1 **Strapline: Original Article** 2 3 Phylogenetic structure and ecological and evolutionary determinants of species richness for 4 angiosperm trees in forest communities in China 5 Hong Qian<sup>1\*†</sup>, Richard Field<sup>2†</sup>, Jinlong Zhang<sup>3</sup>, Jian Zhang<sup>4</sup>, and Shengbin Chen<sup>5\*</sup> 6 7 8 <sup>1</sup>Research and Collections Center, Illinois State Museum, 1011 East Ash Street, Springfield, IL 9 62703, USA 10 <sup>2</sup>School of Geography, University of Nottingham, Nottingham, UK 11 <sup>3</sup>Flora Conservation Department, Kadoorie Farm & Botanic Garden, Lam Kam Road, Tai Po, New 12 Territories, Hong Kong 13 <sup>4</sup>Department of Renewable Resources, University of Alberta, Edmonton, Alberta T6G 2H1, Canada 14  $^{5}$ Nanjing Institute of Environmental Sciences, Ministry of Environmental Protection, 8 15 Jiangwangmiao Street, Nanjing, 210042, China 16 \*Correspondence: Hong Qian, <a href="mailto:hgian@museum.state.il.us">hgian@museum.state.il.us</a>, Research and Collections Center, 17 18 Illinois State Museum, 1011 East Ash Street, Springfield, IL 62703, USA; or Shengbin Chen, 19 chainpin@126.com, Nanjing Institute of Environmental Sciences, Ministry of Environmental 20 Protection, 8 Jiangwangmiao Street, Nanjing, 210042, China 21 <sup>†</sup>These authors contributed equally to this work. 22 23 Running title: Evolutionary and ecological causes of species richness patterns 24 Word count (Abstract to end of References, excluding Acknowledgments): 7,074 25

### **ABSTRACT**

Aim Site-level species richness is thought to result from both local conditions and species' evolutionary history, but the nature of the evolutionary effect, and how much it underlies the correlation with current environment, are debated. Although tropical conservatism is a widely used explanatory framework along temperature gradients, it is unclear whether cold tolerance is primarily a threshold effect (e.g. freezing tolerance) or represents a more continuous constraint. Nor is it clear whether cold tolerance is the only major axis of conservatism or whether others, such as water-stress tolerance, are additionally important or trade-off against cold tolerance. We address these questions by testing associated predictions for forest plots distributed across 35° latitude.

### Location China.

- Methods We recorded all trees within 57 0.1-ha plots, generated a phylogeny for the 462 angiosperm species found, and calculated phylogenetic diversity (standardized PD), net relatedness index (NRI) and phylogenetic species variability (PSV) for each plot. We tested the predictions using regression, variance partitioning and structural equation modelling to disentangle potential influences of key climate variables on NRI and PSV, and of all variables on species richness.
- **Results** Species richness correlated very strongly with minimum temperature, non-linearly overall but linearly where freezing is absent. The phylogenetic variables also correlated strongly with minimum temperature. While NRI and PSV explained little additional variance in species richness, they accounted for part of the species richness—current climate correlation. Water stress added minimal explanatory power. All these variables showed strong latitudinal gradients.

Main conclusions Minimum temperature appeared to primarily control tree species richness, via both a threshold-like freezing effect and a linear relationship in climates without freezing. We found no clear signal of water-stress effects. The modelled contribution of evolutionary history is consistent with cold-tolerance conservatism, but could not account for all the species richness—climate relationship.

Keywords

Climate, latitudinal diversity gradient, niche conservatism, phylogenetic community ecology, phylogenetic structure, tropical conservatism hypothesis

#### Introduction

The assemblage of species in any given location results from local gain and loss of species through time. Some of this results from relatively rapid responses to local conditions and some from the deeper-time evolutionary history of the region, affecting the pool of species able to reach the site ('regional pool'; we favour this site-specific theoretical definition of the regional pool). Over very large extents, the identities of the species found locally must depend in large part on evolutionary history (Ricklefs, 1987; Algar *et al.*, 2009). In particular, the species in the regional pool that can exist in a given set of environmental conditions may be strongly determined by the climatic regimes under which they evolved, constrained by a tendency for niches to be conserved (Diamond, 1975; Chase & Leibold, 2003; Wiens & Donoghue, 2004).

Because taxa differ in their niches, it is a truism that niches must be conserved on phylogenies, at least to some extent. Niche conservatism as a principle is therefore commonly used for understanding and explaining the effects of evolutionary history on the composition of species assemblages (e.g. Chase & Leibold, 2003).

However, given that evolution happens, it is also a truism that niches are labile, at least to some extent. Much less clear are which aspects of the climatic niche are more strongly conserved, which less so, and whether key niche axes are correlated or subject to trade-offs with respect to niche conservatism; here we aim to make progress towards answering these questions. To date, the focus has been primarily on cold tolerance. Most lineages have evolved under warmer conditions than today and temperature tolerance is thought to be strongly conserved phylogenetically (tropical conservatism hypothesis, 'TCH'). Specifically, it is hypothesized that cold tolerance in angiosperms has mostly evolved since the cooling that started in the early

Eocene (~50 mya; Graham, 2011; Condamine *et al.*, 2012), and only in relatively few lineages that have since diversified to some extent (e.g. Latham & Ricklefs, 1993; Wiens & Donoghue, 2004).

Thus, fewer species are expected to be found today in colder climates, and these should on average be more closely related to each other (more phylogenetically clustered) than in warmer climates—producing predictions 1a–1c in Table 1.

Different versions of the TCH exist, some implying that cold tolerance is a relatively continuous phenomenon, with difficult-to-evolve adaptations required for a broad range of the temperatures that characterize today's gradient from tropical to temperate to high-latitude climates (e.g. Qian, 2014). For convenience, we label this the 'continuous TCH'. Various lines of evidence are consistent with such a non-threshold view of cold tolerance (e.g. Qian, 2014). From this hypothesis, we expect continuous relationships between a range of temperature variables and both species richness and phylogenetic relatedness (Table 1, prediction 2a).

