.9%, C-C was shifted from 284.74 eV to
604 287.58 eV, C-O dropped from 39.8% to 16.4% after PMS activation. This indicated
605 that some of C-O on the aerogel surface were reduced to C=O, which could be attributed
606 to the interaction with Zn/Co-ZIF@GEL and PMS [62, 63]. As shown in Fig. 11b, for
607 the unused XPS spectrum of O 1s, the binding energy of C-OH appeared at 531.52eV,
608 and the peak of C=O/C-O-C was at 533.41 eV. It can be observed that after the reaction,
609 the ratio of C=O/C-O-C increased from 42.1% to 69%, while the ratio of C-OH
610 decreased by 26.9%. This indicated that oxygen actively participated in the catalytic
611 oxidation reaction which means C-OH might interact with PMS during the degradation
612 process, and the increase of C=O/C-O-C proved that cellulose aerogel was involved in
613 the process of protecting Zn/Co-ZIF [64]. The changes in the strength and binding
614 energy of elements before and after the reaction indicated that the aerogel was involved
615 in the activation of PMS, and the surface of the aerogel was oxidized to produce a large
616 amount of strong oxidizing substances, which further participate in the reaction. This
617 was also the reason why aerogel wrapped Zn/Co-ZIF without affecting its degradation
618 efficiency.

619 **4. Conclusion**

620 In this study, Zn/Co-ZIF@GEL was successfully prepared by doping method as a novel
621 and heterogeneous catalyst, the effect of Zn/Co-ZIF@GEL activated PMS for the
622 removal RhB was studied systematically. Different reaction conditions were applied to
623 degrade RhB and Zn/Co-ZIF@GEL/PMS system exhibited an excellent degradation
624 performance. Almost all RhB (50 mg/L) was degraded in a short time (10 min) and a
625 wide range of pH (3-9). Aerogel was successfully converted from sugarcane bagasse
626 and used as catalyst support, which could effectively reduce the leaching of cobalt ions
627 so that avoid secondary pollution. It was found that Zn doping Co-ZIF showed excellent
628 water stability, the reduction of Co content was not affected degradation performance
629 of ZIFs. In addition, radical-quenching tests and EPR measurements manifest that non-
630 radical pathway play a major role for RhB degradation, radical pathways were minor

631 pathways. Moreover, based on the cycle tests, Zn/Co-ZIF@GEL remains good stability
632 with 92.3% RhB removal rate after four successive cycles of reaction. The results
633 illustrated that the as-prepared catalyst performed excellent PMS activation and
634 maintained its stability and reaction activity for a long time in the reaction solution. As
635 a harmless and efficient organic pollutant treatment process, Zn/Co-ZIF@GEL
636 provided new insights for the application of new green aerogels in wastewater treatment.

637 References:

- 638 1. Otoni, C.G., et al., *The Food-Materials Nexus: Next Generation Bioplastics and*
639 *Advanced Materials from Agri-Food Residues*. Adv Mater, 2021: p. e2102520.
- 640 2. Qian, Y., F. Zhang, and H. Pang, *A Review of MOFs and Their Composites-*
641 *Based Photocatalysts: Synthesis and Applications*. Advanced Functional
642 Materials, 2021. **31**(37).
- 643 3. Thai, Q.B., et al., *Cellulose-based aerogels from sugarcane bagasse for oil spill-*
644 *cleaning and heat insulation applications*. Carbohydr Polym, 2020. **228**: p.
645 115365.
- 646 4. Teixeira, S.R., et al., *Valorization of sugarcane bagasse ash: producing glass-*
647 *ceramic materials*. J Environ Manage, 2014. **134**: p. 15-9.
- 648 5. Noor ul, A., et al., *Synthesis and characterization of geopolymer from bagasse*
649 *bottom ash, waste of sugar industries and naturally available china clay*.
650 Journal of Cleaner Production, 2016. **129**: p. 491-495.
- 651 6. Teixeira, S.R., et al., *Sugarcane Bagasse Ash as a Potential Quartz Replacement*
652 *in Red Ceramic*. Journal of the American Ceramic Society, 2008. **91**(6): p. 1883-
653 1887.
- 654 7. Patcharin, W., K. Sriamporn, and A. Kanokkan, *Utilization Biomass from*
655 *Bagasse Ash for Phillipsite Zeolite Synthesis*. Advanced Materials Research,
656 2011. **383-390**: p. 4038-4042.
- 657 8. Xu, Q., et al., *Characteristics and Applications of Sugar Cane Bagasse Ash*
658 *Waste in Cementitious Materials*. Materials (Basel), 2018. **12**(1).
- 659 9. Peiris, C., et al., *Biochar based removal of antibiotic sulfonamides and*
660 *tetracyclines in aquatic environments: A critical review*. Bioresour Technol,
661 2017. **246**: p. 150-159.
- 662 10. Zhu, H., Q. Zhang, and S. Zhu, *Alginate Hydrogel: A Shapeable and Versatile*
663 *Platform for in Situ Preparation of Metal-Organic Framework-Polymer*
664 *Composites*. ACS Appl Mater Interfaces, 2016. **8**(27): p. 17395-401.
- 665 11. Yang, W., et al., *Shapeable three-dimensional CMC aerogels decorated with*
666 *Ni/Co-MOF for rapid and highly efficient tetracycline hydrochloride removal*.
667 Chemical Engineering Journal, 2019. **375**.
- 668 12. Moon, R.J., et al., *Cellulose nanomaterials review: structure, properties and*
669 *nanocomposites*. Chem Soc Rev, 2011. **40**(7): p. 3941-94.
- 670 13. Zhu, H., et al., *Flexible and Porous Nanocellulose Aerogels with High Loadings*
671 *of Metal-Organic-Framework Particles for Separations Applications*. Adv

