

1 **Bioleaching for resource recovery from low-grade wastes like fly and bottom ashes from**  
2 **municipal incinerators: A SWOT analysis**

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11 **Highlights**

- 12 • Circular economy depends on resource recovery from wastes
- 13 • Strengths, Weaknesses, Opportunities, and Threats (SWOT) identified for bioleaching
- 14 • Bioleaching can be a sustainable technique for metal recovery from low-grade wastes
- 15 • Resource recovery demands integrated policy and regulatory framework

16

17 **Abstract**

18 Bioleaching (or microbial leaching) is a biohydrometallurgical technology that can be  
19 applied for metal recovery from anthropogenic waste streams. In particular, fly ashes and  
20 bottom ashes of municipal solid waste incineration (MSWI) can be used as a target material  
21 for biomining. Globally, approximately 46 million tonnes of MSWI ashes are produced  
22 annually. Currently landfilled or used as aggregate, these contain large amounts of marketable  
23 metals, equivalent to low-grade ores. There is opportunity to recover critical materials as the  
24 circular economy demands, using mesophile, moderately thermophile, and extremophile  
25 microorganisms for bioleaching. A Strengths, Weaknesses, Opportunities and Threats (SWOT)  
26 analysis was developed to assess the potential of this biotechnology to recover critical metals

27 from MSWI wastes. Bioleaching has potential as a sustainable technology for resource  
28 recovery and enhanced waste management. However, stakeholders can only reap the full  
29 benefits of bioleaching by addressing both the technical engineering challenges and regulatory  
30 requirements needed to realise and integrated approach to resource use.

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32 **Keywords:** Circular economy; critical raw materials; biotechnology; municipal solid waste  
33 incineration – MSWI; waste management

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## 35 **1. Introduction**

36 Our current patterns of consumption are leading to the exhaustion of planetary resources,  
37 while simultaneously generating pollution and threatening our survival as a species. There is  
38 an urgent need for a change in paradigm in waste management to efficiently recover resources  
39 (energy, metals, nutrients) from waste streams, making industrial and urban processes more  
40 efficient. According to the recent UN Global Resources Outlook, extractive industries are  
41 responsible for half of the global carbon emissions (Oberle et al., 2019). Resource extraction  
42 and processing caused 90% of biodiversity loss and water stress, which puts a more dangerous  
43 level of pressure on climate and natural life support systems than previously thought. Resources  
44 are being extracted three times faster than in 1970, even though the population has only doubled  
45 in that time, according to the same report. The United Nations Sustainable Development Goals  
46 (SDG) set a target to achieve sustainable management and efficient use of natural resources by  
47 2030, by decoupling economic growth from resource use and environmental degradation. This  
48 is to achieved by improved resource efficiency, decreased reliance on raw materials, and  
49 increased recycling to reduce environmental pressure and impact. “Circular economy” and  
50 “zero waste” are buzzwords today, but both goals still look unattainable (Velenturf et al., 2019).

51 Metals are valuable raw materials for the global economy, and key components of various  
52 products such as low-carbon energy technologies, electric vehicles, and electronic and  
53 biomedical devices. Large quantities of critical and scarce metals, such as platinum group  
54 metals (PGM), rare earth elements (REE), cobalt, vanadium, selenium, and tellurium are  
55 required for storage and production of renewable energies, catalytic processes, digital  
56 communication, and green technologies (Hofmann et al., 2018; Işıldar et al., 2019). Raw  
57 materials and metals are considered critical when they have high economic importance  
58 combined with an elevated risk of supply, mainly resulting from high production in countries  
59 with poor governance (geopolitical instability), limited material replaceability, and low end-  
60 of-life recycling rates (EC, 2017).

61 Municipal solid waste incineration (MSWI) has been globally adopted for the management  
62 of vast amounts of waste (a global average of 130 tonnes per year) (Joseph et al., 2018), as it  
63 allows energy recovery and reduction in the volume of waste sent to landfill. However, the  
64 incineration of municipal solid waste destroys technical value given that once the energy value  
65 of waste has been recovered by burning, it is no longer available to the circular economy  
66 (Purnell, 2019). Coarse metals (>2mm) remaining in fly ashes and bottom ashes are typically  
67 recovered, and the residual fraction is recycled as construction material. However, low but  
68 significant quantities of high-tech, high-value metals remain in the residual material. Estimated  
69 annual flows of these high-value metals are in the order of tens of kg and a total content  
70 comparable to low-grade active mines (Funari et al., 2015). The potential for urban mining and  
71 recovery of secondary resources from MSWI residues is, therefore, promising as a way of  
72 closing the loop within a circular economy (Simon and Holm, 2018).