In contrast, tolerance to freezing, specifically, is often stressed as key to tropical conservatism (e.g. Wiens & Donoghue, 2004). In this 'threshold TCH', freezing represents a stepchange in an organism's environmental conditions, requiring specific adaptations that may not readily evolve. Under this hypothesis, the minimum temperatures experienced by organisms become the main focus. For gradients that include sites experiencing freezing, species richness and phylogenetic relatedness should be more strongly correlated with minimum temperature than with other temperature variables (Table 1, prediction 3a; see, for example, Hawkins *et al.*, 2014). From this hypothesis we also expect the relationships between minimum temperature and both species richness and phylogenetic relatedness to be more threshold-like (predictions 3b and 3c).

However, it is not realistic to expect very clear thresholds because freezing tolerance is complex. For trees, it may be manifest in at least three different ways, with the timing of the cooling being a key distinction. First, there are winter frosts. Most trees with at least some freezing tolerance have an acclimation period, in which they become more resistant to colder temperatures (e.g. by increasing the concentration of solutes in their tissues, lowering the freezing point of the water within them). There is wide variation between frost-hardy species in just how low a temperature they can experience and still survive, once they have acclimated, but this can be as low as -80°C (Sakai & Weiser, 1973). However, if acclimation has not happened, the same trees may be killed or badly damaged by much milder temperatures, so the timing of the coldest temperatures is also critical. Second, late-spring frosts can damage newly growing plant tissue. Third, in trees in particular, early-spring warming may cause freezing-related desiccation for example when the sun warms tree crowns while the trunks remain frozen, meaning no supply of water and solutes to the metabolizing crowns. Because these varying causes of damage occur at different temperatures and depend on antecedent conditions and timing, a clear-cut threshold is not expected in the relationship between minimum temperature and response variables that aggregate across species and sites. Even so, threshold-like relationships between minimum temperature and species richness or phylogenetic relatedness may emerge if freezing tolerance is important.

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Some aspects of the climatic niche may be more conserved than others under changing climates, depending largely on the combination of the selection pressure and the difficulty of adapting to change experienced. Freezing is both a strong selective force, because it is often lethal, and difficult for many lineages to adapt to because the nature of freezing damage (e.g.

physical damage caused by the expansion of water as it freezes, and the desiccation of tissues) often requires complex adaptations. It is not surprising, then, that freezing tolerance has received considerable attention in relation to niche conservatism. Not so clear is why cold tolerance should be an important structuring force in places not experiencing frost, though chilling injury is known in some tropical plants at temperatures as high as 12°C (Lyons, 1973). Further, the continuous (non-linear) response of the rate of chemical reactions to temperature may play a role in linking non-freezing temperatures to species richness (e.g. Brown *et al.*, 2004). The reasonable amount of evidence consistent with a continuous TCH probably results from a research focus on cold tolerance generally, typically without explicit consideration of the exact mechanism of cold tolerance involved.

While cold tolerance has received the most attention, it may not be the only major axis of niche conservatism. In theory, any aspect of tolerance to ambient environmental conditions that is both phylogenetically conserved and difficult to evolve may be important in determining patterns of relatedness and, potentially, species richness. Any aspect of the environment experienced by organisms that is now 'harsh' (relative to the conditions prevailing over their evolutionary history) has potential for (partly) explaining composition and richness of current species assemblages via a niche conservatism mechanism (Wiens & Graham, 2005). For example, it is thought that the world has become drier over approximately the same time-period as it has become colder (e.g. Wolfe, 1975). Tolerance of low water availability therefore represents a strong candidate for attention (Wiens & Graham, 2005), yet so far it has been largely ignored. From this 'water-stress conservatism' hypothesis we expect phylogenetic relatedness and species

richness to relate to water stress in similar ways as to cold temperatures (Table 1, predictions 4a and 4b).

A fundamental concept underlying niche conservatism is a trade-off between niche axes, but they may also be correlated; Kraft *et al.* (2007) called for research exploring correlations and trade-offs between traits in the context of community assembly. Many adaptations to freezing may also represent adaptations to drought because (as mentioned above) a key part of freezing stress is lack of liquid water. We may therefore expect some of the same variation in species richness and phylogenetic relatedness to be accounted for by both temperature and water-stress variables (prediction 4c).

Given a large span of latitude, we expect species richness to correlate strongly with both latitude and climate (Table 1, prediction 1c). This prediction is far from unique to the TCH, and there is debate (e.g. Wiens & Graham, 2005; Algar et al., 2009; Brown, 2014) about whether species richness is determined mainly by current climate or mainly by niche conservatism, or whether both contribute strongly. While it is possible that neither is a major determinant of species richness, we consider this very unlikely (e.g. see Field et al., 2009). Saliently, the characters thought to be conserved under the TCH are directly related to climatic tolerances. Therefore, with respect to species richness, it is appropriate to consider the TCH not as an alternative to current climate but as a hypothesis for why current climate is correlated with species richness. Given that tropical conservatism is not the only hypothetical mechanism for species richness—current climate correlations (e.g. see Lavers & Field, 2006; Algar et al., 2009; Jocque et al., 2010; Boucher-Lalonde et al., 2014; Brown, 2014), it is informative to ask how

much of the correlation can be accounted for by phylogenetic relatedness (Table 1, predictions 5a–5c).

Here we test the predictions outlined in Table 1 by analyzing data from forest plots in China, spanning 35° latitude. First, we test whether the patterns of phylogenetic relatedness in our data are consistent with the TCH, and if so, which version of the TCH receives the most support. For this, we focus on which of the (inter-correlated) temperature variables best account for relatedness, whether the relationships are threshold-like, and whether water-stress variables account for additional variation or overlap. Second, using path analysis and variance partitioning (similarly to Algar *et al.*, 2009 and Qian *et al.*, 2015), we assess how much of the spatial variation in species richness is accounted for by direct statistical effects of climate versus direct effects of evolutionary history, the degree of overlap between the two, and the indirect effects of climate routed through evolutionary history (Table 1).

