

1 **A critical review of microplastic pollution in urban freshwater environments**
2 **and legislative progress in China: recommendations and insights**

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23 Freshwater systems are vitally important, supporting diversity and providing a range of
24 ecosystem services. In China, rapid urbanization (over 800 million urban population) has
25 led to multiple anthropogenic pressures that threaten urban freshwater environments.
26 Microplastics (<5 mm) result from intensive production and use of plastic materials, but
27 their effects in urban freshwater environments remain poorly understood. Rising concerns
28 over the ecological effects of microplastics have resulted in increased attention being
29 given to this contaminant in Chinese freshwater systems. Some studies provide
30 quantitative data on contamination loads, but in general relevant knowledge in freshwater
31 environment remains narrow in China, and lacking adequate understanding of threshold
32 levels for detrimental effects. Notably, non-standardized sample collection and
33 processing techniques for point and non-point sources have hindered comparisons of
34 contamination loads and associated risk. Meanwhile, legislative frameworks for
35 managing microplastics in China remain in their infancy. This manuscript critically
36 reviews what is known of the nature and magnitude of microplastic pollution in Chinese
37 freshwater environments, and summarises relevant Chinese legislation. It provides
38 recommendations for improving the legislative framework in China and identifies
39 research gaps that need to be addressed to improve management and regulatory strategies
40 for dealing with microplastic pollution in Chinese urban freshwater environments.

41 **Keywords:** microplastics, urban freshwater environment, abundance, China, legislation,
42 policy

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70 **1. Introduction**

71 Annual global plastic resin and fibre production accelerated from 2 million metric tonnes in
72 1950 to 3.81 billion tonnes in 2015, contributing to the 6.3 billion tonnes of global plastic waste
73 produced by 2015 (Geyer *et al.*, 2017). The ubiquity of plastics results in serious environmental
74 problems worldwide, with over 99 million metric tonnes mismanaged yearly and approximately
75 79% of total plastic waste accumulating in landfills or in the natural environment (Geyer *et al.*,
76 2017; Lebreton and Andrady, 2019). Plastic pollution can be conveyed between different
77 ecosystems, with for example around ten million metric tonnes of terrestrial-based plastic litter
78 estimated to enter the oceans every year (Jambeck *et al.*, 2015). Freshwater systems are a major
79 pathway for delivering plastic pollution to the marine environment (Crawford and Quinn,
80 2017b).

81 China is the largest plastic producer in the world, with monthly plastic production
82 reaching at 5-12 million metric tonnes by 2019 (Garside, 2019). Since the early 1950s, the
83 utilisation of plastic mulch in agriculture has become widespread across China, while plastic
84 tableware and bags became prevalent in industrial and domestic sectors after the economic
85 reforms in the late 1970s (Zhou *et al.*, 2016). The growth of E-Commerce further increased
86 China's plastic consumption, especially the fast food delivery services which are part
87 responsible for the approximately 60 million items of plastic tableware daily used in China,
88 most of which is single-used (Industry, 2019). Thus, China plays an important role in global
89 plastic production and consumption (Wang *et al.*, 2018). The Chinese plastic material market
90 has resulted producing more than 8.82 million metric tonnes of mismanaged plastic waste
91 annually, which ranked as the global highest (Jambeck *et al.*, 2015). Based on current trends,
92 global mismanaged plastic waste is estimated to reach 155-265 million metric tonnes annually

93 by 2060s, with China remaining as one of the major sources (Everaert *et al.*, 2018; Lebreton
94 and Andrady, 2019; van Wijnen *et al.*, 2019).

95 “*Microplastics*”, defined as plastic debris smaller than 5 mm in diameter, can be
96 ingested by organisms, and abrade and clog breathing and feeding apparatus (Eerkes-Medrano
97 *et al.*, 2015; Li *et al.*, 2018). Given the size of microplastics, they can be transported for long
98 distances (i.e. more than thousands of km) and have been found in geographically remote
99 regions (i.e. polar areas and waterbodies on undeveloped plateaus) (Lusher *et al.*, 2015; Horton
100 *et al.*, 2017a; Baptista Neto *et al.*, 2019; C. Jiang *et al.*, 2019). Microplastics also have the
101 potential to adsorb other contaminants (including hydrophobic persistent organic pollutants,
102 pathogenic microorganisms and antibiotics) and transport these pollutants over a large spatial
103 area (Lambert and Wagner, 2018; Arias-Andres *et al.*, 2019).

104 Microplastics encompass a highly diverse group of materials (e.g. Polyethylene
105 Terephthalate (PET), Polystyrene (PS) and Polypropylene (PP)), morphologies (e.g. fragments,
106 fibres and beads), colours and sizes (usually 1 µm to 5 mm). Microplastics are classified as
107 being primary (manufactured at micro-size) or secondary (smaller fragments that have been
108 eroded or weathered from the larger plastics) (Eerkes-Medrano *et al.*, 2015; Horton *et al.*,
109 2017b; Sharma and Chatterjee, 2017). Despite the potential significance of microplastic
110 pollution, much remains unknown about its sources, pathways, fate and impacts on receptors.
111 Major sources of plastic wastes and the burning and breakdown of those mismanaged larger
112 plastic litter (e.g. via microbeads, fibres, etc.) is estimated to be the largest contributor of
113 microplastics (Horton *et al.*, 2017b; Conley *et al.*, 2019), and is likely to be greater in urban
114 areas. Urban freshwater environments may therefore be of great significance as a source of

115 microplastics, rapidly conveyed from urban discharge into fluvial systems and eventually to
116 marine systems via estuaries and deltas (Zhao *et al.*, 2014).

117 Zhao *et al.* (2014) were the first study to quantify microplastic pollution in freshwater
118 environments in China, but there has been limited subsequent research. Some studies of Chinese
119 freshwater bodies have reported higher microplastic concentrations than many other countries
120 (Su *et al.*, 2016; Zhang *et al.*, 2018; X. Jiang *et al.*, 2019). Rivers in East Asia are predicted to
121 carry the highest annual microplastic loads by 2050, approximately four times higher than the
122 microplastic emission from other OECD (Organization for Economic Cooperation and
123 Development) countries (van Wijnen *et al.*, 2019). Because of its size and population, China
124 plays the most important role in the East Asia region. Chinese cities have large populations and
125 substantial plastic use, but often with poor disposal management and limited knowledge of
126 microplastic concentrations in urban freshwater environment. China's current legal framework
127 still does not specifically cover the management of microplastic (Zhang *et al.*, 2019).
128 Legislation aimed at reducing microplastic pollution can have multiple positive effects in
129 China's, but an essential prerequisite for this is awareness of contamination sources, pathways
130 and levels.

131 Microplastic pollution has been reported worldwide and, according to the reviews of
132 Eerkes-Medrano *et al.* (2015) and Horton *et al.* (2017b), is constantly increasing in freshwater
133 environments. Zhang *et al.* (2018) were the first to review what is known of microplastic
134 pollution in Chinese inland water systems; they also considered the State-of-the-Art approaches
135 for sampling microplastics. Fu and Wang (2019) reviewed research approaches, characteristics,
136 sources and fate of microplastics in the Chinese freshwater environments and provided
137 recommendations for the Chinese Government and public to reduce freshwater microplastic

138 pollution. Fok *et al.* (2020) studied the investigating approaches used in microplastic studies in
139 China, while Fu *et al.* (2020) synthesised knowledge of microplastic pollution in various
140 ecosystems and provided an overview of policies related to plastic and microplastic
141 management in China. These reviews established a thoughtful basis for future development of
142 microplastic controls in the Chinese freshwater environment, but understanding of
143 microplastics in Chinese urban catchments still remains limited. However, available literatures
144 are still general on the discussion of microplastic management and current legislation do not
145 provide specific guidance on developing appropriate policies.

146 Given the circumstances above, this manuscript reviews microplastic pollution in urban
147 freshwater catchments in China, with a particular focus on legal frameworks for managing the
148 problem. The aim of the review is to identify major knowledge and policy gaps that need to be
149 filled to improve understandings of the environmental risks of microplastics. Specific
150 objectives are:

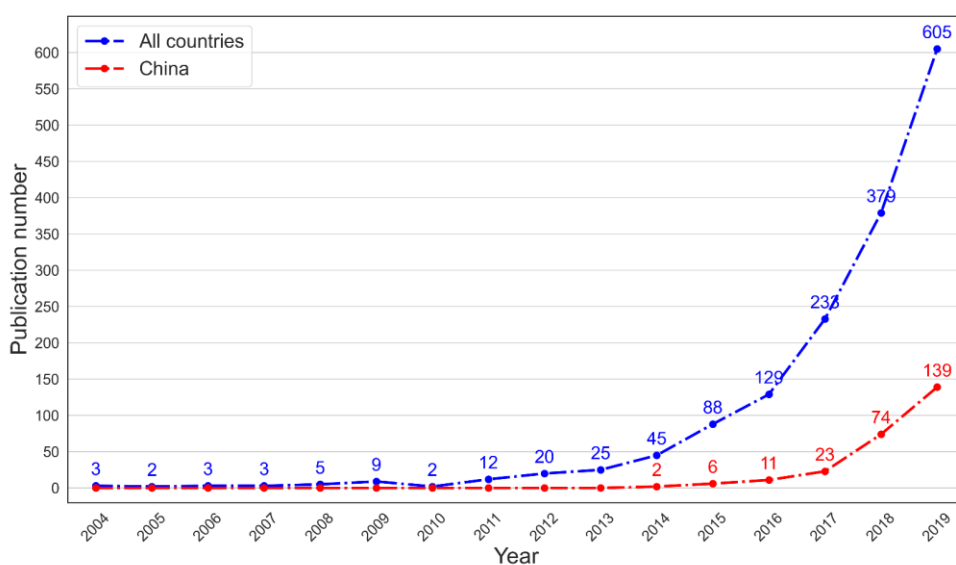
- 151 1) To review what is known of microplastic abundance and characteristics in China's
152 freshwater environment and identify current knowledge gaps, especially related to
153 urban catchments.
- 154 2) To review key existing legislation and policies related to microplastics worldwide
155 and in China.
- 156 3) To provide recommendations for managing microplastic pollution in China's
157 freshwater environments, and specifically for dealing with pollution of urban
158 catchments.

159 **2. Current knowledge of urban freshwater microplastics in China**

160 ***2.1. The foci of China's microplastics research***

161 Following the first use of the term '*Microplastics*' (Thompson *et al.*, 2004), 1563 papers about
162 microplastic pollution have been published (up an and including 2019; Web of Science). The
163 trajectory during the last 16 years reflects a rapidly growing concern about microplastics
164 worldwide (Fig. 1). The first data of microplastic loads in China were not published until 2014,
165 when Zhao *et al.* (2014) reported on loads in both freshwater and seawater zones of the Yangtze
166 River Estuary System. By the end of 2019, 255 papers concerning microplastics in China had
167 been published, with these using a wide variety of sampling and sample processing approaches
168 (Zhang *et al.*, 2018). Since then, China has become a significant contributor to the literatures,
169 producing more than 16% (according to Web of Science) of the global microplastics research
170 (China is ranked first, followed by USA, UK and Germany). This indicates that China is starting
171 to play an important role in the understanding of microplastics, and consequently may influence
172 future research directions. This section reviews Chinas 255 contributions to the global
173 literature.

