

1 **Elemental composition of yerba mate (*Ilex paraguariensis* A.St.-Hil.) under low input**
2 **systems of southern Brazil**

3
4 **Highlights**

- 5 • Portions of Fe, Ti, As, Mo, Li, V, and Pb found in yerba mate originate from dust.
- 6 • 23% of yerba mate leaves exhibited Cd levels close to or above MAL (0.4 mg kg⁻¹).
- 7 • Soil parent material impacted yerba mate foliar elemental composition.
- 8 • Yerba mate exhibited a great capacity to extract and accumulate Ca and Mn.

9
10 **ABSTRACT**

11 Elemental composition of food can be used to determine nutritional potential as well as
12 guiding legislation for establishing maximum acceptable limits (MAL) of metals in
13 consumption products. This study aimed to determine the elemental background levels of
14 yerba mate (*Ilex paraguariensis* A.St.-Hil.) under varied geologic formations in southern
15 Brazil. Mature leaves were randomly collected from four wild-grown plants at thirty native
16 sites in three states and analyzed for 32 elements. Since yerba mate is not washed to obtain
17 the final product, leaves were analyzed with and without washing to assess foliar deposition.
18 Concentration values of As, Ag, Be, Cs, Cr, Li, Se, Tl, U, and V were near detection limits,
19 indicating low potential as a source and/or toxicity to the consumer. Washing decreased
20 concentrations of Fe, Ti, As, Mo, Li, V, and Pb, suggesting atmospheric contributions/dust
21 deposition. Concentrations of Mn (very high), Zn (high), and Ni (high) demonstrated that
22 leaves could be an important source of these elements. Soil parent material affected elemental
23 composition with basalt providing higher concentrations of Mn, P, and Co while Rhyodacite
24 provided higher concentrations of K and Na. All samples exhibited Pb values below the MAL
25 of 0.6 mg kg⁻¹, but 23% of washed leaves and 20% of unwashed leaves had Cd

26 concentrations close to or above the MAL value of 0.4 mg kg⁻¹. Study results indicated that
27 Cd MAL values for yerba mate in southern Brazil should be reassessed.

28 **Keywords:** Toxic elements, Heavy metals, ICP-MS, Food security, Legislation.

29

30 **1. Introduction**

31 Yerba mate (*Ilex paraguariensis* A.St.-Hil.) is a plant native to southern Brazil, Paraguay
32 and northeastern Argentina. Distribution of this species primarily coincides with the
33 subtropical Araucaria forest [*Araucaria angustifolia* (Bertol.) Kuntze] encompassing
34 approximately 540,000 km². During pre-Columbian times, many native populations used the
35 leaves and fine branches as an infusion product (Cardozo-Junior and Morand, 2016).

36 Several cycles of yerba mate exploitation have occurred in the five centuries since
37 European colonization of the region. Yerba mate has played key economic and social roles in
38 many remote communities in Brazil where more than 180 thousand farmers are estimated to
39 be cultivating this species, especially on small farms (Marques, 2014). Also, yerba mate is
40 part of the daily diet for a large portion of populations in Uruguay, Argentina, Paraguay, and
41 Brazil (Marques, 2014). Recently, pharmaceutical, cosmetic, and beverage industries have
42 increased research on potential new products derived from yerba mate (Cardozo-Junior and
43 Morand, 2016).

44 Yerba mate is not cultivated on a commercial scale elsewhere. Thus, local production has
45 been driven by both internal consumption and exportation. There are two very distinct yerba
46 mate production systems. The first is an agroforestry system (low-input) with spontaneous
47 yerba mate growth under the Araucaria forest canopy without fertilization and liming. The
48 second is monocultures of planted seedlings without forest protection from full sun exposure.
49 However, low-input agroforestry systems have been historically preferred by process

50 industries since they are more lucrative compared to yerba mate monocultures under full sun
51 exposure (Marques, 2014).

52 Remaining native Araucaria forests can no longer be logged, and yerba mate harvesting
53 and Araucaria seed gathering are the primary uses of these native Araucaria forests (Danner
54 et al., 2012). In the past, different degrees of thinning have been applied to Araucaria forests,
55 especially from logging. These thinning's enabled sunlight to pass through the canopy,
56 allowing the growth of grass for raising cattle, sheep, pigs, and horses (Marques, 2014). By
57 maintaining yerba mate during forest thinning or by natural regeneration, a combination of
58 forest, yerba mate, and pasture has been established and preserved by farmers for generations
59 (Marques, 2014; Pinotti et al., 2018).

60 Yerba mate is often found on acidic soils under native conditions (pH 3 to 4), and farmers
61 choose acidic soils to plant commercial orchards with low lime application rates (Ca and Mg
62 supply) or with no applied lime (Reissmann et al., 1999). Adaption to acidic soil conditions
63 has been confirmed in a recent study that indicated increased seedling root growth when
64 applying Al to hydroponic solutions (Benedetti, 2012). In this regard, yerba mate displays
65 characteristics similar to tea (*Camellia sinensis* L.) (Fung et al., 2008; Bora et al., 2019).

66 As result of low soil pH, yerba mate leaves may have Al concentration higher than 1000
67 mg kg⁻¹ (Reissmann et al., 1999) and Mn concentrations higher than 3000 mg kg⁻¹ (Fossati,
68 1997; Jacques et al., 2007). However, these concentrations can vary from site to site as well
69 as within a site (Fossati, 1997). Additionally, soil acidity cannot be the only factor impacting
70 Al uptake by yerba mate since soil exchangeable Al was not directly correlated to Al
71 concentrations found in plants (Fossati, 1997; Reissmann et al., 1999). Researchers have
72 suggested that plant root exudation influences nutrient and toxic element availability in yerba
73 mate (Reissmann et al., 1999; Benedetti, 2012).

74 Yerba mate samples from marketplaces have been used to study differences in mineral
75 composition between countries (Pozebon et al., 2015; Baran et al., 2018), states (Saidelles et
76 al., 2010; Santos et al., 2017), and municipalities within the same state (Costa et al., 2009;
77 Tatsch et al., 2010; Bragança et al., 2011). However, yerba mate may be transported long
78 distances after drying, and there is no guarantee that processing facilities represent the
79 location of origin. Although the influence of washing tea leaves on metal concentrations has
80 been evaluated (Tsushida and Takeo, 1977; Jin et al., 2005; Han et al., 2006), there is a lack
81 of research on this topic for yerba mate. This aspect of the yerba mate production chain
82 should be investigated since the washing step is omitted during industrial processing
83 (Valduga et al., 2019a).

84 Recently, highly restrictive legislation for heavy metal allowance in yerba mate was
85 imposed by the Common Market of the South (MERCOSUL), which established
86 concentration limits of 0.60 mg kg⁻¹ and 0.40 mg kg⁻¹ for Pb and Cd, respectively (ANVISA,
87 2013). This legislation change has raised many concerns regarding metal contamination.
88 Some investigations have detected Pb and Cd in yerba mate above these limits (Santos et al.,
89 2017; Valduga et al., 2019b). However, no studies have evaluated naturally-occurring toxic
90 element backgrounds under native conditions. Therefore, this study was designed to
91 determine background values of nutrients and trace elements in yerba mate leaves cultivated
92 under native or low anthropic intervention systems. In addition, changes in these values due
93 to leaf washing was investigated.