### **Materials and Methods**

Species data

Forest plots were sampled in 15 areas (Appendix S1), 14 of which are nature reserves, spanning 35° latitude from tropical rain forests to boreal forests. In each area, four 20 x 50 m (0.1 ha) forest plots were sampled and latitude, longitude and elevation of each were recorded. Woody individuals with diameter at breast height at least 3 cm were identified to species, all of which are native. Species-level botanical nomenclature was standardized using the Flora of China (Wu et al., 1994–2013). Three forest plots with fewer than two angiosperm tree species were

excluded because some phylogenetic metrics used (see below) require at least two species. The remaining 57 plots contained 462 angiosperm tree species in 187 genera and 64 families.

### Phylogeny

We generated a phylogeny for the 462 species, based on the species-level phylogeny of Zanne *et al.* (2014), which includes 30,535 angiosperm species and was time-calibrated using seven gene regions and fossil data. Orders and families in the phylogeny were constrained according to APG III (2009). All the families and 97% of the genera in our dataset were included in Zanne *et al.*'s phylogeny. Of the six missing genera, *Ellipanthus* is the only genus of its family in our data and thus was represented by its family branch in our phylogeny; the remaining five genera were randomly distributed among tips within their respective families. Zanne *et al.*'s phylogeny includes 294 (64%) of our 462 species. Of the other 168, 37 belong to genera with only one species in our data and were represented by branches of their respective genera. Thus 331 (72%) of the 462 species were completely resolved in our phylogeny. The remaining 131 species were randomly distributed among tips within their genera, following recent literature (e.g., Brunbjerg *et al.*, 2014). We checked sensitivity of results to the method used for placing unresolved species in the phylogeny; see Appendix S2

### Metrics of phylogenetic diversity and phylogenetic structure

We used Faith's (1992) phylogenetic diversity (PD) metric to quantify each plot's phylogenetic diversity as the total phylogenetic branch length joining the basal node (here the angiosperm) to the tips of all the species in the plot. Because PD correlates strongly with species richness,

following previous studies (e.g. Slik *et al.*, 2009) we calculated standardized PD, independent of species richness. Specifically, we randomly selected 10 species from each plot and calculated PD (for the 36 plots with 10 or more species). We repeated this 1000 times per plot and calculated the mean of the 1000 randomized PD values.

We used two indices, net relatedness index (NRI; Webb, 2000) and phylogenetic species variability (PSV; Helmus et~al., 2007), to quantify phylogenetic structure (evolutionary inertia) per plot (also see Qian et~al., 2015). NRI is commonly used to measure the standardized effect size of mean phylogenetic distance (MPD), which estimates the average phylogenetic relatedness between all pairs of taxa in an assemblage. Webb (2000) defined NRI as: NRI =  $-1 \times (MPD_{observed} - MPD_{randomized})/(sdMPD_{randomized})$ , where MPD<sub>observed</sub> is observed MPD, MPD<sub>randomized</sub> is the expected MPD of randomly generated assemblages (n = 1000) generated by drawing a number of species randomly from across the phylogeny equal to the observed number of species in the assemblage, and sdMPD<sub>randomized</sub> is the standard deviation of the MPD for the randomized assemblages. To generate randomized (null) assemblages, species in each forest plot were randomly drawn from the pool of all species in the phylogeny (i.e. model 2 of Phylocom; <a href="http://phylodiversity.net/phylocom">http://phylodiversity.net/phylocom</a>). Positive NRI indicates lower MPD than expected by chance (species more closely related than expected): phylogenetic clustering of species. Conversely, negative NRI (species more distantly related than expected by chance) indicates phylogenetic evenness or overdispersion. NRI was calculated using Phylocom (Webb et~al., 2008).

Helmus *et al.* (2007) defined PSV as: PSV =  $(n\text{trC} - \Sigma C)/(n(n-1))$ , where *n* is the species richness, C is a covariance matrix summarizing the correlation structure of the community phylogeny, trC is the trace (sum of the diagonal elements) of C,  $\Sigma$ C is the sum of all elements in C.

PSV is standardized to vary from zero (maximum relatedness or clustering) to one (minimal relatedness or maximal evenness: all species from disparate parts of the phylogenetic tree) (Algar *et al.*, 2009). It is independent of species richness (Helmus *et al.*, 2007; Savage & Cavender-Bares, 2012) and was calculated using Picante (Kembel *et al.*, 2010).

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### Environmental variables

Temperature and precipitation are typically the environmental variables most strongly correlated with species richness for terrestrial plants and animals (Hawkins et al., 2003; Field et al., 2009; Wang et al., 2011) and were used in previous studies assessing the relative importance of evolutionary and ecological effects on species richness (e.g. Algar et al., 2009; Qian et al., 2015). Mean temperatures for all 12 months and their derivations BIO1–BIO11 in the WorldClim database (Hijmans et al., 2005; http://www.worldclim.org) were all strongly correlated in our data, but the minimum coldest-month temperature (BIO6, hereafter 'minimum temperature') is the most appropriate for assessing freezing tolerance and cold tolerance more generally (e.g. Algar et al., 2009; Hawkins et al., 2014; Qian et al., 2015). We therefore used this variable in testing all the predictions relating to temperature (i.e. all the predictions in Table 1 except 1b, 4a, 4b and 5a). To test predictions 2a, 2b, 3a and 3c we also used mean annual temperature (BIO1, hereafter 'mean temperature'), mean annual potential evapotranspiration ('PET', calculated using the method of Willmott et al., 1985; see Gavin & Hu, 2006), temperature seasonality (BIO4) and the maximum warmest-month temperature ('maximum temperature'; BIO5). As a preliminary analysis, we examined the bivariate relationships between all the response and temperature variables.