73 Given the low metal concentrations, economic feasibility is dependent on using low cost,  
74 sustainable technologies, such as bioleaching. Bioleaching is commercially used for the  
75 recovery of metals from low and waste grade ores, in particular for copper (Gomes et al., 2018).

76 These ores would typically be uneconomic to process using conventional comminution-  
77 concentration-flotation routes. Bioleaching uses microorganisms isolated from natural settings  
78 (e.g. extreme environments, acid mine drainage) to generate mineral or organic acid (as  
79 metabolites) and improve metal solubility by enzymatic reactions. Bioleaching can be  
80 performed by direct contact (primary bioleaching) and by indirect leaching (or secondary  
81 bioleaching). The latter only uses the acid produced by bacteria to recover metals without a  
82 direct inoculation. This approach may be best suited in some circumstances, e.g. for alkaline  
83 wastes where the conditions are not favourable for the survival of the typically acidophilic  
84 bacteria. However, both approaches are acknowledged to lessen environmental and economic  
85 drawbacks during treatment of anthropogenic waste, compared to using mineral acids (Funari  
86 et al., 2017).

87 This paper aims to assess opportunities and limitations associated with bioleaching of low-  
88 grade wastes for metal recovery, in the context of the circular economy. It is beyond the scope  
89 of this study to provide a comprehensive review of the latest developments of bioleaching. In  
90 recent years, there has been an increasing amount of literature on bioleaching, with several  
91 extensive reviews covering the topic (Auerbach et al., 2019; Pollmann et al., 2018; Sethurajan  
92 et al., 2018; Srichandan et al., 2019). Our objectives are to focus on practical aspects and  
93 limitations that remain obstacles for the full implementation of bioleaching as a crucial tool for  
94 the circular economy.

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## 96 **2. Bioleaching as an alternative for resource recovery from MSWI waste**

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98 Current alternatives used at industrial scale for resource recovery from MSWI fly ashes are  
99 based on acid leaching (e.g. FLUWA process), and washing processes for salt recovery (e.g.  
100 HALOSEP®) (Quina et al., 2018). At lab scale, hydrometallurgical processes for copper and

101 zinc recovery (Tang et al., 2018), electro dialytic processes (Kirkelund et al., 2015), and  
102 treatment processes combining leaching, selective extraction and adsorption (Tang et al., 2019)  
103 have shown promising results, but are still at low technology readiness levels (TRL).

104 A recent review of the different removal techniques for metals from fly ash compared  
105 bioleaching, carrier in pulp method, chemical extraction with different acids, alkaline leachates  
106 and chelating agents, chloride evaporation process, electro dialytic and thermal treatments  
107 (Meer and Nazir, 2018). The authors concluded that the selection of the best process depends  
108 on the type of fly ash and target metal(s), but also of factors like cost, time and energy (Meer  
109 and Nazir, 2018).

110 Bioleaching of bottom ash (BA) and fly ash (FA) with a mixed culture isolated from a  
111 natural system showed good yields of metal extraction, with more than 90% Zn, Cu, and 10%  
112 Pb removed from FA; while 100% Cu, 80% Zn and 20% Pb were removed from BA samples  
113 (Funari et al., 2019). Bioleaching of bottom ashes with pure cultures *Acidithiobacillus*  
114 *ferrooxidans* or *Leptospirillum ferrooxidans*, or a mixture of *Acidithiobacillus thiooxidans* and  
115 *Acidithiobacillus ferrooxidans* in batch tests showed that Al, Cr, Cu, Ni, Mn and the rare earth  
116 elements Ce, La, and Er were significantly more extracted with iron-oxidizing bacteria  
117 compared to abiotic controls (Auerbach et al., 2019). The results are encouraging for industrial  
118 application to recover concentrated metals like Al and Cu, simultaneously reducing the cost of  
119 landfilling the remaining residues. Continuous heap bioleaching at lab scale showed leaching  
120 yields for zinc and copper between 18–53% and 6–44% (Mäkinen et al., 2019), but also  
121 highlighted the need for further optimisation, in particular regarding acid addition and aeration.