94

95 **2. Material and methods**

96 *2.1. Study area*

97 Thirty sites in southern Brazil were selected for sampling yerba mate leaves and soils (Figure
98 1). Three states in the region (*Paraná* – PR, *Santa Catarina* – SC, and *Rio Grande do Sul* –

99 RS) have native occurrence of yerba mate and account for more than 99% of commercial
100 yerba mate production in Brazil (IBGE, 2018). Climatic conditions based on Köppen
101 classification were Cfa (Subtropical Humid, without dry season with hot summers) or Cfb
102 (Subtropical Humid without dry season with temperate summers), with an average altitude
103 above 650 m for the region (Alvares et al., 2013). All sites had Mixed Ombrophylous Forest
104 (Araucaria Forest) as native vegetation.

105 Four criteria were sought when selecting the 30 sites: 1) close to native conditions with
106 no history of chemical fertilizer, lime application, and herbicide use (note that *Barão de*
107 *Cotegipe* site was an organic farm system that used pig manure); 2) represent major geologic
108 formations where yerba mate occurs naturally; 3) within major areas of commercial
109 production; and 4) large variation in plant genotype (i.e., no established cultivar, no cloning
110 or plant selection). Table S1 shows compiled information about sites and management
111 conditions.

112

113 2.2. Leaf and soil sampling

114 Samples were collected between 03 July and 18 September 2015. This period represents a
115 state in yerba mate growth where all leaf tissues are mature. At the *Itaiópolis* and *Lebon*
116 *Régis* sites, plants had not been previously harvested, were greater than 15 m in height, and
117 were part of a preserved forest. The *Lebon Régis* forest differed from others sites in the
118 amount of ground residue from the natural senescence of “*taquara*” (*Merostachys*
119 *multiramea* Hackel). As a member of the Poaceae family, “*taquara*” can dominate the
120 understory canopy by creating dense cane thickets which naturally died 28 years after
121 blooming.

122 At each site, four plants were randomly selected and 12 branches were cut from each
123 plant from all sides and heights (low, medium, and top) with manual tree pruning shears. The

124 cut branches were placed in paper bags for transport. At the laboratory, leaves were manually
125 removed from branches, mixed, and divided into two portions for each site. One half was
126 placed in labelled paper bags and the other half was placed on a plastic tray. The tray samples
127 were washed twice with tap water and once with purified water by emersion before
128 placement in another labelled paper bag.

129 Three soil samples were collected from a meter around from each yerba mate plant and
130 mixed to make a composite soil sample from these 12 single site samples. For sampling, a
131 manual soil auger collected samples to a depth of 0–20 cm.

132

133 *2.3. Leaf analysis*

134 Two paper bags representing washed and unwashed samples, were placed in a forced air
135 oven at 65°C until constant weight. After drying, samples were ground with a Willey mill to
136 pass a 1 mm sieve and placed in plastic containers.

137 Leaf tissue was analyzed for Ag, Al, As, B, Ba, Be, Ca, Cd, Co, Cr, Cs, Cu, Fe, K, Li,
138 Mg, Mn, Mo, Na, Ni, P, Pb, Rb, S, Se, Sr, Ti, Tl, U, V, and Zn. An Ethos 1 Microwave
139 Digester (Milestone SrL, Bergamo, Italy) platform with a capacity to hold 48 vessels was
140 utilized. Multi-elemental analysis was performed using Inductively Coupled Plasma Mass
141 Spectrometry (ICP-MS; Agilent 7500ce, Santa Clara, CA, USA).

142 Each 48 sample run consisted of 42 ground leaf samples (200 mg), 2 internal standards, 2
143 tomato leaf standards (NIST Tomato 1573A, National Institute of Standards and Technology,
144 NIST, Gaithersburg, MD, USA), and 2 reagent blanks. Samples were weighed into vessels
145 before adding 2 mL HNO₃ (re-distilled from 69% ‘Aristar’ grade; VWR Ltd, Lutterworth,
146 UK), 1 mL of 18.2 MΩ cm water, and 1 mL of 30% (100 volumes) hydrogen peroxide (Trace
147 Select Ultra, Fluka, Gillingham, UK). Operational sample blanks were run to determine limit

148 of detection values (LOD; 3*standard deviation, SD) in digested composite samples.
149 Percentage of recovery LOD values are shown in Table S2.

150 The microwave heating programme was: 10 min at 500 W to 55° C; 3 min at 100 W to
151 65° C; 14 minutes at 1200 W to 115° C; 20 minutes at 1200 W to 140° C, 35 bar; and cooling
152 for 15 minutes. Under atmospheric pressure, samples were digested in vessels comprised of
153 polytetrafluoroethylene (PTFE) and polyethylethylketone (PEEK) pressure jackets with a
154 safety devices that activate at a pressure of 35 bar. After cooling, contents of each vessel were
155 washed into a tared 15 mL plastic tube (Greiner Bio-One, Stonehouse, UK) for storage. After
156 all samples were digested, 1:5 dilutions were created by transferring 2 mL of sample solution
157 to a bottle and adding 8 mL of Milli-Q H₂O.

158

159 *2.4. Soil analysis*

160 Samples were air dried and passed through a 2 mm sieve. Soil chemical analysis
161 consisted of: pH-CaCl₂ 0.01 M (1:2.5 – soil:solution ratio), soil buffer to pH 7 (H + Al),
162 exchangeable Al, Ca, and Mg (KCl 1 M – extraction), available P, K, Fe, Mn, Zn, and Cu
163 (Mehlich I – extraction), and organic C according to procedures of Marques and Motta
164 (2002). The following instrumentation was used to perform analysis: atomic absorption
165 spectrophotometer for Ca, Mg, Fe, Mn, Zn, and Cu; flame spectrophotometer for K; and
166 UV/Vis spectrophotometer for P. Soil granulometric analysis was performed for clay, silt,
167 and coarse and fine sand, according to Donagema et al. (2011) using pipette methodology.
168 The sum of soil bases (BS = Ca + Mg + K), effective cation exchange capacity (CEC)
169 (CEC_{ef} = SB + Al), pH 7.0 CEC (CEC pH 7.0 = BS + (H + Al)), percentage of base
170 saturation ($V = BS/CEC \text{ pH } 7.0 * 100$), percentage of Al saturation [$m = Al/(BS + Al) * 100$]
171 were calculated. Soil chemical and physical properties are shown in Table 1. Soil color was
172 determined from 2 mm dried sieved soil samples using a Munsell color chart.

173 Total soil chemical composition was determined by digesting a 0.2 g soil sample with
174 HNO₃ and HClO₄, followed by HF, in duplicate. The final solution was solubilised with
175 HNO₃ and analyzed by inductively coupled plasma mass spectrometry (ICP-MS; Agilent
176 7500ce, Santa Clara, CA, USA). Montana soil was used in each batch and percentages of
177 recovery were: Na – 89.7; Mg – 77.4; P – 101.3; K – 93.6; Ca – 89.2; Al – 77.4; V – 97.0; Cr
178 – 78.2; Mn – 91.1; Fe – 94.2; Co – 91.5; Ni – 90.2; Cu – 88.7; Zn – 111.6; As – 107.3; Se –
179 100.0; Rb – 92.9; Sr – 89.5; Mo – 98.8; Ag – 91.2; Cd – 103.7; Cs – 88.9; Ba – 95.0; Ti –
180 101.1; and U – 91.8.