Mean annual precipitation is commonly used to measure water availability in species richness studies (e.g. Field *et al.*, 2009; Algar *et al.*, 2009; Qian *et al.*, 2015) and we used it to test predictions relating to current climate (5b and 5c in Table 1). To test water stress-related conservatism (predictions 4a–c), water availability during the (main) growing season and, particularly, measures of water stress are more appropriate. We obtained various water-related variables, measuring water deficit, water surplus, actual evapotranspiration, precipitation and precipitation seasonality, including WorldClim BIO12–BIO19. We initially examined bivariate relationships between all the response and water variables, but only entered a few water variables into our modelling, based primarily on theoretical reasoning and secondarily on the bivariate relationships: mean annual precipitation (BIO12, for comparison with previous studies), annual water deficit, warmest-quarter precipitation (BIO18), driest-month precipitation (BIO14), water surplus in summer (specifically August, which showed stronger relationships with some response variables than water surplus in other summer months). Appendix S3 lists all environmental variables obtained and initially examined, and their sources.

### Data analysis

We first examined whether the observed values of NRI and PSV differ from null expectation. For each of the 57 forest plots, we generated 1000 null assemblages by randomly shuffling the species on the tips of the phylogeny. From these null assemblages we computed the mean NRI and PSV for each plot, to create the variables NRI<sub>null</sub> and PSV<sub>null</sub>.

We took three approaches to assessing the relationships between species richness and the evolutionary and ecological variables. First, we used Pearson's correlation to quantify the

bivariate relationships between variables (multiple R for quadratic relationships). Second, for prediction 5b (Table 1) we conducted partial regressions (Legendre & Legendre, 1998) to partition the variance in species richness into four portions: uniquely accounted for by (a) ecological variables (current climate), (b) evolutionary variables (NRI and PSV), (c) jointly accounted for by the ecological and evolutionary variables, and (d) accounted for by none of the variables. We used SAM version 4.0 (Spatial Analysis in Macroecology; Rangel *et al.*, 2010) for correlation and regression analyses, including variance partitioning. Using this software, we checked for spatial structure in the residuals.

Third, we used path analysis (within structural equation modelling, 'SEM') to estimate the direct effects of these variables on species richness (predictions 5a and 5b), and the indirect effect of climate routed through evolutionary history (prediction 5c). The framework of our path analyses is based on Algar *et al.* (2009) and Qian *et al.* (2015), in which species richness was the response, climate variables were exogenous and NRI and PSV were endogenous variables. Unlike those studies, we did not transform species richness because this produced good normality and homoscedasticity, better in both respects than using any of the common transformations. We used Amos (http://amosdevelopment.com) for our SEM analyses.

### Results

The number of angiosperm tree species per genus varied from 1 to 22 (Fig. 1). As expected, PD was strongly correlated with species richness of angiosperm trees (r = 0.918), which was strongly correlated with latitude (r = -0.880; Fig. 2A), and PD was strongly correlated with latitude (r = -0.881). Similarly, standardized PD (accounting for species richness), was strongly correlated with

latitude (Fig. 2B), and with minimum temperature (Table 2), as expected (Table 1, prediction 1a). NRI and PSV were negatively correlated with each other (r = -0.832), and both were correlated with latitude (Fig. 2C, D). Both were also correlated with species richness (Table 2), in line with prediction 1b except that the negative species richness—NRI correlation was not very strong. Species richness was lower in colder climates (Table 2), consistent with prediction 1c. Values of NRI and PSV derived from the null assemblages were not significantly correlated with species richness, latitude, minimum temperature or mean annual precipitation (P ranging from 0.142 to 0.805).

The freezing tolerance and chilling tolerance versions of TCH predict that minimum temperature correlates more strongly with species richness, NRI and PSV than do other temperature variables (Table 1, prediction 3a), while other forms of the continuous TCH predict no such difference (prediction 2b). With species richness and NRI, minimum temperature was slightly more correlated than the other temperature variables (Table 2), with two exceptions (temperature seasonality for species richness and mean annual temperature for NRI), both of which were extremely co-linear with minimum temperature (|r| = 0.99; Table 2). PSV was no more strongly correlated with minimum temperature than with other temperature-related variables. The proportion of variation accounted for by (minimum) temperature was lower for evolutionary history than for species richness: in regressions with minimum temperature as the only explanatory variable,  $r^2$  was 0.841 with species richness as response, 0.729 for PSV<sup>2</sup> and 0.576 for NRI (Table 2, Fig. 3).

The continuous and freezing tolerance TCHs are also distinguished by their predictions about whether relationships between temperature (especially minimum temperature) and

species richness, NRI and PSV show threshold-like behaviour (prediction 2a vs 3b and 3c). For species richness, the relationship with minimum temperature (and with the highly co-linear temperature seasonality) showed some evidence of a threshold (Fig. 3A, B). Evidence for threshold-like relationships between temperature and either NRI or PSV was weak at best; the most non-linear relationship for each is shown in Fig. S2. However, all significantly clustered communities (i.e. those forest plots with NRI > 1.96, the critical Z-score for significance at P = 0.025 in a one-tail test) experience minimum temperatures below freezing, most well below freezing (Fig. S2A).

There was little evidence for species in more water-stressed forest plots being more closely related (see Table 1, prediction 4a). Mean annual precipitation did correlate positively with standardized PD (Table 3), but became non-significant when minimum temperature was accounted for (*P*=0.58; multiple regression with standardized PD as response, minimum temperature and its square fit as covariates, N=36). Further, standardized PD was uncorrelated with warmest-quarter precipitation and actually positively correlated with water deficit (though only marginally significant; Table 3). Relationships between water variables and both NRI and PSV were mostly non-significant or weak (Table 3); neither NRI nor PSV correlated significantly with the most direct measure of water stress, annual water deficit, in contradiction to prediction 4b.