- 672 Mater, 2016. **28**(35): p. 7652-7.
- 673 14. Liu, C., et al., *Novel carbon based Fe-Co oxides derived from Prussian blue*
674 *analogues activating peroxymonosulfate: Refractory drugs degradation without*
675 *metal leaching*. Chemical Engineering Journal, 2020. **379**.
- 676 15. Xin, S., et al., *High efficiency heterogeneous Fenton-like catalyst biochar*
677 *modified CuFeO₂ for the degradation of tetracycline: Economical synthesis,*
678 *catalytic performance and mechanism*. Applied Catalysis B: Environmental,
679 2021. **280**.
- 680 16. Li, Y., et al., *Trace pyrolyzed ZIF-67 loaded activated carbon pellets for*
681 *enhanced adsorption and catalytic degradation of Rhodamine B in water*.
682 Chemical Engineering Journal, 2019. **375**.
- 683 17. Chi, H., et al., *Ferrous metal-organic frameworks with stronger coordinatively*
684 *unsaturated metal sites for persulfate activation to effectively degrade dibutyl*
685 *phthalate in wastewater*. J Hazard Mater, 2019. **377**: p. 163-171.
- 686 18. Li, J., et al., *Enhancement of the degradation of atrazine through CoFe₂O₄*
687 *activated peroxymonosulfate (PMS) process: Kinetic, degradation*
688 *intermediates, and toxicity evaluation*. Chemical Engineering Journal, 2018.
689 **348**: p. 1012-1024.
- 690 19. Li, C., et al., *Tunable titanium metal-organic frameworks with infinite 1D Ti-O*
691 *rods for efficient visible-light-driven photocatalytic H₂ evolution*. Journal of
692 Materials Chemistry A, 2019. **7**(19): p. 11928-11933.
- 693 20. Zhou, C., et al., *Strategies for enhancing the perylene diimide photocatalytic*
694 *degradation activity: method, effect factor, and mechanism*. Environmental
695 Science: Nano, 2021. **8**(3): p. 602-618.
- 696 21. He, C., et al., *Rational design to manganese and oxygen co-doped polymeric*
697 *carbon nitride for efficient nonradical activation of peroxymonosulfate and the*
698 *mechanism insight*. Chemical Engineering Journal, 2022. **430**.
- 699 22. Ghanbari, F. and M. Moradi, *Application of peroxymonosulfate and its*
700 *activation methods for degradation of environmental organic pollutants: Review*.
701 Chemical Engineering Journal, 2017. **310**: p. 41-62.
- 702 23. Lin, K.-Y.A. and H.-A. Chang, *Zeolitic Imidazole Framework-67 (ZIF-67) as a*
703 *heterogeneous catalyst to activate peroxymonosulfate for degradation of*
704 *Rhodamine B in water*. Journal of the Taiwan Institute of Chemical Engineers,
705 2015. **53**: p. 40-45.
- 706 24. Hu, P. and M. Long, *Cobalt-catalyzed sulfate radical-based advanced oxidation:*
707 *A review on heterogeneous catalysts and applications*. Applied Catalysis B:
708 Environmental, 2016. **181**: p. 103-117.
- 709 25. Howarth, A.J., et al., *Chemical, thermal and mechanical stabilities of metal-*
710 *organic frameworks*. Nature Reviews Materials, 2016. **1**(3).
- 711 26. Fang, C., et al., *Co-Ferrocene MOF/Glucose Oxidase as Cascade Nanozyme*
712 *for Effective Tumor Therapy*. Advanced Functional Materials, 2020. **30**(16).
- 713 27. Hwang, J., et al., *Controlling the morphology of metal-organic frameworks and*
714 *porous carbon materials: metal oxides as primary architecture-directing agents*.
715 Chem Soc Rev, 2020. **49**(11): p. 3348-3422.