122 Bioleaching using alkaline autochthonous extremophiles isolated from a fly ash landfill site  
123 showed that *Alkalibacterium sp.* TRTYP6 could recover 52% of Cu (Ramanathan and Ting,  
124 2016). The use of fungi for fly ash bioleaching showed that fungal morphology of *Aspergillus*

125 *niger* was significantly affected during one-step and two-step bioleaching, with precipitation  
126 of calcium oxalate hydrate crystals at the surface of hyphae (Xu et al., 2014).

127 Several factors can influence bioleaching efficacy, such as pH, temperature, pulp density,  
128 redox potential, microorganisms or communities involved, particle size, oxygen and iron  
129 concentrations, and wastes mineralogy (Sethurajan et al., 2018). Also, metal bioleaching is  
130 influenced by biomass concentration, metal tolerance of microorganisms, type and amount of  
131 metabolic products released into the medium, contact time, and pretreatment (e.g. heating) and  
132 has to be assessed case by case (Pollmann et al., 2018). Nevertheless, bioleaching can be a  
133 flexible and environmentally friendly alternative for conventional processes, as it allows the  
134 recovery of valuable resources, but can also reduce the toxicity of the waste for further reuse  
135 in other applications (e.g. aggregate materials) (Auerbach et al., 2019).

### 136 **3. SWOT analysis**

137 Strengths, Weaknesses, Opportunities, and Threats (SWOT) analysis is a framework  
138 technique used in business to facilitate the development of a sustainable market niche by  
139 uncovering new outlooks and identifying problems that would hinder progress (Miller, 2007).  
140 Table 1 summarises a SWOT analysis for the use of bioleaching for metal recovery from MSWI  
141 fly ashes and bottom ashes.

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144 **Table 1. Strengths, Weaknesses, Opportunity and Threats (SWOT) analysis for**  
 145 **bioleaching for resource recovery from MSWI residues.**

<b>Strengths</b>	<b>Weaknesses</b>
<ul style="list-style-type: none"> <li>- Lower environmental footprint (avoids strong mineral acids used in hydrometallurgy methods, lower energy consumption)</li> <li>- Not labour intensive</li> <li>- Different microorganisms (fungi, isolate/mixed acidophilic/alkaline bacteria) can be used for bioleaching</li> <li>- Bioleaching can be achieved in one-step, two-step and spent medium-step</li> <li>- Minimal investment and low operating costs compared with hydrometallurgy methods</li> <li>- Can be performed <i>in situ</i></li> <li>- Technology readiness level 9 for primary ores</li> <li>- Can be used to reduce contamination of wastes/biostabilization</li> </ul>	<ul style="list-style-type: none"> <li>- Depends on quantities/concentrations of metals in wastes</li> <li>- Presents slow dissolution kinetics and low metal leaching yield</li> <li>- Heap bioleaching can be space demanding</li> <li>- Adaptation of microorganisms to waste materials is critical</li> <li>- Alkaline wastes are not favourable to the growth of acidophile bacteria</li> <li>- More data on alkaline bioleaching are needed</li> <li>- Not fully reproducible, as it depends on the feedstock material</li> <li>- Inhibitory layers hinder cell-mineral interaction (passivation effects)</li> <li>- Lack of research dedicated to process development and reactor design</li> <li>- No pilot-scale applications for anthropogenic streams such as FA and BA from MSWI</li> </ul>
<b>Opportunities</b>	<b>Threats</b>
<ul style="list-style-type: none"> <li>- Recovery of metals from low-grade ores and wastes can be a potential offset for remediation and operation costs in both operating and legacy sites</li> <li>- Development and use of bioelectrochemical systems for energy production and metal recovery</li> <li>- Minimise the environmental impacts of raw materials extraction</li> <li>- Fine-tuned bioleaching to enhance selectivity to targeted metals</li> <li>- Accelerate carbonation and carbon sequestration using microorganisms</li> </ul>	<ul style="list-style-type: none"> <li>- High volatility of markets and metal prices</li> <li>- Low cost of waste disposal in landfill</li> <li>- Low mineral extraction costs (“mineralogical barrier”)</li> <li>- No alternatives for the downstream processing of excessive biomass produced</li> <li>- Lack of satisfactory coverage for substrate materials and inocula</li> <li>- Lack of work practices may lead to poor reproducibility</li> <li>- Higher bioavailability of potentially toxic metals</li> </ul>