For species richness, relationships with water variables were stronger, but still weaker than with temperature variables (Tables 2, 3), and importantly the correlation with water deficit was not significant. Of the water variables, mean annual precipitation was the strongest correlate of species richness, but was also strongly correlated with minimum temperature (Table 3) and only uniquely accounted for 1.6% of the variance in species richness in partial regression (data

not shown). Partitioning the species-richness variance between that accounted for by minimum temperature (quadratic) and the water-stress variables, annual water deficit was not significant (*P*=0.175). Warmest-quarter precipitation correlated quite strongly with species richness (Table 3), accounting for 45% of the variance, almost all of which overlapped with minimum temperature (Fig. 4A), in line with prediction 4c. Similarly, all the variance in PSV² that warmest-quarter precipitation could account for overlapped with minimum temperature (no significant independent contribution). In contrast, for NRI, both water-stress variables were significant in the variance partitioning, together accounting for 9% of the variance in NRI, none of which was shared with minimum temperature (Fig. 4B). Here, the warmest-quarter precipitation had a positive partial correlation with NRI—rather surprisingly in the same direction (positive correlation with NRI) as water deficit. In the path analysis, none of the water-stress variables added significantly to the models shown (Figs 5, S4).

Predictions 5a–c (Table 1) concern the amount of shared, and unique, explained variance between evolutionary and climatic (ecological) variables. We tested these predictions using partial regressions (Figs 4C, S3) and by comparing the direct and indirect effects on species richness as modelled in path analysis (Figs 5, S4). Overall explained variance was high (84–90%), but in the partial regressions the variables measuring evolutionary history (NRI and PSV) uniquely accounted for very little (1–3%). Shared explained variance always exceeded 50% of the total variance, and the unique contribution of the climate variables was approximately one third of the total. Similarly, in each path analysis, the direct effects of the evolutionary history variables were weak, though they were significant, while the direct effect of minimum temperature was strong

(Figs 5, S4). The indirect effect of climate, routed through evolutionary history, was also quite strong.

### Discussion

Minimum temperature appeared to primarily control tree species richness, via both a threshold-like freezing effect and a linear relationship in climates without freezing. This one climatic variable accounted for 84% of the variation in species richness. Adding variables measuring evolutionary history (NRI and PSV) and/or water surplus (August) raised this only slightly (to 87–90%; Figs 4–5, Appendix S2), but more than half of the total variance was shared between climate and evolutionary history. Thus, around two thirds of the climate—species richness correlation may be attributable to niche conservatism. The core predictions of the tropical conservatism hypothesis (TCH) were all met (predictions 1a–c and 5a, Table 1), while our findings also indicate an important direct role of minimum temperature (predictions 5b–c not met).

The strongest models involved a quadratic relationship between minimum temperature and species richness, in which species richness was very low when minimum temperature was below about -10°C, rapidly increasing at higher minimum temperatures. Minimum temperature could also account for much of the variation in NRI and PSV. These results are consistent with the freezing tolerance TCH (predictions 3a–c), and also with the chilling tolerance version of the continuous TCH (prediction 2a but not 2b). Water variables tended not to add much explanatory power to any of these models, in these mostly mesic forest plots, and overall there was little support for water-stress conservatism (predictions 4a–c).

Before proceeding, we note that our R<sup>2</sup> values are very high for fine-grained plot data (Field *et al.*, 2009). Concern exists that very strong spurious relationships between richness, climate and phylogenetic metrics can result from repeated co-occurrences of species across plots (B.A. Hawkins, pers. comm.). On average, each species occurred in only 3.8 of our 57 plots (mean: 3.79, standard deviation: 3.89), so we consider this potential problem to be minimal in our analysis.

*Tropical conservatism (predictions 1–3)* 

Consistent with the TCH, minimum temperature outperformed other temperature variables that were not excessively co-linear with it, in accounting for species richness. Similarly, Wang *et al.* (2011), found that species richness covaried more with minimum temperature than their other five temperature-related variables in 2500-km² cells across China—accounting for 10% more of the variation (69%) than mean annual temperature (59%). Thus, the two studies, focusing on very different spatial scales, are consistent with each other and with the notion that freezing tolerance is important in structuring species richness patterns (Latham & Ricklefs, 1993; Wiens & Donoghue, 2004)—but also with any limitation associated with minimum temperatures, such as chilling tolerance. At approximately the same scale (grain and latitudinal extent) as our study, Hawkins *et al.* (2014) examined TCH predictions for clade age of angiosperm tree species in forest plots in the USA. They found clade age correlated more strongly with minimum temperature than mean or maximum temperature, and it correlated very strongly with cold tolerance. Taken together, these findings support the notion that the TCH can account for a considerable part of the relationship between climate and tree species richness and composition at the plot level.

We found evidence for threshold-like behaviour in the relationship between minimum temperature and species richness (Fig. 3A): above minimum temperature values of about -5°C, species richness values were much higher than below -15°C (unfortunately we had no data between these two values)—consistent with the fact that water containing solutes freezes at slightly lower temperatures than 0°C. Above -5°C, the relationship was continuous (near-linear; Fig.3A), suggesting that freezing tolerance is not the only temperature-related effect on species richness, and that a continuous, positive relationship with temperature exists—superimposed on a threshold. Further research should be focused on this. Minimum temperature linearly related to both NRI and PSV above -5°C, with no evidence for thresholds. Overall, then, while our results are consistent with freezing tolerance as an important structuring force in tree assemblages, they also emphasize a more continuous effect of (minimum) temperature.

Water-stress conservatism (predictions 4)

Water variables generally added little to minimum temperature in accounting for species richness, NRI or PSV (Table 3; Figs. 4–5, see also Figs. S3–S4 in Appendix S2). The dessication-related link between cold tolerance and drought tolerance (see Introduction) was expected to lead to shared explained variance between minimum temperature and water-stress variables (prediction 4c). However, water deficit was not even significantly correlated with species richness, NRI or PSV (Table 3). Overall, we found no compelling evidence for trade-offs or correlations among niche axes within niche conservatism. Our study forests were mostly mesic, though annual water deficit ranged up to 150 mm and 9 of the 57 plots had 5 or 6 months of water deficit; these relatively water-stressed plots did not have high residuals from fits with minimum

temperature, suggesting little effect of water stress. Further, the water variable adding most to minimum temperature in accounting for both species richness and NRI was a measure of summer water surplus (Appendix S2), not water stress.