181

182 2.5. Data analysis

183 Student t-tests between washed and unwashed samples were performed for each element.
184 Correlation among elements was conducted for washed and unwashed samples. A Principal
185 Component Analysis (PCA) with a correlation matrix of total soil elements (data transformed
186 by box-cox and standardized) was performed. Also, a PCA of leaf elemental composition was
187 performed to identify which elements primarily explain data variance (data not shown). For
188 selected elements (Mn, P, Co, Ni, Mg, Al, Fe, K, Na, Ca, and Ba), Discriminant Linear
189 Analysis (LDA) was used to discriminate samples by region and soil parent material.
190 Elements used in LDA were compared according to soil parent material by the Kruskal
191 Wallis test, followed by Dunn's test (when $p < 0.05$). Statistical analyses were performed in
192 software R, version 3.6.0 (R Core Team, 2019).

193

194 3. Results and Discussion

195 3.1. Elemental composition

196 Levels of many elements were close to or below detection limits (Table S2). Among
197 analysed elements (Table 2), Ag and U were below detection limits in both unwashed and

198 washed samples. For As, Co, Cr, Mo, Na, Ni, Pb, Se, Tl, and V, values were close to or
199 below detection limits. Note that the observed low value of Se is similar to findings of
200 Marcos et al. (1998) and Nookabkaew et al. (2006) for Camellia tea, which is another
201 common infusion product. Although Se is essential for animal nutrition, both yerba mate and
202 Camellia tea are low potential Se sources for human consumption.

203 The established allowance limit for As is 0.6 mg kg⁻¹ (ANVISA, 2013), but there is no
204 established allowance limit for Cr in yerba mate in South America. Camellia tea is another
205 common infusion product; by comparison, Chinese legislation sets allowance limits for Cr at
206 5.0 mg kg⁻¹ and As at 2.0 mg kg⁻¹ (Bugang and Woolsey, 2010). Our observed values (Table
207 2) for both of these elements were below allowance values. Pozebon et al. (2015) reported
208 maximum values of 0.70 (Cr) and 0.06 (As) mg kg⁻¹ for 54 processed commercial yerba mate
209 samples from four South American countries. Low values of Cr from native samples
210 observed in our study (Table 2), as well as commercial yerba mate values reported by
211 Pozebon et al. (2015), indicate no contamination from soil or from yerba mate commercial
212 processing. In general, observed Cr and As values for yerba mate leaves (Table 2) and
213 commercial yerba mate samples (Pozebon et al., 2015) were lower than for tea (Zheng et al.,
214 2014; Moreda-Piñeiro et al., 2003; Zhong et al., 2015; Seenivasan et al., 2008; Marcos et al.,
215 1998; Brzezicha-Cirocka et al., 2016; Han et al., 2005).

216 Lead average levels for yerba mate leaves (Table 2) were lower than the Pb allowance
217 concentration of 0.6 mg kg⁻¹ adopted by Mercosul for yerba mate (ANVISA, 2013) and far
218 below the 5.0 mg kg⁻¹ level for tea from China (Bugang and Woolsey, 2010). However, some
219 of our yerba mate samples had concentrations above 0.6 mg kg⁻¹. In other words, yerba mate
220 can have natural leaf Pb concentrations above the limit established by Mercosul legislation.
221 Corroborating results of our study, values below the allowance were observed by others for
222 yerba mate commercial samples (Pozebon et al., 2015; Saidelles et al., 2010; Heinrichs and

223 Malavolta 2001). These yerba mate observations suggest that variation in soil parent material,
224 management practices, and commercial processing has low impact on Pb levels.

225 Many samples had Cd values above the Mercosul allowance level of 0.4 mg kg⁻¹ (Table
226 2). Five sites had concentrations above that allowed by legislation and two other sites were
227 near the limit (Table S3). Since 29 sites did not receive fertilizer or other manufactured
228 agricultural products that could account for Cd contributions, these observed values represent
229 background concentrations under natural conditions. The *Barão do Cotegipe* site that
230 received lime and pig manure had values of 0.34 mg kg⁻¹ (Table S3). Given this noted
231 background level, the 0.4 mg kg⁻¹ limit adopted by Mercosul is a very restrictive value since
232 natural concentrations in yerba mate leaves are often very close to or higher than this value. If
233 the adopted limit was 1.0 mg kg⁻¹ for yerba mate, all samples sites would be below this limit
234 since the highest value of washed samples was 0.475 mg kg⁻¹ (Table 2). Overall, the general
235 mean value of Cd (0.27 mg kg⁻¹) was below those for Brazilian, Uruguayan, and Argentinian
236 commercial yerba mate reported by Pozebon et al. (2015) and were within the 0.35 to 1.21
237 mg kg⁻¹ commercial yerba mate range observed by Saidelles et al. (2010) for three southern
238 Brazilians states used in our study. If the established MRV was closer to natural background,
239 23% of washed leaves and 20% unwashed leaves would still exceed the standard.

240 Commercial yerba mate products and leaves used in our study generally had higher Cd
241 concentrations than those reported for tea (Han et al., 2005; Marcos et al., 1998; Tsushida and
242 Taeko, 1997; Seenivasan et al., 2008; Zhong et al., 2015; Brzezicha-Cirocka et al., 2016).
243 Only Zheng et al. (2014) reported Cd in Camellia tea above our average value. On the other
244 hand, Pb concentrations in our study were generally lower than for tea (Zheng et al., 2014;
245 Moreda-Piñeiro et al., 2003; Zhong et al., 2015; Seenivasan et al., 2008; Marcos et al., 1998;
246 Han et al., 2006; Tsushida and Taeko, 1977; Brzezicha-Cirocka et al., 2016, Chen et al.,
247 2010). Thus, soil, climate, or specific plant characteristic could influence differences in toxic

248 metal concentrations for tea and yerba mate. It is important to highlight that observed natural
249 concentrations of Pb and Cd in yerba mate leaves attained levels that could compromise the
250 production chain under current legislation.

251 We observed means and ranges for Rb (46.4; 14.1 to 132.3 mg kg⁻¹) and Sr (44.1; 15.0 to
252 88.9 mg kg⁻¹) that were higher than those reported by Pozebon et al. (2015), but similar to
253 their Ba values (64.4; 32.5 to 97.5 mg kg⁻¹). Marcos et al. (1998) observed similar values for
254 Rb (58.8; 23.1 to 114.6 mg kg⁻¹) and smaller values for Sr (18.5; 5.3 to 43.3 mg kg⁻¹) and Ba
255 (27.5; 5.0 to 57.7 mg kg⁻¹) in tea. Despite differences, observed values may have little effect
256 on infusion quality. There are no allowance values for these elements in yerba mate and tea,
257 and observed levels were similar to other plant part levels (Kabata-Pendias, 2011).

258 Supporting findings of other studies (Giulian et al., 2007; Toppel et al., 2018), Al
259 concentrations in yerba mate leaves were high (Table 2) due to low pH of most evaluated
260 soils (Table 1). However, the *Barão de Cotegipe* soil had almost no exchangeable Al (Table
261 1) but had high levels in plant tissues. This illustrates the difficulty of associating soil Al
262 availability with leaf concentration (Fossati, 1997; Reissmann et al., 1999), and may suggest
263 that root exudation plays an important role in rhizospheric change and elemental uptake by
264 yerba mate. Although Al values of yerba mate leaves can be considered high, mean values
265 failed to reach the 1000 mg kg⁻¹ reported by Fossati (1997) and Reissmann et al. (1999).
266 Also, Al values were lower than those found in commercial tea (Marcos et al., 1998;
267 Fernandez-Caceres et al., 2001; Nookabkaew et al., 2006) and mature tea leaves (Fung et al.,
268 2009; Ruan et al., 2006; Zhong-lei et al., 2001).