- 716 28. Sun, L., M.G. Campbell, and M. Dinca, *Electrically Conductive Porous Metal-*
717 *Organic Frameworks*. *Angew Chem Int Ed Engl*, 2016. **55**(11): p. 3566-79.
- 718 29. Ding, M., X. Cai, and H.L. Jiang, *Improving MOF stability: approaches and*
719 *applications*. *Chem Sci*, 2019. **10**(44): p. 10209-10230.
- 720 30. Jhung, S.H., et al., *Microwave Synthesis of Chromium Terephthalate MIL-101*
721 *and Its Benzene Sorption Ability*. *Advanced Materials*, 2007. **19**(1): p. 121-124.
- 722 31. Wang, Q. and D. Astruc, *State of the Art and Prospects in Metal-Organic*
723 *Framework (MOF)-Based and MOF-Derived Nanocatalysis*. *Chem Rev*, 2020.
724 **120**(2): p. 1438-1511.
- 725 32. Burtch, N.C., H. Jasuja, and K.S. Walton, *Water stability and adsorption in*
726 *metal-organic frameworks*. *Chem Rev*, 2014. **114**(20): p. 10575-612.
- 727 33. Gu, A., et al., *The preparation of Ag/ZIF-8@ZIF-67 core-shell composites as*
728 *excellent catalyst for degradation of the nitroaromatic compounds*. *Applied*
729 *Surface Science*, 2020. **516**.
- 730 34. Chen, B., et al., *Zeolitic imidazolate framework materials: recent progress in*
731 *synthesis and applications*. *J. Mater. Chem. A*, 2014. **2**(40): p. 16811-16831.
- 732 35. Liu, Y., et al., *Activation of persulfate with metal-organic framework-derived*
733 *nitrogen-doped porous Co@C nanoboxes for highly efficient p-Chloroaniline*
734 *removal*. *Chemical Engineering Journal*, 2019. **358**: p. 408-418.
- 735 36. Zhong, G., D. Liu, and J. Zhang, *The application of ZIF-67 and its derivatives:*
736 *adsorption, separation, electrochemistry and catalysts*. *Journal of Materials*
737 *Chemistry A*, 2018. **6**(5): p. 1887-1899.
- 738 37. Ma, W., et al., *Nitrogen, phosphorus, and sulfur tri-doped hollow carbon shells*
739 *derived from ZIF-67@poly (cyclotriphosphazene-co-4, 4'-sulfonyldiphenol) as*
740 *a robust catalyst of peroxymonosulfate activation for degradation of bisphenol*
741 *A*. *Carbon*, 2018. **137**: p. 291-303.
- 742 38. Wang, C., et al., *Applications of water stable metal-organic frameworks*. *Chem*
743 *Soc Rev*, 2016. **45**(18): p. 5107-34.
- 744 39. Ren, W., et al., *Recyclable metal-organic framework/cellulose aerogels for*
745 *activating peroxymonosulfate to degrade organic pollutants*. *Chemical*
746 *Engineering Journal*, 2018. **349**: p. 766-774.
- 747 40. Chen, Y.Z., et al., *From Bimetallic Metal-Organic Framework to Porous*
748 *Carbon: High Surface Area and Multicomponent Active Dopants for Excellent*
749 *Electrocatalysis*. *Adv Mater*, 2015. **27**(34): p. 5010-6.
- 750 41. Zhou, K., et al., *Characterization and properties of Zn/Co zeolitic imidazolate*
751 *frameworks vs. ZIF-8 and ZIF-67*. *Journal of Materials Chemistry A*, 2017. **5**(3):
752 p. 952-957.
- 753 42. Butova, V.V., et al., *Zn/Co ZIF family: MW synthesis, characterization and*
754 *stability upon halogen sorption*. *Polyhedron*, 2018. **154**: p. 457-464.
- 755 43. Taheri, M., T.G. Enge, and T. Tsuzuki, *Water stability of cobalt doped ZIF-8: a*
756 *quantitative study using optical analyses*. *Materials Today Chemistry*, 2020. **16**.
- 757 44. Li, H., et al., *Enhanced hydrostability in Ni-doped MOF-5*. *Inorg Chem*, 2012.
758 **51**(17): p. 9200-7.
- 759 45. Wang, H., et al., *Preparation of tough cellulose II nanofibers with high thermal*