Putting our results together with others from Asia and the New World, we find little evidence for water-stress conservatism being a strong factor in determining the species richness and composition of angiosperm trees at continental extents. Wang *et al.* (2011), analyzing a gradient from deserts to tropical rainforests in China, found that minimum temperature accounted for 9% more of the variation in species richness than did the best water-related variable. Hawkins *et al.* (2014) found that minimum temperature was a much stronger predictor of clade age of angiosperm tree species in USA forest plots than annual precipitation, precipitation of the driest or warmest quarters, or soil moisture. Moles *et al.* (2014) concluded that plant traits correlated more strongly with mean annual temperature than mean annual precipitation. We interpret these results as consistent with ideas of resource-based increase in the capacity to support more species (e.g. Wright, 1983; Currie & Paquin, 1987; Lavers & Field, 2006). Additional research, focusing on more water-stressed areas, is needed to investigate this further.

Beyond niche conservatism (predictions 5)

The path analyses (Figs 5, S4) suggest some variance in species richness is due to direct effects of evolutionary history on species richness. This is consistent with the idea that current climate sorts species according to tolerances built up over their evolutionary history, these phylogenetic patterns in turn partly determining species richness (Wiens & Donoghue, 2004). It is not clear

from the various forms of niche conservatism hypothesis why (minimum) temperature should correlate so much less strongly with NRI and PSV than with species richness—the opposite to Qian *et al.* (2015), who found minimum temperature correlated much more strongly with PSV than species richness or NRI for North American angiosperm trees in 1° grid cells. The necessarily imperfect correlation between current minimum temperature and past minimum temperature may be part of the explanation. Regardless, a substantial proportion of the variation in species richness was accounted for by minimum temperature directly, with no apparent connection to evolutionary history. Thus, at most, niche conservatism explains only part of the species richness—climate correlation. Further, Boucher-Lalonde *et al.* (2014) found that, while bird species richness in 10,000-km² grid cells in the Americas seasonally tracks environment, individual species' seasonal geographical distributions do not—implying minimal role for niche conservatism in accounting for species richness patterns. Clearly, niche conservatism can only be part of the explanation for species-richness patterns (Wiens & Graham, 2005).

Algar *et al.* (2009, their Fig. 3) and Qian *et al.* (2015, their Fig. 2) both used path analysis to separate direct effects of climate on species richness from indirect effects via evolutionary history. They drew opposing conclusions: Algar *et al.* (2009) found minimal direct effects of evolutionary history, while Qian *et al.* found strong ones. Our parallel analyses (Figs S3A, S4A) produced results intermediate between the two, while suggesting a possible freezing-related threshold. It is likely that the balance of direct current climate and evolutionary history effects in explaining species richness and its patterning varies with taxon, scale and study region.

Absolute numbers of species should be differentiated from spatial species-richness patterns. In two regions with the same richness—temperature correlation, the number of species

at a given temperature may be much higher in one region than the other (Latham & Ricklefs, 1993). Although our study is limited to one region, the distinction is important because it affects possible explanations for direct effects of climate on species richness. One possibility invokes carrying capacity for species richness (e.g. Brown, 2014), perhaps set by productivity or combinations of resource states (Lavers & Field, 2006). Although the carrying capacities may not be absolute (Etienne *et al.*, 2012), such explanations require better understanding of why carrying capacities are apparently so different in different parts of the world with very similar climates.

Other candidates exist for explaining directs effect of current climate on species richness without involving (deep-time) evolutionary history. Climate may affect dispersal and its balance with ecological specialization (Jocque *et al.*, 2010), for example. Another interesting possibility is that climate determines stochastic immigration and extinction rates (Boucher-Lalonde *et al.*, 2012). Evolutionary history may have resulted in different-sized species pools in different biogeographical realms, and, if rates of immigration to (and/or local extinction within) areas within these realms are modified by climate, then richness–climate correlations may emerge that result directly from climate (Boucher-Lalonde *et al.*, 2014). The degree to which these correlations are modified by local environmental filtering (e.g. Qiao *et al.*, 2015) based on conserved niches may vary with scale, taxon and study region.

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495	
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613	

615	SUPPORTING INFORMATION
616	Additional Supporting Information may be found in the online version of this article:
617	
618	Appendix S1 Map showing where the 57 forest plots used in the present study were sampled.
619	Appendix S2 Additional analyses and results.
620	Appendix S3 Environmental variables obtained and their sources.
621	
622	BIOSKETCHES
623	Hong Qian's research is multidisciplinary and particularly lies at the interface of ecology and
624	biogeography ( <u>www.researchgate.net/profile/Hong_Qian3</u> ).
625	Richard Field's main interests are in biodiversity patterns, conservation biogeography
626	(particularly with reference to tropical rainforests) and island biogeography.
627	
628	H.Q. and R.F. conceived the ideas, conducted most of the statistical analyses, and wrote the
629	paper. S.C. collected the plant data. Both J.Z. conducted the phylogenetic analyses. All authors
630	contributed to revisions.
631	
632	Editor: Peter Linder
633	

# Tables

 Table 1. Specific predictions tested herein. The predictions are grouped by numbers corresponding to the hypotheses or debate investigated, subdivided by letters into a set of specific predictions for each. 'TCH' = tropical conservatism hypothesis; 'PSV' = phylogenetic species variability; 'NRI' = net relatedness index; 'minimum temperature' = minimum temperature of the coldest month.