174



175

176 *Figure 1. Number of academic publications about microplastics from 2004 to 2019 (data source: Web of Science).*

177 *Blue bar represents global annual publication numbers and orange bars means yearly publication numbers from*
 178 *China.*

179

180 The VOSviewer software (Leiden University, Netherlands) was used to provide an
 181 overview of current microplastics research in China (Figure 2). The academic terms repeatedly
 182 occurring in each paper were collected from the 255 publications and analysed in this software.
 183 The frequency of occurrence of each keyword and the co-occurrence of pairs of keywords were
 184 used to indicate the foci of published work; in the resulting schematic (Fig. 2), the foci of
 185 proximity between terms indicates the frequency of co-occurrence, the size of each term
 186 illustrating the occurrence frequency, and the colours indicate temporal patterns. This analysis
 187 demonstrates that the ‘*Marine Environment*’ was a very common focus and was closest to
 188 ‘*Microplastics*’ and ‘*Pollution*’, indicating that a large proportion of microplastics research in
 189 China (and globally) has been conducted in marine environments. The red colour of

204 (database: Web of Science) from 2014 to 2019, where colours represent the average publication time of each
 205 keyword (blue to red: early to current). Database: Web of Science from 2004 to 2019.

206 **2.2. Microplastics in freshwaters in China**

207 To further understand current progress in microplastics research in China's freshwater
 208 environments, twenty-one papers reporting microplastic loads in various types of Chinese
 209 freshwater environments were analysed in detail (Table 1).

210 Microplastic pollution has been investigated in freshwater systems including the Poyang
 211 Lake (the largest freshwater lake in China), Dongting Lake (the second largest freshwater lake
 212 in China), Qinghai lake (the largest inland lake in China), Yangtze River (the largest river
 213 catchment in China), Pearl River (the largest river catchment in Southern China), and other
 214 large waterbodies in the country (Zhao *et al.*, 2014; Lin *et al.*, 2018; Wang *et al.*, 2018; Xiong
 215 *et al.*, 2018; Zhou *et al.*, 2018; Yuan *et al.*, 2019). These studies provided quantitative evidence
 216 of microplastic pollution but so far none has investigated loads in any environments in an
 217 integrated way, so as to characterise contamination in waterbodies, sediments and biota;
 218 moreover, small- to medium-sized freshwater systems remain under-represented.

219

220 *Table 1. Twenty-one publications involving investigations of microplastic abundances in freshwater environments*
 221 *in China*

| No. | Location | Sampling time | Abundances | Dominant types | Citations |
|-----|---|-------------------|---|---|-----------------------------|
| 1 | The Yangtze Estuary System and East China Sea (involving urban catchment) | July-August, 2013 | Freshwater (1 m depth surface water): 500-10200 n/m ³ (average: 4137.3 n/m ³) Marine water (333 µm mesh-size neuston net for surface layer water): 0.03-0.455 n/m ³ (average: 0.167 n/m ³) | Materials: N/A Colours: transparent in freshwater and coloured in marine water Shape: fibres Size: 0.5-2.5mm | (Zhao <i>et al.</i> , 2014) |

| | | | | | |
|---|--|-----------------------------------|---|---|--------------------------------|
| 2 | Mingjiang, Jiaojiang and Oujiang estuaries (involving urban sections) | July, 2013 | Surface water (30 cm depth): Minjiang-A: 1245.8±531.5 particles/m ³ ; Minjiang-B: 1170.8±953.1 particles/m ³ ; Jiaojiang: 955.6±848.7 particles/m ³ ; Oujiang: 680.0±284.6 particles/m ³ | Materials: PP and PE Colour: coloured Shape: fibres and granules Size: <2mm | (Zhao <i>et al.</i> , 2015) |
| 3 | Beijiang River littoral zone, Qingyuan City (involving urban section) | March, 2015 | Surface layer of sediments (top 2 cm): 178±69 items/kg to 544±107 items/kg | Materials: PE, PP, and copolymer Colour: blue Shape: particles Size: N/A | (J. Wang <i>et al.</i> , 2017) |
| 4 | Siling Co Basin, Tibet | May-June, 2015 | Sediment (top 2 m): Siling Co: 4-1219 items/m ² ; Geren Co: 42±47 items/m ² ; Wuru Co: 117±126 items/m ² ; Mujiu Co: 17±20 items/m ² | Materials: PE and PP Colours: N/A Shape: N/A Size: 1-5mm in the most sits excluding in Wuru Co. | (Zhang <i>et al.</i> , 2016) |
| 5 | The Taihu Lake (involving urban catchment) | Aug, 2015 | Plank samples: 0.01x10 ⁶ -6.8x10 ⁶ items/km ² Surface water (less than 0.3 m deep): 3.4-25.8 items/L Sediments: 11.0-234.6 items/kg (dry weight) Clams: Summer (Aug.): 1.3-12.5 items/kg; Winter (Nov.): 0.2-9.6 items/kg | Materials: cellophane and PET Colours: blue in plank and water samples; white/transparent in sediments and organisms Shape: fibres Size: 0.1-1mm in water, sediment and organism samples; 0.333-5mm in plank samples | (Su <i>et al.</i> , 2016) |
| 6 | Xiangxi River, Yangtze River, TGR area, Hubei Province (involving county/town sections) | Apr, July and Oct 2015; Jan, 2016 | Surface water: 0.55x10 ⁵ -3.42x10 ⁵ items/km ² Sediment: 80-864 items/m ² Fish: none or 0.33-1.5 items in each fish | Materials: PP in water and sediments, PE in fish Colour: blue in sediments Shape: sheet and fragment in water, sheet in sediment Size: 1-5mm in water and sediment | (Zhang <i>et al.</i> , 2017) |
| 7 | Surface water of 20 major urban lakes and urban sections of Yangtze River and Hanjiang River in Wuhan city, China (urban catchments) | April 2016 | Surface water (0-20 cm depth): From 1660.0±639.1 n/m ³ to 8925±1591 n/m ³ in different waterbodies | Materials: PET and PP Colours: coloured Shape: fibre Size: <2mm | (W. Wang <i>et al.</i> , 2017) |
| 8 | Qinghai Lake (inland waterbody) | July, 2016 | Surface water (112µm mesh size net): 3090-757,500 particles/km ² Sediment (top 2 cm): 50±50 to 1292±582 particles/m ² Fish: 5.4±3.6 particles/individual | Materials: PP and PE Colour: coloured Shape: fibres Size: 0.1-0.5mm in surface water of lake; 1-5mm in river samples | (Xiong <i>et al.</i> , 2018) |
| 9 | Middle- | Aug-Oct, | Surface water (0-12cm depth): 0.5-3.1 | Materials: PS | (Su <i>et al.</i> , |

| | | | | | |
|----|---|--|--|--|-----------------------------|
| | lower Yangtze River Basin (involving urban sections) | 2016 | particles/L Sediment (top 10 cm): 15-160 particles/L Asian clam: 0.3-4.9 particles/g or 0.4-5.0 particles/individual | Colour: transparent and blue Shape: fibres Size: 0.25-1mm | 2018) |
| 10 | Poyang Lake section of Le'an River (involving industrial and residential areas) | Dec, 2016 | Sediment (5cm depth): Average: 1800 ind/kg; upper stream: 1121 ind/kg; branch stream: 2871 ind/kg; downstream: 1366 ind/kg | Materials: PE Colour: white Shape: fragment Size: <1mm | (Zhou <i>et al.</i> , 2018) |
| 11 | Mainstream Pearl River, its three major tributaries and Pearl River Estuary | March to May, July to August, November to January, 2016-2017 | River water: 0.57±0.71 items/L Riverbed sediments: 685±342 items/kg (dry weight) Estuarine sediment: 258±133 items/kg (dry weight) | Materials: PP Colour: white and transparent Shape: sheet Size: <0.25mm | (Fan <i>et al.</i> , 2019) |
| 12 | Changjiang Estuary (CE) and the East China Sea (including urban sections) | Feb, May and July, 2017 | Surface water (30 cm depth): Changjiang Estuary: 157.2±75.8 n/m ³ ; The East China Sea: 112.8±51.1 n/m ³ | Materials: PE and PA Colour: N/A Shape: fibres Size: < 1000 µm | (Zhao <i>et al.</i> , 2019) |
| 13 | The urban section of the Pearl River along Guangzhou City (urban catchment) | July 2017 | Surface water (50 cm depth): 379-7924 items/m ³ (average: 2724 items/m ³) Sediment (top 5 cm): 80-9597 items/kg (average: 1669 items/kg) Influent of wastewater treatment plant: 0.5-4.2 items/L Effluent: 0.3-2.7 items/L | Materials: PP and PE in surface water and sediments Colours: white in surface water samples and yellow in sediments Shape: fibres in surface water, sediments and effluent samples Size: 0.02-1mm in surface water and sediment samples | (Lin <i>et al.</i> , 2018) |
| 14 | Dongting Lake and Hong Lake (involving urban catchments) | September 2017 | Dongting Lake (0-20 cm depth surface water): 900-2800 n/m ³ (average: 1911.7 n/m ³) Hong Lake (0-20 cm depth surface water): 1250-4650 n/m ³ (mean: 2282.5 n/m ³) | Materials: PP and PE Colours: coloured Shape: fibres Size: 0.05-0.33mm | (Wang <i>et al.</i> , 2018) |
| 15 | Poyang Lake (involving urban sections) | Nov, 2017 | Surface water (0-1 m depth): 5-34 items/L Sediment: 54-506 items/kg (dry weight) Fish: 0-18 items/fish | Materials: PP and PE Colours: coloured Shape: fibres Size: 0.1-0.5mm in water, 0.1-0.4mm in sediment and 0.5-1mm in fish | (Yuan <i>et al.</i> , 2019) |
| 16 | Wei River, Wushan County, Gansu Province (including county sections) | Winter of 2017 | Surface water: 3.67-10.7 items/L Sediments: 360-1320 items/kg | Materials: PET Colour: N/A Shape: fibres Size: <0.5 mm | (Ding <i>et al.</i> , 2019) |

| | | | | | |
|----|--|-------------------------------------|---|--|---------------------------------|
| 17 | Poyang Lake, Jiangxi Province (involving urban sections) | December 2016 and April, July, 2018 | Sediments (top 2 cm): 11-3153 items/kg (dry weight); average: 1134 items/kg (dry weight) | Materials: N/A Colour: N/A Shape: fragments Size: <1mm | (Liu <i>et al.</i> , 2019) |
| 18 | Buqu River, Naqu River, Lhasa River, Brahmaputra River, and Nyang River in Tibet Plateau | July, 2018 | Surface water: 483-967 items/m ³ Sediments (top 2 cm): 50-195 items/kg | Materials: PET Colour: transparent Shape: fibres Size: <1mm | (C. Jiang <i>et al.</i> , 2019) |
| 19 | Yangtze Delta area, Shanghai (urban sections) | April-September, 2018 | Overall surface water: 0.08-7.4 items/L; Freshwater: 1.8-2.4 items/L Coastal and estuarine water: 0.9 items/L | Materials: Shape: fibres Colours: blue and red Size: 100-1000 μ m | (Luo <i>et al.</i> , 2019) |
| 20 | Middle and lower reaches of Yangtze River (involving urban sections) | Aug-Sep, 2018 | Surface water: 240 items/m ³ to 1800 items/m ³ Sediments: 90-580 items/m ³ | Materials: PP Colour: blue Shape: fibres Size: <1mm | (Li <i>et al.</i> , 2019) |
| 21 | Urban lakes in Changsha (urban waterbodies) | N/A | Surface water: 2425 \pm 247.5 to 7050 \pm 1060.6 items/m ³ | Materials: PP and PE Colour: transparent Shape: linear Size: <2mm | (Yin <i>et al.</i> , 2019) |