269 Yerba mate is well known to accumulate large amounts of Mn in leaves without affecting
270 yield or resulting in toxicity (Fossati, 1997; Jacques et al., 2007; Barbosa et al., 2018).
271 However, the *Lebon Régis*, *Catanduvas*, *Bituruna*, and *Getúlio Vargas* sites had values that
272 exceeded the maximum expected value (3000 mg kg⁻¹) by 1.5 to 3.0 times (Table S3). The

273 highest Mn value was found at the *Lebon Régis* site, which had a unique combination of large
274 litter accumulation from “*taquara*” (*M. multiramea*) and native yerba mate plants that had
275 never been cut. This site also had the lowest soil pH among sampled sites. The fact that plants
276 were never cut indicated that leaves should be old and contain large amounts of Mn, similar
277 to observations for Mn in tea (Ruan et al., 2006). Two other sites with high Mn (*Bituruna* and
278 *Getúlio Vargas*) also had large amounts of decaying residue from recent hand thinning
279 intended to favour yerba mate growth. However, this same thinning practice at the *Ilópolis*
280 site did not result in similar Mn levels. Thus, more research should be conducted to evaluate
281 the influence of organic residue decay on Mn levels in yerba mate. From a nutritional aspect,
282 yerba mate could be an alternative Mn source, but it is unknown if high Mn concentrations
283 could result in oversupply for heavy yerba mate drinkers (Barbosa et al., 2015).

284 In reviews of this particular aspect, Mn concentrations were noted to be much higher than
285 Fe in yerba mate leaves (Reissmann et al., 1999; Barbosa et al., 2018) and commercial
286 products (Pozebon et al., 2015), with similar results for tea (Dambiec et al., 2013; Marcos et
287 al., 1998; Moreda-Piñeiro et al., 2003; Brzezicha-Cirocka et al., 2016). Since availability of
288 both elements can be related to soil pH and redox potential, it appears that both yerba mate
289 and tea crops are more efficient in the uptake of Mn. This has been observed for hydroponic
290 growth of yerba mate (Benedetti, 2012).

291 Remarkable concentrations of Zn were observed at many yerba mate sampling sites
292 (Table 2). No Zn was applied at any site, except the organically grown yerba mate at *Barão*
293 *do Cotegipe*. High Zn concentrations have been previously reported for a field study (Fossati,
294 1997) and under controlled hydroponic conditions (Benedetti, 2012). However, two sites had
295 values below 20 mg kg⁻¹ of Zn, which can be considered low. Jacques et al. (2007) found that
296 low light intensity/shade increases Zn concentration in yerba mate leaves, which could help

297 explain some of our observed values. Another important point to highlight is that many of our
298 soils were basalt derived and acidic, which results in good supply of Zn (Table 1).

299 Like Zn, the average Ni concentration of 3.6 mg kg⁻¹ was considered high. Similar values
300 were observed by others for yerba mate leaves (Toppel et al., 2018; Barbosa et al., 2019) and
301 commercial products (Pozebon et al., 2015). In general, tissue Ni concentration can be below
302 1 mg kg⁻¹ (Kabata-Pendias, 2011) and in our study only two samples were below this value.
303 This characteristic is comparable to commercially cropped tea (Scancar et al., 2013; Moreda-
304 Piñeiro et al., 2003; Brzezicha-Cirocka et al., 2016). Also soil acidity could play an important
305 role in explaining these findings.

306 In contrast to the wide variation in soil Cu seen in Table 1, foliar concentrations displayed
307 a more narrow range (7.3 to 14.0 mg kg⁻¹). Supporting our findings, Pozebon et al. (2015)
308 and Giulian et al. (2007) reported commercial yerba mate values between 7 to 20 mg kg⁻¹ for
309 different soils and climate conditions. A much wider range (2.0 to 447.5 mg kg⁻¹) was
310 reported by Han et al. (2005) for 800 Chinese tea samples, while others found mean values
311 close to 20 mg kg⁻¹ (Marcos et al., 1998; Zhong et al., 2015; Seenivasan et al., 2008). Higher
312 levels of Cu in tea could be related to Cu containing chemical products used to control fungal
313 diseases which are not used with yerba mate. However, Tsushida and Takeo (1977) reported
314 a low mean Cu value of 11.4 mg kg⁻¹ for 139 Camellia tea samples from Japan.

315 For B, the range concentration in leaves of 29.4-80.5 mg kg⁻¹ was within or above normal
316 levels. Similar values for B have been observed under field (Reissmann et al., 1999) and
317 hydroponic conditions (Benedetti, 2012) suggesting a high absorption capacity for yerba
318 mate. In contrast, the very low Mo concentrations found in leaf tissue confirm previous
319 observations made by Pozebon et al. (2015).

320 Phosphorous levels presented in Table S3 reflect low soil P availability (Table 1). The
321 majority of foliar samples (19 of 30 sites, Table S3) had values below the critical

322 concentration of 1000 mg kg⁻¹. Other researchers found low values of P in yerba mate leaves
323 (Barbosa et al., 2018) and commercial products (Giulian et al., 2007; 2009), which were far
324 below that found in tea (Brzezicha-Cirocka et al., 2016). As expected, the highest value was
325 observed at *Barão de Cotegipe* (Site 23), which was the only site with very high P
326 availability due to pig manure applications (Table 1). Also, *Barão de Cotegipe* was the only
327 site where P concentration was higher than S (Table 2). This could suggest that the majority
328 of our study sites had low P and high S supply. Assessed soils had an abundance of organic
329 matter as reflected by elevated C concentrations (Table 1) and deep, well defined soil A
330 horizons. High organic matter often indicates an abundant S supply since more than 90% of S
331 was found in organic matter in two soils under subtropical conditions (Paraná state, Brazil)
332 and in 6 soils from the state of Iowa, USA (Neptune et al., 1975). Additionally, the largest S
333 values were found at Site 5 (Table S3) where the soil is derived from black shale parent
334 material and the site is in proximity to black shale industrial mining/exploration; this
335 proximity could indicate a possible atmospheric contribution from these operations.

336 The lowest K concentration in yerba mate leaves was close to 10,000 mg kg⁻¹ (Table S3)
337 and similar to the lowest values reported by Reissmann et al. (1999) and Barbosa et al. (2019)
338 for leaves and Pozebon et al. (2015) for commercial yerba mate. These values can be
339 considered normal and were expected since only two soils had K availability between low to
340 medium, while all others were medium to high. The highest observed value (25,319 mg kg⁻¹)
341 was also close to findings of Reissmann et al. (1999) and could be related to luxury
342 consumption of K due to high soil availability under native conditions. Also, results (Table
343 S3) confirmed that K was by far the most abundant nutrient among those evaluated and that
344 yerba mate can be a good K source for consumers (Barbosa et al., 2015; Giulian et al., 2007;
345 Heinrichs and Malavolta, 2001). Despite this abundance, yerba mate had lower K

346 concentration compared to tea (Dambiec et al., 2013; Chen et al., 2009), which was likely due
347 to tea receiving K fertilizer applications by farmers.