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### 148 3.1 Strengths

149 Bioleaching is a technology considered environmentally friendly, for being less aggressive  
150 (lower use of concentrated acids/bases) and energy-intensive than traditional  
151 hydrometallurgy methods. It can be performed *in situ*, as heap bioleaching, with minimal  
152 investment, and low operating costs. The process is not labour-intensive since only a few  
153 process parameters need essential monitoring and control. Unskilled operators can perform  
154 the key maintenance operations with minimal risks (e.g., no use of strong acids like  
155 hydrofluoric acid), including the manipulation of bacterial matter. MSWI bioleaching  
156 bacteria show low biological hazard and risk of contamination, with no identified risks to  
157 humans, especially those cultures isolated from natural systems of well-known  
158 characteristics. Moreover, bioleaching combines traditional hydrometallurgical methods  
159 effortlessly and in a cost-effective manner, because it does not require highly sophisticated  
160 monitoring and control instruments, the implementation of which can be expensive. Pure  
161 bacterial cultures, mixed (acidophilic or alkaline) bacteria, archaea, fungi, algae and plants  
162 metabolites can be used for bioleaching of different metal-bearing wastes. The bioleaching  
163 process can be achieved by one-step, two-step and spent medium-step, both in batch or  
164 continuous modes and showed promising results on the extraction of several metals (Ni, Co,  
165 Mo, V, Fe, Zn, Cu, Cr, Cd, W, Pb and Mn) (Srichandan et al., 2019). Metal extraction takes  
166 place directly by electron supply (e.g. oxidation or reduction reactions) or indirectly by  
167 metabolic products of the microorganisms (inorganic or organic acids or excretion of  
168 complexing agents) (Auerbach et al., 2019). Bioleaching is a fully developed technology  
169 (TRL 9 for primary ores), and has been used at industrial level since 1960 for extraction of  
170 copper from sulphide ores. Currently, approximately 20% of the worldwide Cu is extracted  
171 using bioleaching (Latorre et al., 2016). Bioleaching can reduce contamination of wastes,  
172 contributing to their biostabilization and can be used to extract metals from ores or

173 secondary wastes that are too low grade and therefore uneconomic to processing, using  
174 hydrometallurgical or pyrometallurgical methods.

### 175 **3.2. Weaknesses**

176 For further implementation of bioleaching as a technology for the circular economy, there  
177 is a need to improve process dissolution kinetics and metal leaching yields. Currently,  
178 bioleaching is slower than traditional extraction methods. Dissolution kinetics can be  
179 accelerated by optimisation of process parameters such as pre-treatments, reaction time, pH,  
180 temperature, mass transfer rate, nutrient requirements, pulp density, and particle size. Further  
181 developments in understanding the biotic factors controlling the inhibitory effects of pulp  
182 density on metal extraction are needed (Valix, 2017). Both single-stage, multistage, batch, and  
183 continuous stirred tank reactors can be developed and tested. To facilitate implementation on  
184 waste matrices, there is also a need to assess the process efficiency in larger, commercial  
185 relevant scale reactors, whilst focusing on process development and reactor design.

186 Fine-tuned processes assisted with microorganisms do exist, leading to the production of  
187 engineered inocula tailored to the target material and capable of high metal tolerance and  
188 improved selectivity towards the metals wanted as secondary resource (especially Cu, Co, Mn,  
189 V, Zn). The growing use of indigenous bacteria adapted to the environments instead of well-  
190 known strains from lab collections may overcome some of the limitations in terms of  
191 processing times (Pollmann et al., 2018). However, it will be challenging to use bioleaching  
192 for recovery of just one element, as further separation and purification techniques will be  
193 needed for circularity. The use of mixed culture instead of pure strain shows remarkable  
194 synergistic effects, especially against heavy metals that inhibit biomass growth, although pure  
195 cultures might demonstrate improved selectivity for the recovery of individual or groups of  
196 elements. The use of iron and sulphur (for acidophilic bioleaching) or organic sources (fungal  
197 and cyanogenic bioleaching) might increase bioleaching costs, as well as the need to add acid