222

223 The properties of microplastics are important indicators of sources. Microplastics
224 detected in Chinese freshwaters consists of diverse materials including Polypropylene (PP),
225 Polystyrene (PS), Polyethylene Terephthalate (PET (also abbreviated PETE)), Polyethylene
226 (PE) and Polyvinyl Chloride (PVC). PP and PE were dominant in most investigated freshwater
227 environments (PP in 12 waterbodies and PE in 11 waterbodies; Table 1). This condition fits
228 with the current state of the Chinese plastics market, with the annual yields of PP and PE
229 accounting for 27.21 and 30.04 million tonnes and representing more than 40% and 30% of
230 global totals respectively by 2018 (Yin and Zhang, 2019; Zhang, 2019). Although published
231 research used different sorting strategies, plastic fibres were dominant in 14 cases (Table 1).
232 Size ranges of microplastics were also variable, with some suggesting smaller microplastics
233 (e.g. < 1 mm) are disproportionately abundant (Su *et al.*, 2016; Yuan *et al.*, 2019).

234

235 Microplastic abundance also varied substantially between different geographic areas.
236 For example, concentrations in surface water of rivers on the Tibet Plateau were 483 - 967
237 particles/m³ compared to 4,137.3 particles/m³ in the Yangtze Estuary (Table 1). Such cases also
238 imply that densely populated areas (e.g. urban areas) have higher microplastic concentrations
239 compared to remote areas. Concentrations are also variable over smaller geographic areas. For
240 instance, within the same catchment, Lin *et al.* (2018) found that concentrations along the urban
241 section (Guangzhou) of the Pearl River varied from 379 to 7,924 particles/m³ in surface water
242 (Table 1); a concentration of 0.167 particles/m³ was recorded offshore from the Yangtze River
243 in comparison to 4,137.3 particles/m³ in the Yangtze estuary (Zhao *et al.*, 2014). The spatial
244 variation in the Yangtze is likely due to dilution (Mendoza and Balcer, 2018; Wang *et al.*, 2018).
245 No consistent relative patterns have been found in microplastic concentrations in water versus
246 sediment in fluvial or limnetic environments (Zhou *et al.*, 2018; Ding *et al.*, 2019; Zhao *et al.*,
247 2019), which suggests complex sources and pathways and patterns of accumulation.

248

249 The properties of microplastics, hydrological conditions, surroundings and
250 meteorological conditions have been investigated as the four major factors influencing
251 microplastics' distribution patterns in waterbodies in China (W. Wang *et al.*, 2017).
252 Unfortunately, no clear patterns or consensus has yet been reached. For example, recent studies
253 indicated that the lower microplastic concentrations during the wet season (summer) in the Pearl
254 River, China could be attributed to dilution by higher precipitation and river flows (Fan *et al.*,
255 2019). Conversely Zhao *et al.* (2019) found an increased concentration during wet periods,
256 presumably due to more runoff washing plastic particles into waterbodies. As with suspended
257 sediment, microplastic concentrations in recorded river water may be dictated by preceding

258 conditions (e.g. whether a preceding flood has caused washout), such discharge-concentration
259 relationships are not ways clear.

260

261 Sources differ greatly from place to place. For instance Peng *et al.* (2018) explained that
262 polyester, rayon and other fibres detected in their samples were from clothes washing (based
263 on the materials and shapes of microplastics), while Xiong *et al.* (2018) believed that the PE
264 and PP microplastics detected in the Qinghai Lake, China originated from tourists, due to these
265 types plastics being commonly used in packaging. Nevertheless, a problem that may contribute
266 to different interpretations of sources and loads is the lack of consistent collection, identification
267 and analytical approaches (Luo *et al.*, 2019). Zhao *et al.* (2015) used a 333 μm pore-size sieve
268 to filter water samples from the Minjiang, Jiaojiang and Oujiang estuaries, while Wang *et al.*
269 (2018) selected a 50 μm mesh-size steel sieve for the investigation in the Dongting Lake and
270 the Hong Lake, China. In most microplastics studies in Chinese freshwater environments,
271 smaller-sized microplastics ($< 2\text{ mm}$) are usually most abundant (see Table 1). Such differences
272 cause large variation in reported concentrations.

273 ***2.3. Microplastic dynamics in Chinese freshwater environments***

274 Rivers are often conceptualised as conveyor belts, transporting water, sediments and
275 contaminants to the oceans. However, transport is not continuous, with material stored
276 periodically (e.g. sediment deposited and stored for a period of time on the riverbed, before the
277 next competent event leads to its onward conveyance). This and the different course areas –
278 some of the, point sources and some of them diffuse - may underpin spatial variation reported
279 to date. Wang *et al.* (2018) reported with high concentrations at the confluence between the
280 Dongting Lake and Yangtze River. Based on work in the Pearl River, Lin *et al.* (2018) argued

281 that tributaries transport microplastics to the mainstream and result in high concentration in
282 confluence zones. Peng et al. (2018) observed that the microplastics transported by the Yangtze
283 River and Huangpu River stagnated and accumulated at the plume front area formed between
284 freshwater and seawater of the East China Sea. These findings not only indicate the significance
285 of Chinese fluvial systems as a pathway, delivering microplastics towards trunk streams and
286 marine environments, but also illustrate that confluence areas may be contamination ‘hotspots’.

287

288 The transportation of microplastics may be critical to assessing and understanding
289 health risks. Because of their lipophilic features and high surface area to volume ratio, which
290 enable them to absorb chemical pollutants, including persistent organic pollutants (e.g.
291 pesticides and antibiotics), as well as pathogenic bacteria, fungi and viruses (Zou *et al.*, 2017),
292 microplastics pose risks to ecosystems and human health. Additionally, toxic plastic additives,
293 such as flame retardants, pigment and ultraviolet stabilizer can be release once plastics are in
294 the freshwater environment (Gabriella, 2019). Risk partly depends storage dynamics. For
295 instance, exposure of benthic organisms to microplastics depends how much material may be
296 stored within the bed matrix, and the duration of residence here before being remobilised during
297 high flows (van Cauwenberghe *et al.*, 2015). Residence times are likely to be longer in stable
298 sediments that are infrequently disturbed, so lentic environments can act as a longer term store
299 for microplastics than fluvial environments (Eerkes-Medrano *et al.*, 2015). Movement of
300 microplastics can also be altered by absorption of other materials or colonisation by microbial
301 communities, changing particle density and causing microplastics to settle more easily (Lin *et*
302 *al.*, 2018; Wang *et al.*, 2018; Fan *et al.*, 2019). Research foci are now changing from simple
303 assessment of loads to efforts understand pathways, including those in the subsurface (i.e. via

304 hyporheic zone and groundwater), although such work still remains limited in China (Zhao *et*
305 *al.*, 2015; Su *et al.*, 2016).

306

307 River Basin Management for navigation purposes, alongside with other factors such as,
308 water supply, flood control and hydropower production alter flow regimes and flow hydraulics,
309 and hence, may modify transport and storage dynamics of microplastics. Dams and reservoirs
310 trap sediments, and therefore are likely to also trap microplastics (Crawford and Quinn, 2017b).
311 Microplastics accumulated in stable freshwaters will not stop degrading, potentially generating
312 and releasing smaller-sized, secondary microplastics that that may be more easily ingested by
313 organisms.

314

315 Another issue relates to the temporal dimension. Most studies lack long-term, repeat
316 sampling and measurement. This is important because, when studied, significant temporal
317 differences in microplastic concentrations have been found to exist; this raises concerns over
318 the representativeness of single date or spot sampling (Stanton *et al.* 2019). Such temporal
319 differences may arise due to seasonal and/or weather-related factors, or dues to activities of
320 organisms (Crawford and Quinn, 2017b). For example, in China cyclonic effects (i.e. typhoons)
321 have been reported to increase microplastic loads in freshwaters and conveyance to the marine
322 environment (e.g. especially along the Southern and Eastern coastline of China; Wang *et al.*,
323 2019). Studies are needed in China to elucidate the causes of temporal variation in microplastics
324 and the implications of this for accurate assessment of loads and risk.

325

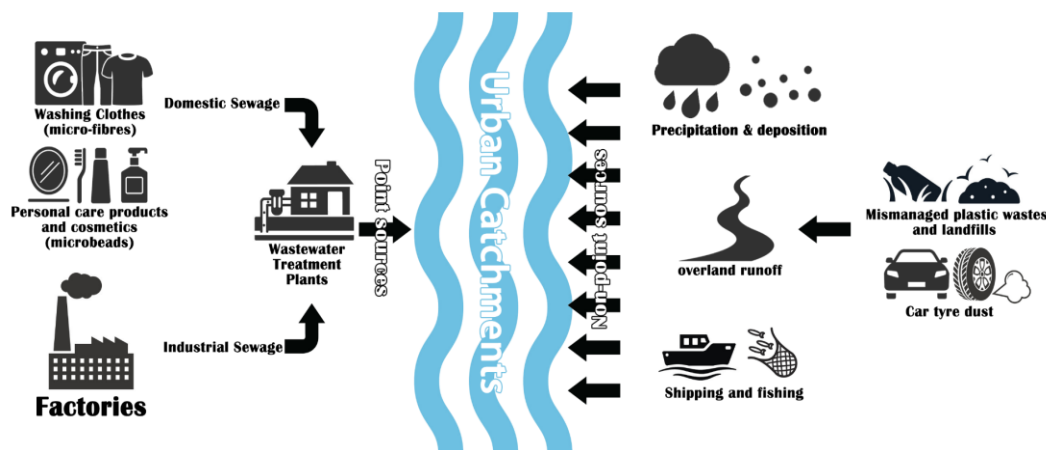
326 **2.4. Sources of microplastics in urban freshwaters**

327 Urban areas play a key role as sources of microplastics in China (Zhang *et al.*, 2018; Fu and
328 Wang, 2019). These sources are summarised in Figure 3. Primary microplastics are usually
329 discharged from industrial areas associated with plastic production or residential areas through
330 wastewater treatment plants (WWTPs) (Figure 3). Microplastics could be removed from
331 sewage during the primary treatment and adsorbed by activated sludge during secondary
332 treatment in WWTPs, but even if removal reached 95-99%, as recorded in some developed
333 countries, the remaining material released to the environment could still be problematic
334 (Talvitie *et al.*, 2015). Many WWTPs do not attain 99% removal efficiency of microplastics in
335 China (Lin *et al.*, 2018); for example, a municipal WWTP equipped with activated sludge
336 techniques in Wuhan only removed about 64.4% of microplastics (Liu *et al.* 2019). Another
337 issue is that 60% of polluted sludge from WWTPs in China is disposed of in landfill sites (Sun
338 *et al.*, 2019). Microplastics can leach from landfills and find routes through soils to contaminate
339 freshwaters (He *et al.*, 2019). Thus, due to the relatively underdeveloped techniques and poor
340 management of waste treatment, WWTPs play significant roles in Chinese urban microplastic
341 pollution, as documented by W. Wang *et al.* (2017) and Lin *et al.* (2018), respectively.