348 The observed concentration of Mg (Table S3) was close to the 3600 – 7800 mg kg⁻¹ found
349 by Reissmann et al. (1999) in yerba mate leaves, and close to 4041 - 5574 mg kg⁻¹ found in
350 commercial yerba mate (Pozebon et al., 2015). Benedetti (2012) reported foliar
351 concentrations of 4550 - 5880 mg kg⁻¹ Mg for leaves of yerba mate grown in hydroponic
352 solution. Thus, the Mg range observed in this study (Table 2) can possibly be normal. These
353 findings are impressive considering that 25 sites had 0.6 cmol_c dm⁻³ or less exchangeable Mg,
354 indicative of low or very low availability (Table 1). Yerba mate appears to have a high
355 capacity for extracting less exchangeable Mg from soil. Also, our results indicated that yerba
356 mate has more Mg than tea. The concentration of Mg in tea has been shown to be below 3000
357 mg kg⁻¹ for leaves of plants grown in hydroponic solution (Fung et al. 2008) and in
358 commercial tea (Dambiec et al., 2013; Modera-Pineiro et al., 2003; Brzezicha-Cirocka et al.,
359 2016).

360 Although exchangeable Ca availability was low in 8 soils (Table 1), most leaf Ca values
361 (Table 2) fell within the 2600 - 8759 mg kg⁻¹ range for yerba mate leaves and products
362 reported by others (Reissmann et al., 1999; Giulian et al., 2007; Pozebon et al., 2015; Toppel
363 et al., 2018; Barbosa et a., 2019). Our values were also higher than observed in tea products.
364 Additionally, yerba mate appears to have a high capacity for Ca uptake from poor soils.

365

366 *3.2. Washing effect on elemental composition*

367 Leaf washing had a major impact on Fe, Ti, As, V, Pb, Li, Mo, and Al by decreasing
368 average, minimum, and maximum concentrations (Table 2 and Figure 2). In addition,
369 determined coefficients between washed and unwashed for Fe, Ti, As, and V had low values,
370 suggesting this influence increased variability among samples (Figure 4).

371 Dust deposition on leaf surfaces may have been responsible for variation in two sample
372 groups related to soil material from nearby roads, tillage activities, and possible atmospheric
373 deposition from fields at greater distances. Mineral contamination from soil has been related
374 to dust deposited on leaf surfaces and soil/plant element ratio related to harvest contamination
375 (Mayland and Sneva, 1983). Although Si, along with Fe, Al, and Ti, are major constituents of
376 all soil types (Kabata-Pendias, 2011), relatively low concentrations were observed in non-
377 washed yerba mate leaves (Table 2). However, high soil/plant ratios for Fe, Al, and Ti,
378 indicate that soil had a large influence on plant concentration as a contaminant in our study.
379 Many soils in this study (Table 1) also had a high degree of weathering, which increases
380 concentrations of Fe, Ti, and Al and reduces concentrations of K, Ca, Mg, and P. Soil/plant
381 ratios of K, Ca, Mg, and P that were close to or smaller than one could possibly explain the
382 lack of influence from washing. Additionally, soils originating from basalt (many soils in our
383 study), have high concentrations of Fe, Ti, and V (Mineropar, 2005), which increases the
384 chance of contaminating plant tissue concentrations.

385 Hoog (1981) analysed soil from Araucarian forests sites in Brazil (major yerba mate
386 ecosystem), and found the following range concentrations (mg kg^{-1}): 34250 - 122889, Fe;
387 60750 - 112571, Al; 850 - 4957, K; 250 - 1057, Ca; trace - 889, Mg; 214 - 689, P; and 125 -
388 789, Mn. These values were similar to those observed in our study. Additionally, the smallest
389 and highest reported values for total Fe were related to sandstone and basalt parent material,
390 respectively. *Paraná* state soil survey reported the following minimum, maximum, and
391 average concentrations: 6156, 32176, and 15726 mg kg^{-1} for Ti; 81, 856, and 370 mg kg^{-1} for
392 V; 15, 45, and 23 mg kg^{-1} for Pb; 2.40, 29.4, and 7.25 mg kg^{-1} for As; and 0.048, 0.457, and
393 0.187 for Cd (Mineropar, 2005). Soil concentrations were clearly much higher for Fe, Ti, and
394 Al or higher for V and As than corresponding concentrations found in yerba mate leaves,
395 while soil K, Ca, Mg, P, Mn, and Cd had similar or lower concentrations than plant leaves.

396 In this study, the high concentrations of Mn and Al found in yerba mate leaves (Table S3)
397 eliminate or reduced the possibility of plant contamination. Also, it is important to highlight
398 that Al, unlike Fe, Ti, and Mn, was found in soil as aluminium silicate that may not be
399 solubilised by plant tissue acids.

400 The average Fe concentration for unwashed samples was 50% higher than for washed
401 (Figure 2), suggesting a large portion of Fe found in commercial yerba mate comes from
402 external deposition of soil particles. In our study cut branches were placed into bags with
403 minimal soil contact. Our samples may have had less Fe contamination than commercial
404 products since their field handling, transportation, and loading procedures were likely
405 avenues of soil contamination. This could explain the large values observed for commercial
406 yerba mate samples reported by Pozebon et al. (2015) and Giulian et al. (2007; 2009).

407 Although the majority of collection sites had low industrial activity and were far from
408 major road traffic, washing influence on Pb concentration suggested that atmospheric
409 deposition may contribute to small increases. Many instances of atmospheric contributions of
410 Pb have been reported for tea (Han et al., 2006; Tsushida and Takeo, 1977; Jin et al., 2005).

411

412 3.3. Soil parent material effect on elemental composition

413 Considering washed leaf concentrations of Mn, P, Co, Ni, Mg, Al, Fe, K, Na, Ca, and Ba
414 according to sample region, there was no effective discrimination of groups (84 % precision
415 in LDA), primarily confusing samples from *Paraná* with north of *Santa Catarina*, as well
416 north of *Rio Grande do Sul* with west of *Santa Catarina* (Figure 3A). On the other hand,
417 when discriminating samples according to soil parent material there was better discrimination
418 (94 % precision – Figure 3B). This result indicates that the confusion generated in attempting
419 to separate observations by sample region was linked to the effect of soil parent material. In
420 others words, yerba mate composition (Mn, P, Co, Ni, Mg, Al, Fe, K, Na, Ca, and Ba) was

421 dependent on soil parent material, with high values of Mn, P, Co, Ni, and Ba in basalt soils,
422 Ni, Mg, Al, and Fe in sedimentary soils, and Fe, K, Na, Ca, and Ba in Rhyolite/Rhyodacite
423 soils (Table 3). Some of these elements also presented differences according to Kruskal
424 Wallis test, followed by Dunn's Test. Additionally, Table 3 shows that the yerba mate leaf
425 concentration often followed total soil concentration (Table S5).

426 Therefore, differences in elemental composition of yerba mate reported as geographical
427 differences (countries, states, or municipality) (Pozebon et al., 2015; Baran et al., 2018;
428 Kahmann et al. 2017; Saidelles et al., 2010; Bragança et al., 2011; Costa et al., 2009; Tatsch
429 et al., 2010) were probably linked to the soil parent material where the species was cultivated.