198 to keep the medium pH low (Srichandan et al., 2019). Adaptation of microorganisms to the  
199 waste materials can be critical, primarily due to heterogeneous feedstock composition, high  
200 buffering, and passivation effects. Alkaline wastes like MSWI ashes are unfavourable for  
201 bioleaching using acidophile cultures, so more data on alkaline bioleaching are needed other  
202 than fungal bioleaching, which showed limited performances in terms of metal yield,  
203 biosorption capacity, and volumes of biomass produced (Luo et al., 2019). Controlling the  
204 formation of inhibitory layers may overcome cell growth disruption and decrease metal  
205 extraction where a reasonable trade-off between microbial community succession and their  
206 energy types metabolisms can be maintained.

207 Regarding microbial development, there is also a need to better understand partnering of  
208 organisms and cell adaptation to the toxic effects of not only metals, but also toxic organic  
209 contaminants in the wastes (e.g. polychlorinated dibenzo-p-dioxins) (Valix, 2017). Metal  
210 separation from the bio-leachate also demands cost-effective and selective processes to recover  
211 metal ions. A major challenge is the recovery of low concentrations of metals from large  
212 volumes of dilute leachates (Pollmann et al., 2018).

### 213 **3.3. Opportunities**

214 Bioleaching has potential to recover metals from low-grade ores and wastes, but also to  
215 offset remediation (legacy sites) and operation costs by valorising wastes. This can also  
216 contribute to minimising the environmental impacts of raw materials extraction as potentially  
217 toxic elements are not discharged to the environment, but recovered for the circular economy.  
218 The importance of resource recovery in reducing carbon emissions could receive increased  
219 attention and should be leveraged to support the development of bioleaching. It can be expected  
220 that the growing demand for hi-tech elements driven by development and uptake of renewable  
221 energy technologies, will further promote research on bioleaching for metal recovery from  
222 wastes (Pollmann et al., 2018). Similarly, resource recovery will play a key role in securing the

223 future availability of critical metals, and further investment in recovery technologies that have  
224 a high potential for implementation in multi-step processes for cost-effective mineral  
225 beneficiation, is likely.

226 Further opportunities reside in fine-tuned bioleaching to enhance selectivity to targeted  
227 metals, especially critical raw materials. Extension of bioleaching methods from two steps to  
228 three and four steps could mitigate the bacteriostatic effects of waste (Valix, 2017). Another  
229 area of development is accelerated carbonation and carbon sequestration in alkaline wastes  
230 using microorganisms (Mayes et al., 2018). Recent advances in microbial electrochemical  
231 technologies for energy production metal recovery are also promising (Huang et al., 2019;  
232 Pollmann et al., 2018). Similarly, reductive bioleaching of oxidised ores and urban biomining  
233 of electronic wastes (Sethurajan et al., 2018) can promote further research and implementation  
234 of bioleaching in municipal solid waste incineration residues.

### 235 **3.4. Threats**

236 Bioleaching efficacy can be compromised by the release of potentially toxic metals from  
237 wastes, which may affect the microbes used in bioleaching. Adaptation of microorganisms is  
238 critical for higher effectiveness of this biotechnology. The current lack of work practices,  
239 especially for wastes at pilot scale, may lead to poor reproducibility of metal recoveries in  
240 different matrices. Further research is needed to satisfactory cover more substrate materials and  
241 inocula. There is also a need to assess the downstream processing in case where excessive  
242 biomass is produced in the process.

243 The major threats for the implementation of bioleaching are statutory and financial factors,  
244 such as the volatility of markets and metal prices, currently low mineral extraction costs  
245 (“mineralogical barrier”), and the current low cost of landfill disposal; all of which are not  
246 favourable to resource recovery. Finally, the possibility of biological hazards needs accurate

247 assessment at the prototype phase via standardised ecotoxicity tests that still have to be fully  
248 developed and should adapt to the proposed technologies.