342 Secondary microplastics usually reach freshwater environments through non-point
343 sources. In Chinese urban areas, mismanaged plastics are the major terrestrial non-point source,
344 which includes dumping and littering of solid wastes (W. Wang *et al.*, 2017; Zhou *et al.*, 2018;
345 Fan *et al.*, 2019). Fragmentation of large plastic waste generates secondary microplastics on
346 land (Zhang *et al.*, 2015), while road dusts containing car tyre fragments contribute to loads
347 entering freshwater systems as a result of runoff (Zhang *et al.*, 2018). Atmospheric dispersal of
348 microplastic may also be important in urban areas (Dris *et al.*, 2016). For example, Zhou *et al.*

349 (2017) found microplastic fibres in atmospheric deposition in Yantai, China. Shipping and
 350 fishing in urban catchments can also release microplastics directly to aquatic environments,
 351 where fragments and fibres from fishing nets or gear have been observed in urban waterbodies
 352 in Changsha City (Yin *et al.*, 2019).

353



354

355 *Figure 3. Potential microplastic sources towards urban catchment (source by Yuyao Xu)*

356

357 Previous literature has assessed factors influencing microplastic pollution in Chinese
 358 urban catchments. Fan *et al.* (2019) found a direct linear relationship ($R^2 = 0.772$) between
 359 microplastic abundance in water samples and population density in the Pearl River catchment,
 360 where large population centres generated and released more microplastics. Similarly, Wang *et al.*
 361 *al.* (2017) found an inverse linear relationship ($p < 0.001$) between the distance from an urban
 362 centre and microplastic concentration in surface waters of the Yangtze River. Other
 363 investigations have, however, failed to find such relationships (e.g. Shanghai) (Peng *et al.*,
 364 2018).

365 Zhao *et al.* (2015) suggested that different economic structures might lead to different
 366 microplastic sources and abundances; interestingly this hypothesis was supported by the Gross

367 Domestic Product (GDP) and microplastic data presented by Fan *et al.* (2019). Li (2020)
368 reported microplastic concentrations in urban runoff from residential roads were significantly
369 higher than from parking lots and cement pavements, which also indicates that local land-use
370 conditions will affect microplastic pollution levels in urban areas.

371 Rapid urbanisation in China may increase pressures on urban freshwaters, including
372 contamination by microplastics (Zhang *et al.*, 2018). Therefore, understanding the roles of land-
373 use, population density and local economic structure in influencing microplastic distribution
374 patterns in urban waterbodies is an important first step to developing policy measures designed
375 to minimise risk. Taking a precautionary approach via implementation of legislative enactments
376 and guidelines to control microplastics is a growing area of interest in China. This is the focus
377 of the section that follows.

378

379 **3. Management and Legislations on microplastics**

380 ***3.1. General legislation***

381 Plastic microbeads (10 – 500 µm in diameter), used in personal care products (PCPs) (Sharma
382 and Chatterjee, 2017), are a type of primary microplastics. To help reduce risks posed by
383 microbeads to wildlife, Five European countries (i.e. Netherlands, Austria, Luxembourg,
384 Belgium and Sweden) issued a joint statement calling for banning the use of microbeads in
385 PCPs (see Table 2). In 2014, the State Government of Illinois (USA) enacted the first
386 prohibition of production and sales of PCPs that contain microbeads (see Table 2), which
387 subsequently led to the US ‘*Microbead-Free Water Act of 2015*’ (see Table 2). In 2015 Canada
388 also limited the addition of microbeads in PCPs, by adding microbeads as a new toxic substance
389 to the 1999 ‘*Environmental Protection Act*’ (Table 2). European nations also expressed the

390 concerns about Microbead pollution in cosmetics through the proposal ‘*Cosmetics Europe*
 391 *Recommendation on Solid Plastic Particles (Plastics Micro Particles)*’ (Table 2). This
 392 recommendation has implications for future legislation worldwide (i.e. Microbeads in Toiletries
 393 Regulations by Canada in 2017). Following these first pieces of legislation, more countries
 394 (including the UK, France, South Korea, Italy, New Zealand, India and South Africa) joined in
 395 this ‘*Microbead-free*’ action (see Table 2), which will make efforts to reduce the global release
 396 of microbeads. Nonetheless, microbeads are only a small part of the total microplastic load,
 397 approximately accounting for 0.1~ 4.1% (McDevitt *et al.*, 2017) of total global microplastic
 398 pollution in aquatic environments, so broader action is also needed.

399 In October 2019, the Chinese government officially issued the “Economic Structure
 400 Adjustment Guidance Catalogue”, which included the prohibition of light plastic bags (<
 401 0.025mm thickness), disposable-foamed plastic tableware, disposable plastic swabs, daily
 402 chemical products containing microbeads and polyethylene agricultural films (< 0.01 mm
 403 thickness) (Table 2). This catalogue not only highlights microbeads problem in both PCPs and
 404 cosmetics, but also looks at other commonly used in China but potentially polluting plastic
 405 products (NDRC, 2019).

406

407 *Table 2. Brief history of ‘microbeads-free’ activities worldwide (The word ‘Microbead-free’ was from the*
 408 *‘Microbead-free Act’ in USA and used to represent the popularization of banning plastic microbeads in*
 409 *relevant products in different countries in this paper)*

| No. | Time | Laws/Policies | Legislator | Legislative aims | Comments | Citations |
|-----|------|--|-----------------------------|--|---|-----------------------------|
| (1) | 2013 | Joint call to EU for banning of microbeads | Five European Countries, EU | Issuing the public concern about the risks of microbeads in PCPs | This call (by the Netherlands, Austria, Luxembourg, Belgium and Sweden) followed a green paper by | (Crawford and Quinn, 2017a) |

| | | | | | | |
|-----|------|---|----------------------|--|--|--|
| | | | | | EC published in 2013. | |
| | | | | | It led to the following legislative actions on microbeads in Europe and even worldwide. | |
| (2) | 2014 | Legislation banning microbeads | of Illinois, USA | Banning the manufacture (by 2017) and sale (by 2018) of PCPs containing microbeads | Illinois was the first state enacting bans of microbeads in America. The state government also left certain periods for industries and market to prepare. | (Chicago Tribune, 2014; McCormick <i>et al.</i> , 2014; McDevitt <i>et al.</i> , 2017) |
| (3) | 2015 | A proposal under the <i>Canadian Environmental Protection Act</i> (1999) | Canada | Adding microbeads into a list of toxic substances and intending to ban the use of microbeads in PCPs | Canada was the first country who regarded microbeads as toxicities in the world. It is an alternative way to ban microbeads, which avoids superabundant legislative processes of structuring a new policy/law. | (Canada, 1999, 2015) |
| (4) | 2015 | <i>Cosmetics Europe Recommendation on Solid Plastic Particles (Plastic Micro Particles)</i> | Cosmetics Europe, EU | Suggesting banning the use of microbeads in cosmetic products | This is the first official concern about the plastic microbeads in cosmetics. Then, the microbead-free activities started to take cosmetic products into considerations. | (Cosmetics Europe, 2015) |
| (5) | 2015 | <i>Microbead-Free Water Act of 2015 (H.R.1321)</i> | USA | Prohibiting the manufacture and preventing the inputs of rinse-off PCPs and cosmetics containing microbeads into interstate commerce | After ten states banned microbeads and 15 states introduced relevant plans, President Obama signed the microbead-free act. The microbead-free legislation was from state governments at beginning then to the federal government. | (Pallone, 2015; McDevitt <i>et al.</i> , 2017) |

| | | | | | | |
|------|------|--|---------------|---|---|------------------------------|
| | | | | | The local-to-center legislation route is a good example for Chinese relevant legislation. | |
| (6) | 2017 | A unique commercial testing service to assess ecological impacts of microbeads environmental impact assessment | Intertek, EU | Providing an available technique to assess the environmental risks of microbeads in relevant products | This was the first and only commercial service that combined ecological impact assessment and multiple samples polymer testing for microbeads. | (Intertek, 2017) |
| (7) | 2017 | <i>Microbeads in Toiletries Regulations</i> | Canada | Prohibiting the manufacture, import and sale of toiletries used to exfoliate or cleanse that contain plastic microbeads, including non-prescription drugs and natural health products | It was another use of toxic substance list for banning microbeads in relevant products. This was also the first formal legislation about the microbeads in toiletries in the world. | (Canada.ca, 2017) |
| (8) | 2017 | EU Eco-label for detergent | EC, EU | Updating the standards of EU Eco-label for detergent products with the content of microbeads as a new indicator | The products awarded the ecolabel afterwards will not contain plastic microbeads after this updating. | (Zhang <i>et al.</i> , 2019) |
| (9) | 2017 | Prohibition of sales cosmetics containing microbeads | South Korea | Banning the using and selling of cosmetics containing plastic microbeads | South Korea was the first Asian country issuing microbead-free police. | (BeatTheMicrobead, 2020) |
| (10) | 2017 | Ban of microbeads | Taiwan, China | Banning the application of microbeads in all cosmetic products | Taiwan was first area of China promoting microbeads relevant legislation. | (BeatTheMicrobead, 2020) |
| (11) | 2017 | <i>The Environmental protection (Microbeads) (England) Regulations 2017</i> | UK | Banning the manufacture and sale of rinse-off PCPs containing microbeads | After England microbeads regulations, similar regulations also issued in Scotland and Wales in 2018. | (Newground LUS, 2018) |
| (12) | 2018 | <i>The Decree 2017-2019</i> | France | Banning the using of microbeads in PCPs and the producing of plastic cotton buds | France was the first country to pass a legislation about plastic cotton buds productions. | (Zhang <i>et al.</i> , 2019) |

| | | | | | | |
|------|------|--|--------------|--|--|--------------------------|
| (13) | 2018 | Draft legislation to ban microbeads in rinse-off PCPs as well as cosmetics and plastic cotton buds | Italy | Banning the using of microbeads in PCPs and cosmetics and the producing of plastic cotton buds | The ban of plastic cotton buds will take into force on the first day of 2019, which made Italy the first country enforced plastic cotton bans. | (BIOPLASTICS NEWS, 2019) |
| (14) | 2018 | Ban of microbeads in rinse-off PCPs | New Zealand | | | (BeatTheMicrobead, 2020) |
| (15) | 2018 | Microbeads ban | India | Announcing microbead ban, which will enter force in 2020 | India was the first developing country contributing to the microbead-free activities. | (BeatTheMicrobead, 2020) |
| (16) | 2018 | Proposal of microbeads ban | South Africa | Issuing the proposal for microbeads ban legislations | After finding microbeads occurrence in tap water, the government of South Africa claimed their concerns. This is also the first step of microbead-free activities in Africa. | (BeatTheMicrobead, 2020) |
| (17) | 2019 | <i>Industrial structure adjustment guidance catalogue (2019) (exposure draft)</i> | China | Listing light plastic bags (0.025mm), disposable foamed plastic tableware, microbeads toothpaste, rinse-off PCPs and cosmetics as the obsoleted products | This document is a suggestion file, but it represents the concerns from Chinese government. | (NDRC, 2019) |