430

431 *3.4. Correlation among plant elements*

432 Correlations between K and Rb (0.73 and 0.73), Ca and Mg (0.55 and 0.54), Ca and Ba
433 (0.66 and 0.59), Mg and Sr (0.54 and 0.54), Mg and Ba (0.54 and 0.56), and P and Mo (0.51
434 and 0.52) were observed in both washed and unwashed samples (Figure 4), respectively. The
435 monovalent cations K and Rb have similar hydrate size and consequently exhibit similar
436 behaviour during plant uptake, allowing Rb isotopes to be used as substitutes for K in uptake
437 studies (Broadley, 2012). Additionally, total soil concentrations displayed similar distribution
438 patterns in a geologic survey of *Paraná* state (Mineropar, 2005). The two anions of P and Mo
439 can present synergic interactions. The same value or small differences in coefficient of
440 correlation confirmed a low or no influence of dust deposition on values and relationships.

441 Correlations between Fe and As (0.70 and 0.76), Fe and V (0.83 and 0.82), As and V
442 (0.78 and 0.76), and Ti and V (0.56 and 0.81) were observed for washed and unwashed
443 samples, respectively (Figure 4). Also, correlations for Fe and Ti (0.67) and As and Ti (0.64)
444 were only seen in unwashed samples. Plant concentrations of Fe, As, V, and Ti appear to
445 follow similar patterns. The influence of dust deposition was indicated by increased

446 correlations including Ti (i.e., presence of Fe and Ti and As and Ti correlations and increased
447 Ti and V correlations). Other correlations were observed between Al and Pb (0.55 and 0.56)
448 for washed and unwashed samples and Pb and As (0.51) for unwashed.

449

450 *3.5. Total soil concentration*

451 Total soil concentration (Table S5) showed a strong influence of parent material with Fe,
452 V, Ti, Cu, Ni, and Co having high values under basalt, which was expected (Figure 5).
453 Another aspect of basalt derived soils were the high Cu values similar to or greater than Zn,
454 which confirm observations of Motta et al. (2007) who pointed out two to three times higher
455 concentration of Cu vs Zn in basalt derived soils (Table S5). The same can be stated for Co
456 and Ni in some soils of basalt origin. While eruptive acid rocks (Rhyolite/Rhyodacites) can
457 influence K, Rb, Cs, Ba, Na, Li, and Mg (Figure 5). This was especially true for K, Cs, and
458 Rb which exhibit similar characteristics (charge and size) allowing for selective adsorption to
459 2:1 clay type (Zaunbrecher et al., 2015).

460 There was no indication that soil Cd and Pb concentrations were above the reference level
461 for southern soils. The mean, minimum, and maximum concentration values for Cd were
462 0.098, 0.021, 0.285 mg kg⁻¹, and for Pb were 20.89, 7.13, and 33.37 mg kg⁻¹. These values
463 were below the respective 0.59 and 36 mg kg⁻¹ values for Cd and Pb in soil derived from
464 basalt reported by Althaus et al. (2018). In addition, they were in the range of 0.36 - 0.42 mg
465 kg⁻¹ for Cd and 16 - 27 mg kg⁻¹ for Pb in soil derived from other parent material. Thus,
466 obtained values for Cd and Pb in plants were associated with normal soil conditions.

467

468 **4. Conclusions**

469 Soil analysis confirmed low pH and availability of Ca, Mg, and P as well as abundant
470 exchangeable Al, except in the soil under organic management. Although foliar analysis

471 indicated low leaf Cd background, 23% of washed leaves had values close to or above the 0.4
472 mg kg⁻¹ MAL for Brazil. However, three other potentially toxic elements (Pb, Cr, and As) had
473 low concentrations and were below the established MAL values for Pb and As.

474 High Mn concentrations reached up to 9000 mg kg⁻¹, which was 3 to 4 folds higher than
475 previously reported. The highest observed value could be related to particular forest
476 conditions where accumulation and decay of “*taquara*” resulted in very high residue amounts
477 under very old yerba mate trees. The Al concentration in yerba mate was high compared to
478 conventional crop species and reached values up to 1000 mg kg⁻¹ as noted in other studies.
479 High concentrations of Zn were observed in tissue, primarily related to basalt derived soil and
480 high availability under low soil pH. Yerba mate can possibly be an important supply for Mn
481 and Zn, however the possibility of Mn oversupply from heavy consumption of yerba mate
482 drinks needs further investigation. In contrast, high Cu availability did not result in high foliar
483 concentrations. Despite low soil availability, normal concentrations of Ca and Mg reflect the
484 root systems ability to exploit low availability nutrient forms. However, this was not true for
485 P, which had low tissues concentrations that mirrored low soil P availability.

486 The washing process showed that dust residue deposition increased Fe, Ti, As, Pb, Li,
487 Mo, and V concentrations of foliar tissue and that Fe, Ti, As, and V might be trace elements
488 from surface deposition of soil dust. This explains higher Fe values found in commercial
489 yerba mate compared to our study. Also, significant correlations between Fe and As and
490 between Fe and V for washed and unwashed samples and between Fe and Ti in unwashed
491 samples confirmed this relationship.

492 This study clearly shows that washing yerba mate during industrial processing may
493 alleviate some dust contamination issues, but will not reduce levels of some elements (e.g.,
494 Cd) below the current MAL. This work provides some suggested background concentration

495 ranges for native yerba mate that could potentially be used as guidelines for setting more
496 realistic levels in future legislations.

497

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505

506 **References**

507 ANVISA (2013). Resolução - RDC N° 42, de 29 de agosto de 2013. Dispõe sobre o
508 Regulamento Técnico MERCOSUL sobre limites máximos de contaminantes
509 inorgânicos em alimentos., Brasil, DOU p. 1.

510 Althaus, D., Gianello, C., Tedesco, M.J., Silva, K.J.D., Bissani, C.A., Felisberto, R. 2018.
511 Natural fertility and metals contents in Soils of Rio Grande do Sul (Brazil). Revista
512 Brasileira de Ciência do Solo, 42, 1-15.

513 Alvares, C.A., Stape, J.L., Sentelhas, P.C., Gonçalves J.L. de M., Sparovek, G. 2013.
514 Köppen's climate classification map for Brazil. Meteorol. Z. 22, 711-728.
515 <https://doi.org/10.1127/0941-2948/2013/0507>.

516 Baran, A., Gruszecka-Kosowska, A., Kołton, A., Jasiewicz, C., Piwowar, P. 2018. Content
517 and health risk assessment of selected elements in the yerba mate (*Ilex paraguariensis*,
518 St. Hil.). Hum. Ecol. Risk Assess. 24(4), 1092-1114.

519 Barbosa, J.Z., Zambon, L.M., Motta, A.C.V., Wendling, I. 2015. Composition, hot-water
520 solubility of elements and nutritional value of fruits and leaves of yerba-mate. *Ciênc.*
521 *Agrotec.* 39(6), 593-603.

522 Barbosa, J.Z., Motta, A.C., Consalter, R., Poggere, G.C., Santin, D., Wendling, I. 2018. Plant
523 growth, nutrients and potentially toxic elements in leaves of yerba mate clones in
524 response to phosphorus in acid soils. *An. Acad. Bras. Ciênc.* 90(1), 557-571.