Hypothesis or debate	Predictions				
1. TCH (core predictions)	1a. Species in forest plots with lower temperatures are more closely related (i.e. phylogenetically clustered).				
	<ol> <li>Species richness correlates strongly with PSV (positively) and NRI (negatively).</li> </ol>				
	1c. Species richness is lower in colder climates.				
2. Continuous TCH	2a. Continuous bivariate relationships between temperature variables and (i) species richness, (ii) PSV, (iii) NRI.				
	2b. Minimum temperature no more correlated with species richness, PSV or NRI than other temperature variables. [Not true for the chilling tolerance version of the hypothesis.]				
3. Freezing tolerance (threshold) TCH	3a. Species richness, PSV and NRI correlate more strongly with minimum temperature than with other temperature variables. [Also true for chilling tolerance version of the continuous TCH.]				
	3b. Minimum temperature has a threshold-like relationship with species richness, PSV and NRI.				
	3c. Threshold-like relationships with species richness, PSV and NRI are less clear for other temperature variables than for minimum temperature.				
4. Water-stress	4a. Species in more water-stressed forest plots are more closely related.				
conservatism	4b. Water stress correlates negatively with species richness and PSV, and positively with NRI.				
	4c. Water stress overlaps with minimum temperature in accounting for variation in species richness, PSV and NRI.				
5. Evolutionary history	5a. Evolutionary history has strong direct effects on species richness.				
vs other mechanisms for species richness– current climate	5b. Current climate explains little additional variation in species richness once evolutionary history is accounted for.				
correlations	5c. Current climate has strong indirect effects on species richness, routed through evolutionary history, and weak direct effects.				

Table 2. Pearson's correlation coefficients among species richness (SR), standardized phylogenetic diversity (PD<sub>std</sub>), net relatedness index (NRI), phylogenetic species variability (PSV, transformed by being squared) and the five temperature (or related) variables analyzed: mean annual temperature (MeanT), mean annual potential evapotranspiration (PET), temperature seasonality ( $T_{seas}$ ), maximum temperature of the warmest month (MaxT) and minimum temperature of the coldest month (MinT). N = 57 except correlations involving PD<sub>std</sub>, where N = 36. P < 0.001 for all the relationships shown.

	SR	$PD_{std}$	NRI	PSV <sup>2</sup>	MeanT	PET	$T_{Seas}$	MaxT
PD <sub>std</sub>	0.723 <sup>a</sup>							
NRI	-0.650 <sup>b</sup>	-0.813						
PSV <sup>2</sup>	0.742	0.879	-0.869					
MeanT	0.893 <sup>b</sup>	0.816	-0.757 <sup>b</sup>	0.859				
PET	0.859	0.786 <sup>b</sup>	-0.737	0.871 <sup>b</sup>	0.950			
$T_{seas}$	-0.921 <sup>b</sup>	-0.820	0.724 <sup>b</sup>	-0.811	-0.962	-0.864		
MaxT	0.655	0.577	-0.659	0.739	0.826	0.880	-0.646	
MinT	0.917 <sup>b</sup>	0.838	-0.759 <sup>b</sup>	0.854	0.991	0.912 <sup>c</sup>	-0.985	0.759

<sup>650</sup> a When correlated with log(SR), r = 0.812

<sup>651</sup> b Quadratic fit

<sup>&</sup>lt;sup>c</sup> When modelled as quadratic with MinT + MinT<sup>2</sup> as explanatory, (multiple) r = 0.968

Table 3. Pearson's correlation coefficients among species richness (SR), standardized phylogenetic diversity (PD<sub>std</sub>), net relatedness index (NRI), phylogenetic species variability (PSV, transformed by being squared) and the water-related variables analyzed: annual water deficit (Water def.), precipitation of the warmest quarter (Precip<sub>warm</sub>), the surplus of precipitation over potential evapotranspiration in August (Sur<sub>Aug</sub>) and mean annual precipitation (Precip<sub>annual</sub>). Minimum temperature of the coldest month (MinT), as the most important temperature variable (Tables 1,2) is included to indicate the degree of correlation between temperature and water variables. N = 57 except correlations involving PD<sub>std</sub>, where N = 36. Non-significant relationships (P > 0.05) are indicated by 'ns'; P < 0.001 for all other relationships shown, except where otherwise indicated.

	SR	$PD_{std}$	NRI	PSV <sup>2</sup>	Water def.	Precip <sub>warm</sub>	Surplus <sub>Aug</sub>	Precipannual
PD <sub>std</sub>	0.723 <sup>a</sup>							
NRI	-0.650 <sup>b</sup>	-0.813						
PSV <sup>2</sup>	0.742	0.879	-0.869					
Water def.	ns	0.338 <sup>d</sup>	ns	ns				
Precip <sub>warm</sub>	0.674 <sup>b</sup>	ns	ns	0.521	-0.357 <sup>e</sup>			
Surplus <sub>Aug</sub>	0.415	ns	ns	ns	ns	0.506		
Precip <sub>annual</sub>	0.814 <sup>c</sup>	0.655	-0.557	0.762	ns	0.848	0.283 <sup>f</sup>	
MinT	0.917 <sup>b</sup>	0.838	-0.759 <sup>b</sup>	0.854	ns	0.579	ns	0.855

<sup>666</sup> a When correlated with log(SR), r = 0.812

<sup>667 &</sup>lt;sup>b</sup> Quadratic fit

<sup>668 &</sup>lt;sup>c</sup> When correlated with log(SR), r = 0.882

 $^{d}P = 0.043$ 

 $<sup>^{</sup>e}P = 0.007$ 

 $<sup>^{</sup>f}P = 0.033$ 

# 673 Figures

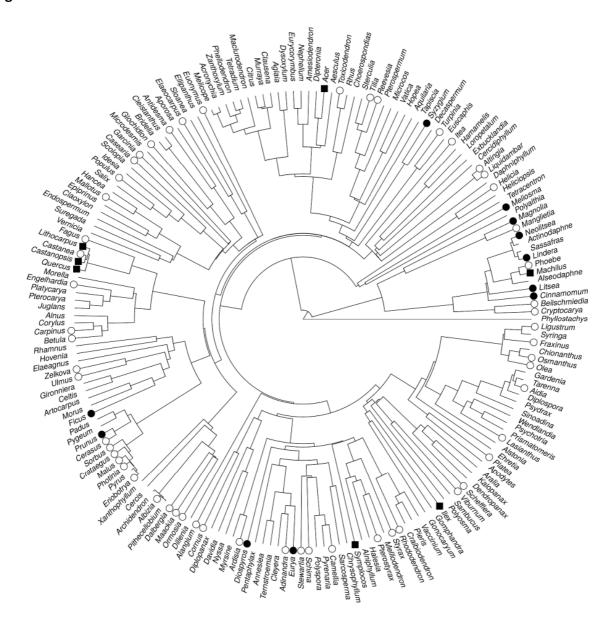


Figure 1 Phylogeny showing the 187 angiosperm tree genera and their species richness in the studied forest plots (for illustrative purposes; analyses are based on a species-level phylogeny, see Materials and Methods). The number of species in each genus is indicated by symbols: tip with no symbol represents 1 species in the genus; open circle is 2–5 species; filled circle is 6–10 species; filled square is >10 species.