410 Note: (The word ‘microbead-free’ was from the ‘Microbead-free Act’ in USA and used to represent the
411 popularization of banning plastic microbeads in relevant products across different countries in this paper)

412

413 As yet, ‘*Microplastics*’ have not been adopted as a formal legislative object in any
414 national or international laws (Zhang *et al.*, 2019). Nevertheless, because of the concern about
415 marine microplastic pollution, several international conventions take microplastics into account
416 (Crawford and Quinn, 2017a). The ‘*Oslo-Paris Convention for protecting and conserving the*
417 *North-East Atlantic and its resources (OSPAR)*’ uses microplastic abundance in seabird

418 stomachs as one indicator of marine ecological quality (see Table 3). In 2014, the United
 419 Nations Environment Assembly (UNEA) also identified microplastics as an emerging marine
 420 pollutant, which has placed marine microplastic management on the agenda of many countries
 421 worldwide (see Table 3). China has enacted a national marine environmental legislative
 422 framework that addressed the issues of shared maritime rights and obligations in the East and
 423 South China Seas, and extends to the management of waste dumping, shipping waste,
 424 construction waste and landfills in territorial waters (see Table 3).

425 Even though microplastics are not mentioned explicitly in these regulations and laws,
 426 their legislative power to reduce solid waste pollution in marine systems covers should help
 427 reduce plastic pollution (Li, 2018; Zhang *et al.*, 2019). Lessons learned from their enactment,
 428 combined with lessons learnt from international conventions, are important for developing
 429 policies related explicitly to microplastics in China.

430

431 *Table 3. The international and Chinese legislation related to plastic and microplastics (mainly for marine*
 432 *environments)*

| No. | Time | Laws/Policies | Legislators | Legislative aims | Comments |
|-----|------|--|---|--|--|
| (1) | 1972 | <i>Convention on the Prevention of Marine Pollution by Dumping of Wastes and Other Matter 1972 (The London Convention; LC'72')</i> | United Nations Conference on the Human Environment (UNCHE); IMO | Controlling marine pollution led by dumping wastes | This convention took into force in 1975. Till 2016, 89 parties signed this convention including China. The London Convention is regarded as a legislative framework of international environmental legislations. |
| (2) | 1973 | <i>International Convention for the Prevention of Pollution from Ships (MARPOL)</i> | IMO | Limiting the dumping and emission of ship garbage, including plastic litters | The convention issued in 1973, the protocol launched in 1978 and this policy finally took into force in 1983. Sometimes, a relatively long period is required for legislations, especially the multilateral policies involving many countries. |

| | | | | | |
|------|------|--|---|---|---|
| (3) | 1982 | <i>United Nations Convention on the Law of the Sea (UNCLOS)</i> | United Nations | Defining the rights and responsibilities of nations with respect to their legal rights in world's oceans, establishing guidelines for businesses, the environments and the managements of marine natural resources. | China signed this convention in 1982. |
| (4) | 1982 | <i>Marine Environment Protection Law of People's Republic of China (Marine Law)</i> | China | Controlling the marine pollution and defining the rights/responsibilities of organizations or individuals to the behaviors within marine environments. | The marine environment law is a basic framework of marine environmental legislation in China. It updated four times in 1999, 2013, 2016 and 2017. |
| (5) | 1985 | <i>Regulations of the People's Republic of China on the Control over Dumping Wastes into the Sea Waters (Dumping Regulations) [Revised]</i> | China | Limiting the dumping of marine waste into sea waters in China. | The regulations updated once in 2017. |
| (6) | 1988 | <i>Regulations of the People's Republic of China on the prevention of environmental pollution by shipbreaking (Shipbreaking Regulation)</i> | China | Controlling the marine pollution led by ship garbage in China | The regulations were updated twice in 2016 and 2017. |
| (7) | 1989 | <i>The Basel Convention on the Control of Transboundary Movements of Hazardous Wastes and their Disposal (the Basel Convention)</i> | The Conference of Plenipotentiaries in Basel, Switzerland | Standardizing the transboundary transportation of solid wastes and prohibiting the illegal transportation of harmful wastes among countries. | Norway government suggested adding 'solid plastic wastes' into the special concern list of wastes in Annex II. |
| (8) | 1990 | <i>Regulations of the People's Republic of China on the prevention and control of pollution damage to the marine environment by land-based pollution</i> | China | Controlling the land-based pollutants from marine ecosystems in China. | Plastic litters were emphasized in this policy. |
| (9) | 1992 | <i>Oslo-Paris convention for the North-East Atlantic Marine Environment (OSPAR)</i> | 15 countries of EU | Protect the ecosystem of west coastal of Europe as well as its catchment | Microplastic occurrence in seabird stomach was regarded as an indicator of EcoQO (marine ecological quality objective) |
| (10) | 1996 | <i>1996 Protocol to the Convention on the Prevention of Marine</i> | LC '72' contracting parties | Update and replace the London convention | The new London Convention (1996) took both marine dumping and |

| | | | | | | | |
|------|------|--|-------------------|--|---|--|--|
| | | | | | <i>Pollution Dumping of Wastes and Other Matter, 1972</i> | according to the status at that time. | land-based solid litters into consideration. It also issued some basic environmental legislative principles, such as 'precautionary approach', 'polluter-pay principle' and 'reverse list'. |
| (11) | 1999 | <i>Elimination of backward production capacity, technology and product catalogue (Patch I)</i> | China | | | Suggest banning the using of disposable foamed plastic tableware | The enforcement of this policy led to some relevant industries or company (producing or selling such plastic products) went into liquidation. |
| (12) | 2002 | Irish plastic bag levy | Ireland | | | Collect tax form the using/consuming of plastic bags | In 2002, Ireland became the first country to impose a plastic bag levy and led to a 90% drop of plastic bag consumption. The plastic bags tax usually cost 5-15 cents each bag and it generate 9.6 million pounds as a green fund supporting environmental projects for Ireland. |
| (13) | 2006 | <i>Regulations on the administration of the prevention and control of pollution damage to the marine environment by marine engineering construction projects</i> | China | | | Control the pollutants leaked during marine engineering constructions. | This policy was established basing on the ' <i>Regulations of the People's Republic of China on the prevention and control of pollution damage to the marine environment by land-based pollution</i> ' and was updated in 2017 and 2018. |
| (14) | 2007 | The San Francisco Plastic Bag Ban | San Francisco, US | | | Limit the use and production of plastic bags | This is the first major law regulating carry-out bags in the US, which desires to reach zero waste by 2020. |
| (15) | 2007 | <i>The Notice of the General Office of State Council on Restricting the Production, Sale and Use of plastic Shopping Bags</i> | China | | | Limit the consumption of plastic bags | Ban the manufacture, sale and usage of plastic bags that are thinner than 0.025mm from 1 st June 2008. |
| (16) | 2011 | <i>The Honolulu Strategy</i> | UNEP and NOAA | | | Establish a global framework for prevention and management of marine litters | This strategy could be regarded as an international legislative transition from the concern of general marine wastes into the concern of small marine litter debris. |
| (17) | 2012 | Global partnership on marine litter (GPML) | UNEP | | | Protect human health and the global environment by | The GPML provided a platform for international |

| | | | | | |
|------|------|---|-------------------------------------|--|---|
| | | | | setting reduction and management of marine litter as its main goal with several specific objectives | collaboration on environmental projects in the future. |
| (18) | 2013 | <i>Industrial structure adjustment guidance catalogue (2013)</i> | China | Update and replace previous catalogues such as elimination catalogue of 1999. | Disposable foamed plastic tableware was not included in this new catalogue. |
| (19) | 2014 | <i>Regulations and decisions adopted by the United Nations Environment Assembly of the United Nations Environment Program at its first session on 27th June 2014</i> | UNEA-1 | Basing on GPML, passed first decisions about marine litters and microplastics. | Since UNEA-1, 'microplastics' as an emerging environmental concern, was first taken into international legislation. |
| (20) | 2015 | <i>Marine Wastes Monitoring and Evaluating Techniques Regulations (in Chinese)</i> | SOA, China | Standardize the classification of marine wastes according to size the materials and monitoring protocols of marine litters | It supplied a technique basis for further research and legislation about microplastic pollution in the future in China. |
| (21) | 2016 | <i>2/11 Marine plastic litter and micro-plastics</i> | UNEA-2 | Encourage governments looking for substitutions for microbeads and the microplastic materials used in PCPs and cosmetics. Call for graduation management for wastes. | UNEA-2 took the microbeads problem into international arena. |
| (22) | 2017 | Set up Marine Waste and Microplastics Research Centre | China | Officially start a national level research of microplastics in China. | Marine Waste and Microplastics Research Centre issued 'Marine microplastic monitoring and evaluation techniques regulations. |
| (23) | 2017 | <i>Foreign Garbage Prohibition: implement of reforming the management system for the import of solid waste plan (in Chinese) (Foreign Garbage Prohibition)</i> | General Office of the State Council | Reduce the pollution led by importing foreign garbage as producers' good | Since 1980s, China have imported a huge amount of solid waste for satisfy the immediate requirement of producers' good. However, foreign garbage has polluted the environments in China. The policy issued for banning the importing of foreign garbage since 2017 and has international influences on global garbage disposal. |
| (24) | 2017 | <i>UNEP/EA.3/Res. 7 Marine litter and microplastics</i> | UNEA-3 | Encourage parties to reduce the marine pollution (especially terrestrial source pollution) by 2015 | |

| | | | | | |
|------|------|--|--------|---|---|
| (25) | 2018 | Strategy on Zero Plastic Waste | Canada | Achieve zero waste object with the vision in a circular economy | It sets stage-targets to reduce the plastic waste per capita per year from 706kg to 490kg by 2030 and to 350kg by 2040. After China issued ‘Foreign Garbage Prohibition’, some developed countries started to establish zero-waste plan or similar plans to reduce the waste amount and encourage higher recycling rates. |
| (26) | 2019 | <i>UNEP/EA.4/L.7 Marine Plastic Litter and Microplastics</i> | UNEA-4 | Pay attention to the impacts of microplastics on wildlife health, ecology, food chain and human health. | The UNEA-4 encouraged the global relevant researchers focusing more on the healthy issues led by microplastics. At the same, UNEA-4 also called for long-term management scheme for microplastics. |

433

434 **3.2. Legislation directly relevant to freshwaters in China**

435 There are multiple legislative efforts to protect and manage freshwater environments in China
436 (see Table 4). Four basic laws established a legislative framework for Chinese freshwater
437 management (“*The Water Pollution Law*”, “*The Water Law*”, “*The Soil and Water*
438 *Conservation Law*”, and “*The Flood Prevention Law*”). Some pieces of legislation relate more
439 specific watersheds or to detailed management plans than to the issue of plastics, while others
440 dealt more with wider hydrological issues connected to economic development (such as flood
441 control, soil erosion, water and soil conservation, and land use demands), rather than tackling
442 the water quality. For example, of 51 rules that make up the “*River Courses Regulations*” only
443 two are concerned with freshwater pollution and wastewater management.