525 Barbosa, J.Z., Motta, A.C.V., Reis, A.R., Corrêa, R.S., Prior, S.A. 2019. Spatial distribution
526 of structural elements in leaves of *Ilex paraguariensis*: physiological and ecological
527 implications. *Trees* 34, 101-110. <https://doi.org/10.1007/s00468-019-01900-y>

528 Benedetti, E.L. 2012. Tolerância da erva mate (*Ilex paraguariensis* St. Hil.) ao alumínio. 72
529 p. Tese (Doutorado em Solos e Nutrição de Plantas), Universidade Federal de Viçosa,
530 Viçosa.

531 Bora, K., Sarkar, D., Konwar, K., Payeng, B., Sood, K., Paul, R.K., Datta, R., Das, S., Khare,
532 P., Karak, T. 2019. Disentanglement of the secrets of aluminium in acidophilic tea plant
533 (*Camellia sinensis* L.) influenced by organic and inorganic amendments. *Food Res. Int.*
534 120, 851-864. <https://doi.org/10.1016/j.foodres.2018.11.049>.

535 Bragança, V.L.C., Melnikov, P., Zanoni, L.Z. 2011. Trace elements in different brands of
536 yerba mate tea. *Biol. Trace Elem. Res.* 144, 1197–1204.

537 Broadley, M., Brown, P., Cakmak, I., Rengel, Z., Zhao, F. 2012. Function of nutrients:
538 micronutrients. In: Marschner's mineral nutrition of higher plants. Academic Press, p.
539 191-248.

540 Brzezicha-Cirocka, J., Grembecka, M., Szefer, P. 2016. Monitoring of essential and heavy
541 metals in green tea from different geographical origins. *Environ. Monit. Assess.* 188
542 (3), 183. <https://doi.org/10.1007/s10661-016-5157-y>.

543 Bugang, W., Woolsey, M. 2010. National food safety standard-maximum levels of
544 contaminants in food. People's Republic of china. Gain Report Number-CH10000.

545 Cardozo-Junior, E.L., Morand, C. 2016. Interest of mate (*Ilex paraguariensis* A. St. Hil.) as a
546 new natural functional food to preserve human cardiovascular health-A review. J.
547 Funct. Foods 21, 440-454.

548 Chen, Y., Yu, M., Xu, J., Chen, X., Shia, J. 2009. Differentiation of eight tea (*Camellia*
549 *sinensis*) cultivars in China by elemental fingerprint of their leaves. J. Sci. Food Agric.
550 89, 2350-2355.

551 Chen, Y, Xu, J. Yu, M., Chen, X, Shi, J. 2010. Lead contamination in different varieties of
552 tea plant (*Camellia sinensis* L.) and factors affecting lead bioavailability. J. Sci. Food
553 Agric. 90, 1501-1507.

554 Costa, A.M.G. da, Nogami, E.M., Visentainer, J.V., Souza, N.E. de, Garcia, E.E. 2009.
555 Fractionation of aluminum in commercial green and roasted yerba mate samples (*Ilex*
556 *paraguariensis* St. Hil.) and in their infusions. J. Agric. Food Chem. 57, 196–200.

557 Dambiec, M., Polechonska, L., Klink, A. 2013. Levels of essential and non-essential
558 elements in black teas commercialized in Poland and their transfer to tea infusion. J.
559 Food Compos. Anal. 31, 62–66.

560 Donagema, G.K., de Campos, D.B., Calderano, S.B., Teixeira, W.G., Viana, J.M. 2011.
561 Manual de métodos de análise de solo. Embrapa Solos-Documents (INFOTECA-E).

562 Danner, M. A., Zanette, F., Ribeiro, J.Z. 2012. O cultivo da araucária para produção de
563 pinhões como ferramenta para a conservação. Pesquisa Florestal Brasileira 32 (72),
564 441-451.

565 Fernandez-Caceres, P.L., Martin, M.J., Pablos, F., Gonzalez, A.G. 2001. Differentiation of
566 tea (*Camellia sinensis*) varieties and their geographical origin according to their metal
567 content. J. Agric. Food Chem. 49, 4775-4779.

- 568 Fossati, L.C. 1997. Avaliação do estado nutricional e da produtividade de erva-mate (*Ilex*
569 *paraguariensis* St. Hil.), em função do sítio e da dioícia. 126 pg.
- 570 Fung, K. F., Carr, H. P., Zhang, J., Wong, M.H. 2008. Growth and nutrient uptake of tea
571 under different aluminium concentrations. *J. Sci. Food Agric.* 88, 1582–1591.
- 572 Fung, K.F., Carr, H.P., Poon, B.H.T., Wong, M.H. 2009. A comparison of aluminum levels
573 in tea products from Hong Kong markets and in varieties of tea plants from Hong Kong
574 and India. *Chemosphere* 75, 955–962.
- 575 Giulian, R., Santos, C.E.I. dos, Shubeita, S. de M., Silva, L.M. da, Dias, J.F., Yoneama, M.L.
576 2007. Elemental characterization of commercial mate tea leaves (*Ilex paraguariensis* A.
577 St.-Hil.) before and after hot water infusion using ion beam techniques. *J. Agric. Food*
578 *Chem.* 55, 741-746.
- 579 Giulian, R., Santos, C.E.I. dos, Shubeita, S. de M., Silva, L.M. da, Yoneama, M.L., Dias, J.F.
580 2009. The study of the influence of industrial processing on the elemental composition
581 of mate tealeaves (*Ilex paraguariensis*) using the PIXE technique. *Food Sci. Technol.*
582 42, 74-80.
- 583 Han, W.Y., Zhao, F.J., Shi, Y.Z., Ma, L.F., Ruan, J.Y. 2006. Scale and causes of lead
584 contamination in Chinese tea. *Environ. Pollut.* 139, 125-132.
- 585 Han, W.Y., Shi, Y.Z., Ma, L.F., Ruan, J.Y. 2005. Arsenic, Cadmium, Chromium, Cobalt, and
586 Copper in Different Types of Chinese Tea. *Bull. Environ. Contam. Toxicol.* 75, 272–
587 277.
- 588 He, H., Bleby, T.M., Veneklaas, E.J., Lambers, H., Kuo, J. 2012. Precipitation of calcium,
589 magnesium, strontium and barium in tissues of four *Acacia* species (Leguminosae:
590 Mimosoideae). *PLoS one*, 7(7).
- 591 Heinrichs, R., Malavolta, E. 2001. Mineral composition of a commercial product from mate-
592 herb (*Ilex paraguariensis* St. Hil.). *Ciênc. Rural* 31, (5), 781-785.

593 Hoog, R.J. de. 1981. Site-nutrition-growth relationships of *Araucaria angustifolia* (Bert.) O.
594 Ktze. in Southern Brazil. Freiburg: Universidade de Freiburg, 161 p. (Doctoral Thesis).
595 IBGE. 2018. Instituto Brasileiro de Geografia E Estatística. Produção Agrícola Municipal
596 (PAM), Erva-mate (ano base 2016), IBGE.

597 Jacques, R.A., Arruda, E.J., Oliveira, L.C.S. de., Oliveira, A.P. de., Dariva, C., Oliveira, J.V.
598 de, Carama, E.B. 2007. Influence of agronomic variables on the macronutrient and
599 micronutrient contents and thermal behavior of mate tea leaves (*Ilex paraguariensis*). J.
600 Agric. Food Chem. 55, 7510-7516.

601 Jin, C.W., He, Y.F., Zhang, K., Zhou, G.D., Shi, J.L., Zheng, S.J. 2005. Lead contamination
602 in tea leaves and non-edaphic factors affecting it. Chemosphere 61, 726–732.