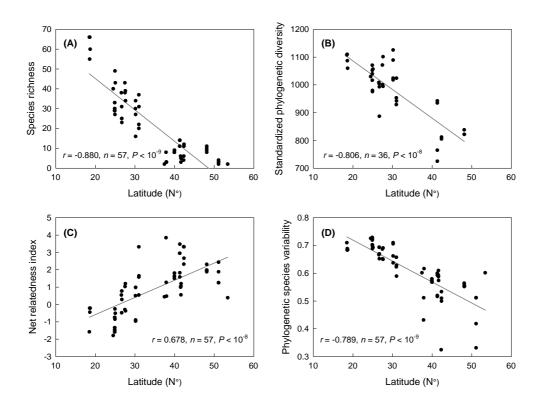


Figure 2 Latitudinal trends in (A) species richness, (B) standardized phylogenetic diversity, (C) net relatedness index and (D) phylogenetic species variability for the angiosperm tree communities examined in this study.

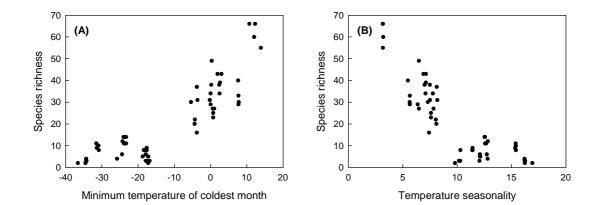


Figure 3 Evidence regarding threshold-like behaviour in the relationships between species richness and (A) minimum temperature and (B) temperature seasonality. In both graphs, N = 57 forest plots.

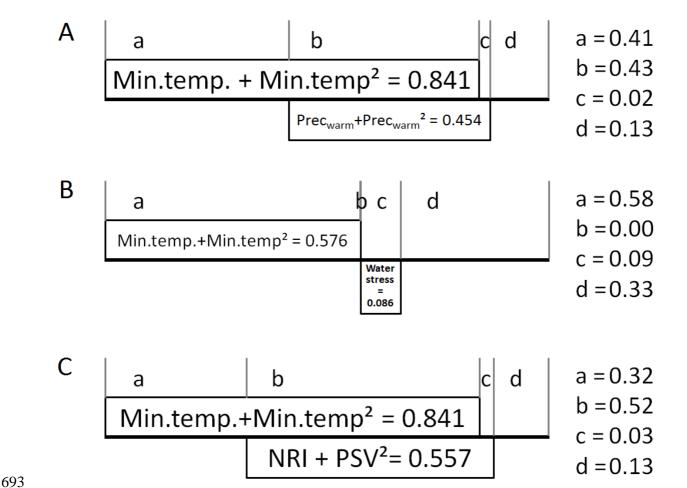


Figure 4 Partial regression results—partitioning the variance: (A) in species richness between temperature (above the thick line) and water-stress (below the line) variables; (B) in NRI between temperature and water-stress variables; (C) in species richness between climatic and evolutionary variables, where climate is represented only by minimum temperature. In each case, the variance in the response variable is partitioned into four portions: (a) uniquely accounted for by variable set 1 (above the thick line); (b) accounted for jointly by variable sets 1 and 2; (c) uniquely accounted for by variable set 2 (below the thick line); and (d) unexplained variance. Explanatory variables are 'Min.temp.' = minimum temperature of the coldest month, 'Prec<sub>warm</sub>' = warmest-quarter precipitation, 'water stress' = both Prec<sub>warm</sub> and annual water deficit, 'NRI' = net relatedness index and 'PSV' = phylogenetic species variability.

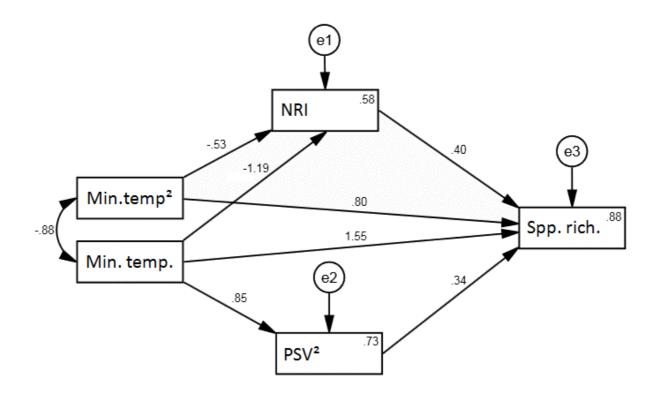


Figure 5 Structural equation model examining the influence on species richness ('Spp. Rich.'; untransformed) of climatic and evolutionary variables, where climate is represented only by minimum temperature ('Min. temp.') and evolutionary history is represented by net relatedness index ('NRI') and phylogenetic species variability ('PSV'; transformed by squaring it). Minimum temperature is modelled as having a quadratic relationship with NRI and species richness but a linear relationship with PSV<sup>2</sup>. For direct effects (single-headed arrows), values are standardized partial regression coefficients; non-significant direct effects were removed. For double-headed arrows (between exogenous variables), values are covariances. For endogenous variables (with error terms), the values given in their top-right corners are R<sup>2</sup>s showing the proportion their variation accounted for.