444 Several laws and regulations, including “*the National Water Law*”, have been amended
445 several times in the past 20 years to meet the needs of China’s current development (see Table
446 4). The “*Environmental Quality Standards for Surface Water (GB3838-2002)*”, which was

447 updated from the older version (*GB3838-1998*), was issued in 2002. This provides an appraisal
 448 system to classify the quality of Chinese surface water, and has been applied to fluvial systems,
 449 groundwater, lakes, and irrigation water quality (Table 4). Other more recent frameworks that
 450 aim to improve water quality include the “*Water Pollution Action of 2015*”, the “*Sponge City*
 451 *Program*” (Chan *et al.*, 2018), and the “*River Chief System*”, these latest developed blue-green
 452 infrastructure and urban water management systems are the initiative to further integrate with
 453 microplastics and plastics control, in prior to improving the urban freshwater quality (Table 4).
 454 Nevertheless, Artificial Polymers (including microplastics) are not considered to be
 455 contaminants in the national water quality standards. Recently, an official letter was issued by
 456 the Ministry of Ecology and Environment to suggest setting up a list of toxic and harmful water
 457 pollutants, which includes heavy metal compounds. This letter reflects an official desire to
 458 update existing freshwater quality standards in China; it may be a timely opportunity to include
 459 artificial polymers as indicators or parameters of water quality in standards.

460

461 *Table 4. Legislation and progress for catchment management in China*

| No. | Time | Laws/Policies | Legislative aims | Comments | Citations |
|-----|------|---|--|---|--------------|
| (1) | 1984 | <i>Law of the People's Republic of China on Prevention and control of water pollution (Water Pollution Law)</i> | Establishing a legislative framework of water pollution management | The water pollution law has been updated three times in 1996, 2008 and 2017. It involved the River Chief System in the version of 2017. The 37 th rule prohibits discharging and dumping industrial garbage, urban litters and other types of wastes into Chinese waterbodies. In the 38 th rule, any deposit of solid waste and other type pollutants under the highest water level line of rivers, lakes, channels or reservoirs are forbidden. | (NPC, 2008) |
| (2) | 1988 | <i>Regulations of the People's Republic of China on the administration of river courses (River Courses Regulations)</i> | Refining regulations for freshwater management | Two regulations of its 51 rules (34 th and 35 th) involved freshwater pollution and wastewater management. However, the rest rules are more about land-use issues and flood management. No punishment aiming | (GOSC, 1988) |

| | | | | | |
|-----|------|---|--|---|----------------------------|
| | | | | at polluting freshwater behaviors in this document. | |
| (3) | 1988 | <i>Water law of the People's Republic of China (Water Law)</i> | Establishing a rational legislative basis for developing, using and protecting water resources | The Water Law was updated in 2002 and 2016. In 2002, the concerns of flood, drought, water pollution and soil erosion were added into the new Water Law of 2002. Water Law of 2002 also emphasized the importance of scientific reasoning in management. Water Law of 2016 involves more rules about water pollution and wastewater discharge, where unauthorized wastewater discharge is regarded as illegal behaviors. | (NPC, 2000) |
| (4) | 1988 | <i>Provisions for the development and construction of water and soil conservation in the areas bordering Shanxi, Shaanxi and Inner Mongolia</i> | Solving the confliction between constructions and local water and soil conservation | This policy focuses on the provisional environmental damage led by constructions but pays more attentions to the water and soil conservation instead of pollution problem. This policy was updated once in 2011. | (GOSC and MWR, 2011) |
| (5) | 1988 | <i>Provisions for the administration of water conservation in cities</i> | Managing urban water consumptions | This policy is the first official concern of urban water resource from Chinese central government, where the major concern was about water shortage and water reusing. | (MOHURD, 2015) |
| (6) | 1989 | <i>Provisions for the prevention and control of pollution in water source protection areas for drinking water</i> | Protecting drinking water quality | This policy was updated once in 2010. It sets conservation areas for drinking water sources. It also stipulates that the water quality in the first level conservation area of drinking water, should achieve the Level I of both surface water environmental standards and drinking water quality standards. It forbids the dumping of industrial, urban-based in protected waterbodies. Any leaks of dangerous/risky pollutants into protected waterbodies were prohibited. | (MWR <i>et al.</i> , 2015) |
| (7) | 1989 | <i>Environmental Protection Law of the People's Republic of China (Environmental Protection Law)</i> | Improving ecosystem quality and preventing environmental hazards | Environmental Protection Law mentioned the management of marine environments in the 21 st rule. It also stipulates the pollutants emission towards waterbodies should obey Water Pollution Law. Environment Protection Law was updated in 2014. | (NPC, 1989) |
| (8) | 1991 | <i>Law of the People's Republic of China on soil and water conservation (Soil and Water Conservation Law)</i> | Providing legislative basis for water and soil conservation in China | Soil and Water Conservation Law aims at protection water and soil resources and reducing hydrological hazards and drought events. This law was published in 1991 and updated in 2010. | (NPC, 2013) |

| | | | | | |
|------|------|--|--|--|---------------|
| (9) | 1994 | <i>Urban water supply ordinance of 1994</i> | Managing water supply system in urban area | This policy directly involved urban water supply system but focused on water supply techniques and the water price rather than the water quality in and after the treatment by the system. | (GOSC, 2015a) |
| (10) | 1995 | <i>Interim regulations on prevention and control of water pollution in the Huaihe River Basin</i> | Improving water quality in Huaihe River Basin | This policy was updated during 2010 and 2011. It standardized the industrial emission/discharge towards Huaihe River Basin by 1997 and pushed the river channel cleaning work by 2000. From 1998, any industrial waste emissions towards waterbodies were banned. Some types of industries such as paper mill, electroplate factory, printing and dyeing mill are forbidden nearby the protected waterbodies. | (GOSC, 2015b) |
| (11) | 1997 | <i>Law of the People's Republic of China on flood prevention (Flood Prevention Law)</i> | Establishing a legal basis for flood prevention work and projects in China | Due to China was in a fast but very beginning developing stage at that time, land-uses and flood/drought management is the major hydrological problems existing in China, especially in urban area. Flood Prevention Law as well as Soil and Water Conservation Law were designed for solving above problems. | (NPC, 2006) |
| (12) | 1999 | <i>Measures for the administration of the Pearl River Estuary</i> | Managing water resources in the Pearl River Estuary | This policy was established according to Flood Prevention Law and River Courses Regulations, while it did not involve water pollution considerations. There were also many other similar measures focusing on river- or catchment- scale hydrological environments without involving enough pollution concerns, which implies the first requirement of hydrological environments at that time in China. | (MWR, 2018) |
| (13) | 2000 | <i>Detailed rules for the implementation of the law of the People's Republic of China on prevention and control of water pollution (Detailed Rules of Water Pollution Law)</i> | Providing detailed stipulations for enforcing water pollution law | This policy was established according to Water Pollution Law. It stipulated the vessel shipping in river channel must be awarded pollution prevention certification and instructed with pollution prevention equipment. Port should have relevant equipment to deal with the wastes and ship garbage as well. It also sets four specific objects: 1) Achieving the requirement of environmental functions of waterbodies; 2) Setting staged due time and targets for law implementation; 3) Designing detailed measures for key water pollution prevention areas and pollution sources; | (GOSC, 2005) |

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|------|------|---|---|--|-----------------------|
| | | | | 4) Improving urban planning with the considerations of drainage system in basin area and sewage treatment system. | |
| (14) | 2002 | <i>Environmental quality standards for surface water (GB 3838-2002)</i> | Standardizing the surface water quality assessment | <p>To satisfy the requirement of Environmental Protection Law and Water Pollution Law, the GB 3838 was published in 1988 and updated in 1999. The GB3838-2002 (a new version) took into force from 1st Jun 2002 and replaced previous versions of GB3838.</p> <p>It classified the water quality into five levels according to a list of indicators:</p> <ol style="list-style-type: none"> 1) Head of water and national natural conservation; 2) Level I conservation area of drinking water, distinct aqua ilia habitats, spawning habitat for fish and shrimps; 3) Level II conservation of drinking water source; 4) Normal industrial water and indirectly contacting entertainment water; 5) Agricultural water and landscape water. Artificial polymers (plastics or microplastic content) did not involve in this standard as an indicator. | (MEE and AQSIQ, 2002) |
| (15) | 2002 | <i>Measures for the demonstration and management of water resources for construction projects</i> | Setting a framework of scientific demonstration for relevant construction projects. | This policy was updated once in 2015. Water quality and hydrological environment are regarded as evaluation criterions in the measures. | (MWR and SDPC, 2018) |
| (16) | 2004 | <i>Measures for the supervision and control of sewage discharge outlets into river</i> | Controlling pollutants emission into river | <p>According to Water Law, Flood Prevention Law and River Courses Regulations, this policy stipulated the files and information required for applying construction or reconstruction of sewage draining exit towards river.</p> <p>It is a concern of freshwater quality from another aspect. However, it did not involve a detailed demand of water quality.</p> | (MWR, 2015a) |
| (17) | 2004 | <i>Measures for the administration of the Yellow River estuary</i> | Managing the water environment of the Yellow River estuary | It is similar with the ' <i>Measures for the administration of the Pearl River Estuary</i> ' and did not involve water pollution consideration. | (MWR, 2015b) |
| (18) | 2005 | <i>Standards for irrigation water quality (GB 5084-2005)</i> | Refining the water quality standards for irrigation water | This standard was established basing on GB3838 and replaced GB 5084-1992. As the same as GB3838, GB5084 did not considered artificial polymer content as a factor of water quality. | (AQSIQ, 2005) |