603 Kabata-Pendias, A. 2011. Trace Elements in Soils and Plants 4th ed. p. 534. CRC Press.
604 Taylor and Francis Group. Boca Raton.

605 Kahmann, A., Anzanello, M.J., Marcelo, M.C.A., Pozebon, D. 2017. Near infrared
606 spectroscopy and element concentration analysis for assessing yerba mate (*Ilex*
607 *paraguariensis*) samples according to the country of origin. Computers and Electronics
608 in Agriculture, 140, 348-360.

609 Marcos, A., Fisher, A., Rea G., Hill, S.J. 1998. Preliminary study using trace element
610 concentrations and a chemometrics approach to determine the geographical origin of
611 tea. J. Anal. Atom. Spectrom. 13, 521–525.

612 Marques, A. da C. 2014. As paisagens do mate e a conservação socioambiental: um estudo
613 junto aos agricultores familiares do planalto norte catarinense. 434 p. (Tese, curso de
614 Doutorado em Meio Ambiente e Desenvolvimento da Universidade Federal do Paraná).

615 Mayland. H.F., Sneva, F.A. 1983. Effect of soil contamination on the mineral composition of
616 forage fertilized with nitrogen. J. Range Manage. 38(3), 286-289.

617 Mineropar, 2005. Geoquímica de solo-Horizonte B. Relatório final de projeto. Curitiba:
618 Mineropar, 2 vol. 453 p.

619 Moreda-Piñeiro, A., Fisher, A., Hill, S.J. 2003. The classification of tea according to region
620 of origin using pattern recognition techniques and trace metal data. J. Food Compos.
621 Anal. 16, 195–211.

622 Motta, A.C.V., Serrat, B.M., Reissmann, C.B., Dionisio, J.A. 2007. Micronutrientes na rocha,
623 no solo e na planta. Curitiba: UFPR. 246 p.

624 Neptune, A.M.L., Tabatabai, M.A., Hanway, J.J. (1975). Sulfur fractions and
625 carbon-nitrogen-phosphorus-sulfur relationships in some Brazilian and Iowa soils. Soil
626 Sci. Soc. Am. J. 39(1), 51-55.

627 Nookabkaew, S., Rangkadilok, N., Satayavivad, J. 2006. Determination of trace elements in
628 herbal tea products and their infusions consumed in Thailand. J. Agric. Food Chem. 54,
629 6939-6944.

630 Pinotti, L.C.A., Hanisch, A.L., Negrelle, R.R.B. 2018. The impact of traditional silvopastoral
631 system on the Mixed Ombrophilous Forest remnants. Floresta Ambient, 25(4),
632 <http://dx.doi.org/10.1590/2179-8087.019217>

633 Pozebon, D., Dressler, V.L., Marcelo, M.C.A., Oliveira, T.C. de, Ferrão, M.F. 2015. Toxic
634 and nutrient elements in yerba mate (*Ilex paraguariensis*). Food Addit. Contam. B 8(3),
635 215-220.<http://dx.doi.org/10.1080/19393210.2015.1053420>

636 R Core Team, 2019. R: A language and environment for statistical computing. R Foundation
637 for Statistical Computing, Vienna, Austria. <<https://www.R-project.org/>>.

638 Reissmann, C.B., Radomski, M.I., Quadros, R.M.B. 1999. Chemical composition of *Ilex*
639 *paraguariensis* St. Hil. under different management conditions in seven localities of
640 Paraná state. Braz. Arch. Biol. Techn. 42, 187-194.

641 Ruan, J., Ma, L., Shi, Y. 2006. Aluminium in tea plantations: mobility in soils and plants, and
642 the influence of nitrogen fertilization. *Environ. Geochem. Health* 28, 519–528.

643 Saidelles, A.P.F., Kirchner, R.M., Santos, N.R.Z. dos, Flores, E.M. de M., Bartz, F.R. 2010.
644 Análise de metais em amostras comerciais de erva-mate do sul do Brasil. *Alim. Nutr.*
645 21(2), 259-265.

646 Santos, L.M.G.D., Neto, V., Alves, S., Iozzi, G., Jacob, S.D.C. 2017. Arsenic, cadmium and
647 lead concentrations in yerba mate commercialized in southern Brazil by inductively
648 coupled plasma mass spectrometry. *Ciênc. Rural* 47(12).
649 <http://dx.doi.org/10.1590/0103-8478cr20170202>.

650 Scancar, J., Zuliani, T., Milacic, R. 2013. Study of nickel content in Ni-rich food products in
651 Slovenia. *J. Food Compos. Anal.* 32, 83–89.

652 Seenivasan, S., Manikandan, N., Muraleedharan, N.N., Selvasundaram, R. 2008. Heavy
653 metal content of black teas from south India. *Food Control* 19, 746–749.

654 Tatsch, F.R.P., Gonçalves, V.C., Meurer, E.J. 2010. Alumínio total e solúvel em amostras de
655 erva-mate comercializadas no Rio Grande do Sul. *Scientia Agraria* 11(1), 83-86.

656 Toppel, F.V., Junior, A.M., Motta, A.C.V., Frigo, C., Magri, E., Barbosa, J.Z. 2018. Soil
657 chemical attributes and their influence on elemental composition of yerba mate leaves.
658 *Floresta* 48(3), 425-434.

659 Tsushida, T., Takeo, T. 1977. Zinc, copper, lead and cadmium contents in green tea. *J. Sci.*
660 *Food Agric.* 28, 255-258. <https://doi.org/10.1002/jsfa.2740280306>.

661 Valduga, A. T., Gonçalves, I. L., Magri, E. 2019a. Analysis of the Presence of Toxic Metals
662 in Yerba Mate Samples: a Case Study from South Brazil. *Water Air Soil Pollut.* 230(7),
663 153.

664 Valduga, A.T., Gonçalves, I.L., Magri, E., Finzer, J.R.D. 2019b. Chemistry, pharmacology
665 and new trends in traditional functional and medicinal beverages. *Food Res. Int.* 120,
666 478-503.

667 Zaunbrecher, L.K., Cygan, R.T., Elliott, W.C. 2015. Molecular models of cesium and
668 rubidium adsorption on weathered micaceous minerals. *J. Phys. Chem. A*, 119(22),
669 5691-5700.

670 Zheng, H., Li, J., Li, H., Hu, G. Li, H. 2014. Analysis of trace metals and perfluorinated
671 compounds in 43 representative tea products from South China. *J. Food Sci.* 79(6),
672 1123-1129. <https://doi.org/10.1111/1750-3841.12470>.

673 Zhong, W., Ren, T., Zhao, L. 2015. Determination of Pb (Lead), Cd (Cadmium), Cr
674 (Chromium), Cu (Copper), and Ni (Nickel) in Chinese tea with high-resolution
675 continuum source graphite furnace atomic absorption spectrometry. *J. Food Drug Anal.*
676 4 (1), 46-55. doi: 10.1016/j.jfda.2015.04.010.

677 Zhong-Lei, X., De-Ming, D., Guo-Zhang, B., Sheg-Tian, W. Yao-Guo, D. Li-Min, Q. 2001.
678 Aluminium content of tea leaves and factors affecting the uptake of aluminium from
679 soil into tea leaves. *Chinese Geogr. Sci.* 11(1), 87-91.

680