- 760 *stability from wood*. Cellulose, 2014. **21**(3): p. 1505-1515.
- 761 46. Schlöder, T., M. Kaupp, and S.J.J.o.t.A.C.S. Riedel, *Can zinc really exist in its*
762 *oxidation state+ III?* 2012. **134**(29): p. 11977-11979.
- 763 47. Lin, H., et al., *Degradation of bisphenol A by activating peroxymonosulfate with*
764 *Mn_{0.6}Zn_{0.4}Fe₂O₄ fabricated from spent Zn-Mn alkaline batteries*. Chemical
765 Engineering Journal, 2019. **364**: p. 541-551.
- 766 48. Wu, Z., et al., *Core-shell magnetic Fe₃O₄@Zn/Co-ZIFs to activate*
767 *peroxymonosulfate for highly efficient degradation of carbamazepine*. Applied
768 Catalysis B: Environmental, 2020. **277**.
- 769 49. !!! INVALID CITATION !!! .
- 770 50. Luo, H., et al., *WS₂ as highly active co-catalyst for the regeneration of Fe(II) in*
771 *the advanced oxidation processes*. Chemosphere, 2021. **262**: p. 128067.
- 772 51. Zhang, M., et al., *Metal-organic framework derived Co₃O₄/C@SiO₂ yolk-*
773 *shell nanoreactors with enhanced catalytic performance*. Journal of Materials
774 Chemistry A, 2018. **6**(24): p. 11226-11235.
- 775 52. Ma, W., et al., *Non-radical-dominated catalytic degradation of bisphenol A by*
776 *ZIF-67 derived nitrogen-doped carbon nanotubes frameworks in the presence*
777 *of peroxymonosulfate*. Chemical Engineering Journal, 2018. **336**: p. 721-731.
- 778 53. Fontmorin, J.M., et al., *Stability of 5,5-dimethyl-1-pyrroline-N-oxide as a spin-*
779 *trap for quantification of hydroxyl radicals in processes based on Fenton*
780 *reaction*. Water Res, 2016. **99**: p. 24-32.
- 781 54. Zhang, M., et al., *Confined pyrolysis of metal-organic frameworks to N-doped*
782 *hierarchical carbon for non-radical dominated advanced oxidation processes*.
783 Journal of Materials Chemistry A, 2019. **7**(20): p. 12547-12555.
- 784 55. Qi, C., et al., *Activation of peroxymonosulfate by base: Implications for the*
785 *degradation of organic pollutants*. Chemosphere, 2016. **151**: p. 280-8.
- 786 56. Li, X., et al., *Single Cobalt Atoms Anchored on Porous N-Doped Graphene with*
787 *Dual Reaction Sites for Efficient Fenton-like Catalysis*. J Am Chem Soc, 2018.
788 **140**(39): p. 12469-12475.
- 789 57. Su, W., et al., *Visible light photocatalysis on praseodymium(III)-nitrate-*
790 *modified TiO₂ prepared by an ultrasound method*. Applied Catalysis B:
791 Environmental, 2008. **77**(3-4): p. 264-271.
- 792 58. Chen, C., et al., *Mn₃O₄ nanodots loaded g-C₃N₄ nanosheets for catalytic*
793 *membrane degradation of organic contaminants*. J Hazard Mater, 2020. **390**: p.
794 122146.
- 795 59. Di, J., et al., *Electrochemical activation combined with advanced oxidation on*
796 *NiCo₂O₄ nanoarray electrode for decomposition of Rhodamine B*. Journal of
797 Water Process Engineering, 2020. **37**.
- 798 60. Sun, H., et al., *Activation of peroxymonosulfate by nitrogen-functionalized*
799 *sludge carbon for efficient degradation of organic pollutants in water*. Bioresour
800 Technol, 2017. **241**: p. 244-251.
- 801 61. Ding, M., et al., *Synergistic Features of Superoxide Molecule Anchoring and*
802 *Charge Transfer on Two-Dimensional Ti₃C₂T_x MXene for Efficient*
803 *Peroxymonosulfate Activation*. ACS Appl Mater Interfaces, 2020. **12**(8): p.

- 804 9209-9218.
- 805 62. Wang, Y., et al., *Activation of peroxymonosulfate by carbonaceous oxygen*
806 *groups: experimental and density functional theory calculations*. Applied
807 Catalysis B: Environmental, 2016. **198**: p. 295-302.
- 808 63. Li, Z., et al., *Biochar-supported nanoscale zero-valent iron as an efficient*
809 *catalyst for organic degradation in groundwater*. J Hazard Mater, 2020. **383**: p.
810 121240.
- 811 64. Li, Z., et al., *Converting untreated waste office paper and chitosan into aerogel*
812 *adsorbent for the removal of heavy metal ions*. Carbohydr Polym, 2018. **193**: p.
813 221-227.
- 814