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|------|------|--|---|--|-----------------------|
| (19) | 2006 | <i>Standards for drinking water quality (GB 5749-2006)</i> | Refining the water quality standards for drinking water | GB5749 also did not involve artificial polymer occurrences as a factor of water quality. | (MOH and AQSIQ, 2006) |
| (20) | 2007 | <i>Regulations of the People's Republic of China on hydrology (Hydrology Regulations)</i> | Standardizing the hydrological management and relevant projects | It was established according to the Water Law and Flood Prevention Law, and updated three times in 2013, 2016 and 2017. It provides a framework for the hydrological planning, construction, forecast, monitoring, data management, instruction protection and legal responsibilities. | (GOSC, 2017c) |
| (21) | 2008 | <i>Plans for the route to the sea for the operation of the three gorges reservoir and measures for the management of water resources and river courses in the reservoir area</i> | Managing the channels of Three Gorges Basin | It forbids dumping wastes and discharging sewage into the reservoir areas according to the Water Law and Flood Prevention Law. | (MWR, 2018) |
| (22) | 2009 | <i>Measures for the administration of the estuary of the New Yongding River, the Haihe River and the Duliujian River</i> | Channel management for estuary of three rivers | With Water Law, Flood Prevention Law and River Courses Regulations as the legislative basis, the waste dumping and disposal are not allowed in those managed catchment areas. | (MWR, 2015c) |
| (23) | 2011 | <i>Measures for the protection of hydrological monitoring environment and facilities</i> | Protecting hydrological monitoring area and facilities | It was updated once in 2015 and banned dumping wastes near to the monitoring area of hydrological conditions. | (MWR, 2015d) |
| (24) | 2011 | <i>Regulations on the administration of the Taihu Lake Basin (Taihu Regulations)</i> | Protect hydrological environment of the Taihu Lake Basin | Taihu Regulations make specific rules for the management of Taihu Lake catchment according to Water Law and Water Pollution Law. The dumping of any types wastes into the Taihu Lake is banned. | (GOSC, 2015c) |
| (25) | 2015 | <i>Action plan for prevention and control of water pollution (Water Pollution Action)</i> | Enhancing prevention of water pollution and strengthening national water security | Water Pollution Action, as known as 'Ten Measures of Water' in China, issued ten demands on water pollution prevention works nationwide: (1) Completely controlling pollutants discharge; (2) Transforming and upgrading the economic structures; (3) Conserving and protecting water resources; (4) Strengthening science and technology support; (5) Giving full play to the role of market mechanisms; (6) Intensifying environmental law enforcement and supervision; (7) Reinforcing management on water environments; | (MEE, 2016) |

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|------|------|--|--|---|-------------------------------|
| | | | | <p>(8) Ensuring ecological safety of water environments;</p> <p>(9) Clarifying and implementing the responsibilities of all parts in governments and organizations;</p> <p>(10) Strengthening public participation and oversight.</p> <p>It also requires that the water quality of over 70% of total water in the major seven basins (Yangtze River, Yellow River, Pearl River, Songhua River, Huaihe River, Haihe River and Liaohe River) reach Level III according to GB3838-2002 by 2020.</p> | |
| (26) | 2016 | <i>Views on the full implementation of the River Chief System</i> | Popularizing the River Chief System scheme to national level | <p>It stipulates to establish provincial, city, county and township level in the River Chief Systems by the end of 2018 nationwide in China.</p> <p>By June 2018, 31 provinces have established the River/Lake Chief System (RCS) and 300,000 river chiefs have been selected for the four levels. By the end of 2018, Beijing, Tianjin, Jiangsu, Zhejiang, Anhui, Fujian, Jiangxi and Hainan had established RCS completely.</p> | (GOSC, 2016) |
| (27) | 2017 | <i>Standards for groundwater quality (GB/T 14848-2017)</i> | Refining the water quality standards for groundwater | It replaced GB/T 14848-1993 and added another 54 indicators for water quality evaluations but did not include any considerations of artificial polymers. | (AQSIQ, 2017) |
| (28) | 2018 | <i>Work Specification of the River Chief System (DB3306/T 015-2018); Work Specification of the Lake Chief System (DB3306/T 016-2018)</i> | Establishing the responsibility framework of RCS/LCS. | They are the first formal legislations for River Chief System scheme even if they were issued by local governments. The popularization of such work specification could be expected. | (Shaoxing City, 2018a, 2018b) |
| (29) | 2018 | <i>Implementation plan for urban black and smelly water treatment</i> | Completing the installation of and improving sewage treatment systems in cities and towns | <p>Basing on Water Pollution Action, this policy aims at eliminating 90% black and smelly waterbodies (BSW) in urban built-up areas for major cities in China by the end of 2018, for prefecture cities by the end of 2020.</p> <p>RCS/LCS is required to play an important role in this implement.</p> | (MOHURD, 2018) |
| (30) | 2019 | <i>Management of urban pipe networks and sewage treatment subsidy funds</i> | Standardizing the grants of subsidy funds for urban pipe networks and sewage treatment from 2018 to 2021 | This subsidy funds will cover Sponge City project, underground pipe network establishment project, urban BSW treatment pilot project, and improvement of sewage systems in the cities and towns of central and western regions, with different specific plans. | (MOHURD and MOF, 2019) |
| (31) | 2019 | <i>The notice of the three ministries on organizing the application of pilot</i> | Standardizing the requirements of applying pilot cities | Basing on the official document 'Implement plan for urban black and smelly water treatment', this | (MOF, 2019) |

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|------|------|--|---|--|----------------|
| | | <i>cities for the treatment of black and smelly waterbodies in urban areas in 2019</i> | for treatment of black and smelly waterbodies | document clarifies the application approaches to becoming treatment urban BSW pilot cities. Sponge City theory and RCS/LCS was encouraged to combine with BSW treatment in this document. | |
| (32) | 2019 | <i>Three-year action plan for improving the quality and efficiency of urban sewage treatment (2019-2021)</i> | Completing the installation of and improving sewage treatment systems in cities and towns | This policy confirms the major objects on the improving sewage treatment systems in cities and towns, where Biochemical Oxygen Demand (BOD) was precisely issued as the indicator representing the quality of sewage treatment systems. | (MOHURD, 2019) |
| (33) | 2019 | <i>Letter soliciting opinions on 'the list of toxic and harmful water pollutants (Batch I) (exposure draft)'</i> | Suggesting establishing a professional list to clarify the types of water pollutants and collecting comments from relevant professional organization in China | By February 2019, nine pollutants were added into the list: dichloromethane, trichloroethane, trichloro ethylene, tetrachloroethylene, cadmium and cadmium compounds, mercury and mercury compounds, hexavalent chromium compounds, lead compounds, and arsenic and arsenic compounds. | (MEE, 2019) |

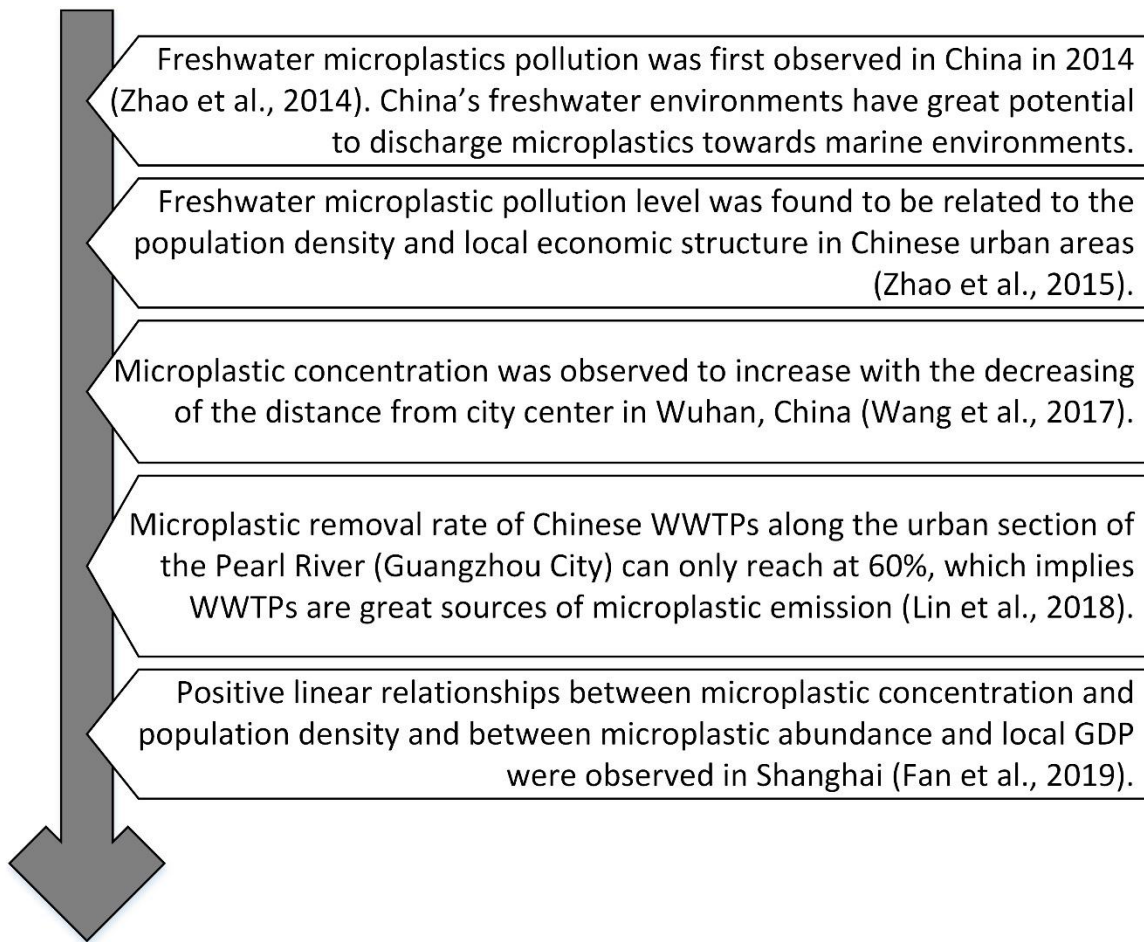
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464 **4. Mitigation of microplastic pollutions in Chinese urban freshwater environment**

465 The growing findings on microplastic pollution, especially in urban catchment (Figure 4), are
 466 likely to motivate legislative action in China (Yuan *et al.*, 2019). However, there are major
 467 challenges to reducing plastic pollution in China, stemming from the many benefits of and
 468 society's reliance on plastics, and a lack of suitable alternatives for some applications. These
 469 challenges are summarised in Figure 5. Reflecting both current knowledge gaps and these
 470 challenges, we put forward a number of recommendations, as detailed in the remainder of this
 471 section.

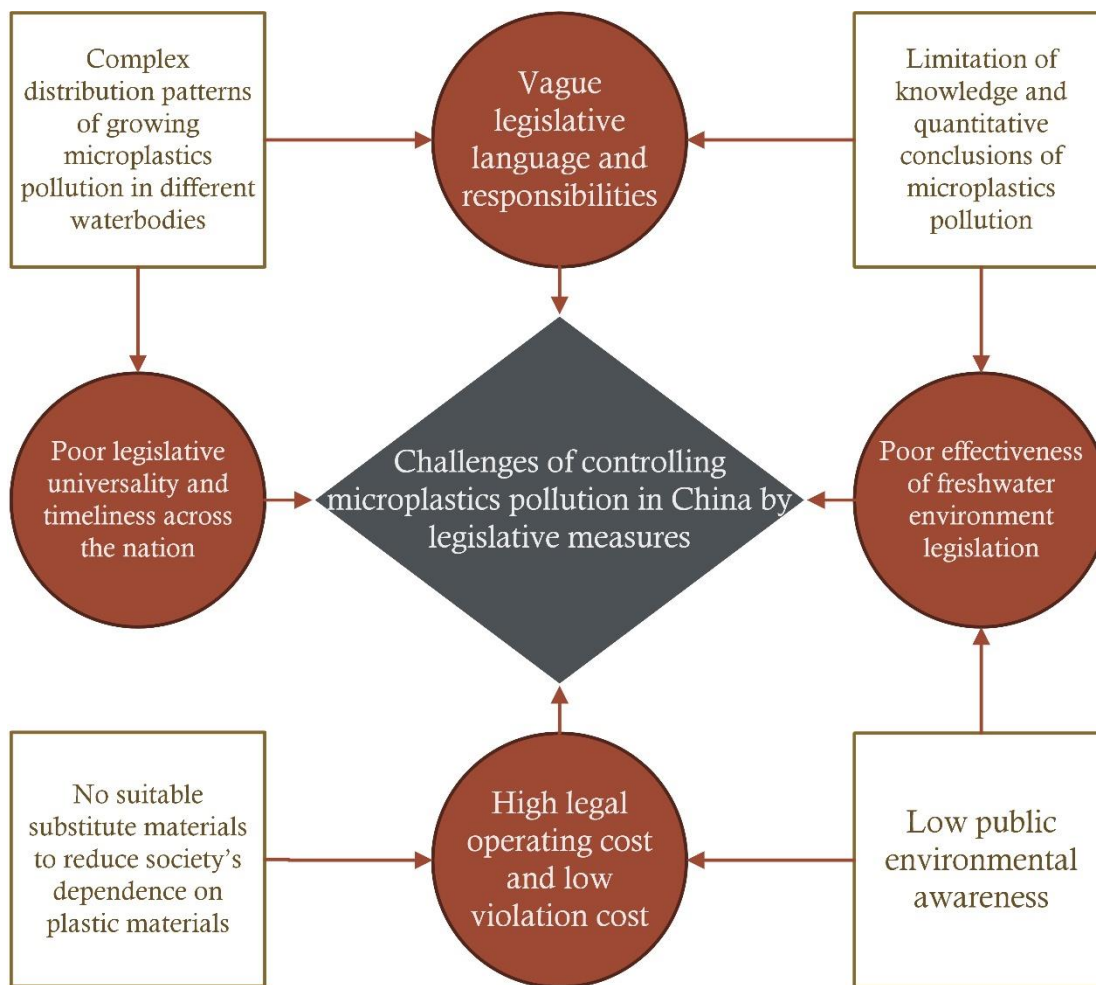
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473

474 *Figure 4. Key research progresses on microplastic pollution in Chinese urban catchments (Source: Yuyao Xu)*

475



476

477 *Figure 5. Reasons of legislative difficulties on microplastic pollution (Source: Yuyao Xu)*

478

479 **4.1. Toxicology.**

480 China needs to build a toxicological (dose-based) or environmental (impact-based) microplastic
 481 pollution risk assessment system, and formulate reasonable treatment plans based on this
 482 quantitative basis. Although microplastics are categorised as toxic pollutants and banned in
 483 Canada (Canada, 2015; Canada.ca, 2017), there are some uncertainties remaining over direct
 484 toxicity to humans and wildlife. A common problem is that many toxicity tests have been based
 485 on doses much higher than found in the environment, so assessing the risks to human and

486 ecological health from current levels of contamination remains problematic (X. Jiang *et al.*,
487 2019). Advanced toxicological studies may either provide the impetus for the Chinese
488 government to list microplastics as toxic pollutants or allay concerns about current levels. Risk
489 assessment systems should also consider the interactions between microplastics and other
490 relevant pollutants (persistent organic matters, microorganisms and heavy metal).

491 ***4.2. Recognising different types of microplastics***

492 China should apply specific management measures for different types of microplastics.
493 Microplastic is actually a general term, encompassing plastic debris with a wide variety of
494 characteristics. These characteristics (e.g. sizes, shapes, chemical composition) affect their
495 distribution patterns and environmental impacts, and require different solutions to manage each
496 type of them. ‘*Microbead-free*’ action is a successful case for the management of a single group
497 of microplastics, but micro-fibres are far more abundant in many freshwaters, especially in
498 urban catchments (see Table 1). Domestic household discharge (e.g. via washing clothes on
499 urban rivers) is one of the main sources of fibres in freshwater environments (see Figure 3).
500 Yang *et al.* (2019) found that polyester fabric releases fewer fibres during laundry compared to
501 polyamide fabric and acetate fabric and so developing textiles that do not as readily shed fibres
502 would be beneficial. Also, improving laundry and fabric filter techniques of washing machines
503 may reduce the amount of fibres from washing machines into sewage pipes in residential areas,
504 is a way to cut the transportation of microplastics. Compared to platen laundry machines, fibres
505 are easier to peel off in pulsator laundry machine (Yang *et al.*, 2019). Another way to cut the
506 transportation of synthetic fibres is to improve the microplastic removal rate of WWTPs in
507 China. Legislation aiming at those three points can effectively control the amount of synthetic

508 fibre emitted in densely populated areas. This approach to controlling fibre pollution could also
509 be extended to other types of microplastics.

510 ***4.3. Consumption rates***

511 We recommend reducing the consumption of plastic products, such as plastic bags and single-
512 use food containers. This will be a considerable challenge as the amount of plastic consumed
513 from these sources is increasing every year in China. By 2017, over 20 million fast-food
514 deliveries were produced per day, where daily plastic bag consumption was enough to cover
515 168 football fields in China (Xue, 2017). In early 2011, the Chinese Government had to spend
516 about 18.5 million yuan per year to control macro-plastic pollution (Zhu, 2011). Promoting the
517 use of recyclable packaging throughout China's e-commerce industry, or supplying recycling
518 services for non-disposable food containers in food delivery businesses, are therefore measures
519 with great potential within China (Hao, 2019). This alone will not be enough and such
520 reductions should be coupled with investment and development of adequate recycling
521 infrastructure. Developing alternative materials for plastics or controllable plastics degradation
522 technology is also an approach to reducing the society's dependence on plastic products.

523 ***4.4. Management strategy***

524 The fourth recommendation is the devolution of management responsibility for microplastic
525 pollution to local governments after central government sets general targets and overarching
526 legislation. As evident in Table 1, microplastic loads vary markedly across different
527 waterbodies and, therefore, local governments are better placed to specify management
528 strategies given their detailed local knowledge and prioritisation of local environmental threats
529 and constraints. This may then inform national action, as was employed in the USA for the
530 "*Microbeads-free Waters Act*" (McDevitt *et al.*, 2017), and a similar approach has been used to

531 develop the “*Sponge City Program*” in China (Chan *et al.*, 2018). As part of a better
532 management strategy, it is useful to consider microplastics and water quality more broadly, in
533 longer-term land-use improvement projects in China. By combining the microplastic risk
534 assessment system and the approach that manages different types of microplastics separately,
535 the abundance of some groups of microplastics (such as PP or PE in particular size range) could
536 be listed as a water quality index of the national freshwater quality standard (e.g. based on the
537 *GB3838-2002 that published by the National Environmental Bureau from the Chinese*
538 *Government*). The updated water quality standards will have a long-term impact on China’s
539 future hydrologic and environmental management, accomplished with current freshwater
540 management strategies such as the “*River/Lake Chief System*” and “*Sponge City Project*”.

541 Complex land-use patterns and a multitude of industrial and urban activities create
542 substantial challenge for employing a unified approach to freshwater management in China. As
543 microplastic pollution is related closely to land-use function and local economic structure in
544 Chinese cities, the country’s urban planning system is well placed as to help integrate
545 considerations of land-use demand and microplastic pollution, as part of strategies to promote
546 more sustainable development.

547

548 **5. Conclusion**

549 In this paper, we have reviewed microplastics in Chinese urban freshwater environments and
550 relevant legislation that aims at managing microplastic pollution. Microplastic properties,
551 hydrological, meteorological and geographical conditions, population size and local land-use
552 functions are critical factors determining microplastic concentrations in urban waterbodies in
553 China. With the growing loads of microplastics entering lakes, streams and rivers, improving

554 management strategies and developing legislation is a significant challenge in China but one that
555 needs to be addressed. Unfortunately, fundamental knowledge of loads, transport pathways and
556 mechanisms, and of the toxicological effects of microplastic, remains limited in China, so there
557 remains a need for more empirical research to help underpin evidence-based legislation.

558 The Chinese government has paid more attention to urban water quality over the last
559 three years, but as yet there is no legislation that deals explicitly with microplastics. This is an
560 important issue, given that China is known to release large quantities of microplastic particles
561 into its freshwaters. The potential ecological and human health risks posed by microplastics,
562 speak to the need for improving legislation and policy frameworks, to better manage current
563 and deal with future threats of plastic waste.

564

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970 **Appendix: Abbreviation of organization names**

971 AQSIQ: Administration of Quality Supervision Inspection and Quarantine (中华人民共和国
972 国家质量监督检验检疫总局)

973 GOSC: General Office of the State Council of the People's Republic of China (中华人民共和国
974 国务院办公厅)

975 IMO: International Maritime Organization

976 MEE: Ministry of Ecology and Environment of the People's Republic of China (中华人民共和国
977 生态环境部)

978 MLR: Ministry of Land and Resources of the People's Republic of China (中华人民共和国
979 国土资源部)

980 MOF: Ministry of Finance of the People's Republic of China (中华人民共和国财政部)

981 MOH: Ministry of Health of the People's Republic of China (中华人民共和国卫生部)

982 MOHURD: Ministry of Housing and Urban-Rural Development of the People's Republic of
983 China (中华人民共和国住房和城乡建设部)

984 MWR: Ministry of Water Resources of the People's Republic of China (中华人民共和国水利
985 部)

986 NDRC: National Development and Reform Commission of the People's Republic of China (中
987 华人民共和国国家发展和改革委员会)

988 NPC: The National People's Congress of the People's Republic of China (中华人民共和国全
989 国人民代表大会)

990 UNEA: United Nations Environment Assembly

991 WHO: World Health Organization

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