249

#### 250 **4. Concluding remarks**

251 Bioleaching allows recovery of critical resources from MSWI residues, such as fly ashes  
252 and bottom ashes, and in this paper, we have identified the Strengths, Weaknesses,  
253 Opportunities, and Threats associated with this biotechnology. Bioleaching is a promising  
254 approach for the recovery of metals, in particular critical raw materials, from wastes.  
255 However, further developments are still needed to enable sustainable and commercial  
256 application to residues from municipal solid waste incineration, to enable scale-up and  
257 demonstration of the commercial value. Advances in reactor design and demonstration at  
258 commercially relevant scales are critical to the adoption of this biotechnology. Techno-  
259 economic analysis, life cycle assessment (LCA), life cycle sustainability assessment (LCSA)  
260 are tools that can be used to establish and develop the application of bioleaching for resource  
261 recovery from wastes.

262 The full implementation of new technologies for resource recovery, such as bioleaching,  
263 needs an integrated policy and regulatory framework at all levels – local, regional, national,  
264 and international, not just scientific and technical advances. The need for regulatory,  
265 economic, and fiscal instruments to enforce and incentivise resource recovery is pressing.  
266 Some of these instruments might include providing a buffer against price volatility for  
267 recovered materials and metals; supporting markets in recyclates; whilst simultaneously  
268 investing in research and development and advancing technology readiness levels. Further  
269 investment in infrastructure and supply chains to enable resource recovery is also required.

270 Finally, good coverage of testing for suitable microorganisms and heterogeneous substrate  
271 materials is essential to fill the existing knowledge gaps and feasibility uncertainties for

272 alkaline waste. Increasing environmental awareness, as well as limitations of traditional  
273 methods for complex materials with low metal content, will expand the development and  
274 application of bioleaching for primary ores. The potential amount of resources, in particular,  
275 critical raw materials such as metals, that can be recovered are relevant and can contribute to  
276 more sustainable waste management practices, whilst simultaneously avoiding unnecessary  
277 raw resource extraction and associated environmental impacts.

278

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284

## 285 **Author contributions statement**

286 **Helena I. Gomes:** Conceptualization, Writing - original draft preparation and finalisation;  
287 **Valerio Funari and Rebecca Ferrari:** Writing - Reviewing and Editing

288

## 289 **References**

290 Auerbach R, Ratering S, Bokelmann K, Gellermann C, Brämer T, Baumann R, et al.  
291 Bioleaching of valuable and hazardous metals from dry discharged incineration slag.  
292 An approach for metal recycling and pollutant elimination. *Journal of Environmental*  
293 *Management* 2019; 232: 428-437.  
294 EC. Communication from the Commission to the European Parliament, the Council, the  
295 European Economic and Social Committee and the Committee of the Regions on the  
296 2017 list of Critical Raw Materials for the EU. COM(2017) 490 final, 2017. Brussels,  
297 13.9.2017  
298 Funari V, Braga R, Bokhari SNH, Dinelli E, Meisel T. Solid residues from Italian municipal  
299 solid waste incinerators: A source for “critical” raw materials. *Waste Management*  
300 2015; 45: 206-216.  
301 Funari V, Gomes HI, Cappelletti M, Fedi S, Dinelli E, Rogerson M, et al. Optimization  
302 routes for the bioleaching of MSWI fly and bottom ashes using microorganisms

303 collected from a natural system. *Waste and Biomass Valorization* 2019, 10(12), 3833-  
304 3842

305 Funari V, Makinen J, Salminen J, Braga R, Dinelli E, Revitzer H. Metal removal from  
306 Municipal Solid Waste Incineration fly ash: A comparison between chemical leaching  
307 and bioleaching. *Waste Manag* 2017; 60: 397-406.

308 Gomes HI, Funari V, Mayes WM, Rogerson M, Prior TJ. Recovery of Al, Cr and V from  
309 steel slag by bioleaching: Batch and column experiments. *Journal of Environmental*  
310 *Management* 2018; 222: 30-36.

311 Hofmann M, Hofmann H, Hagelüken C, Hool A. Critical raw materials: A perspective from  
312 the materials science community. *Sustainable Materials and Technologies* 2018; 17:  
313 e00074.

314 Huang T, Wei X, Zhang S. Bioleaching of copper sulfide minerals assisted by microbial fuel  
315 cells. *Bioresource Technology* 2019; 288: 121561.

316 Işıldar A, van Hullebusch ED, Lenz M, Du Laing G, Marra A, Cesaro A, et al.  
317 Biotechnological strategies for the recovery of valuable and critical raw materials  
318 from waste electrical and electronic equipment (WEEE)—A review. *Journal of*  
319 *Hazardous Materials* 2019; 362: 467-481.

320 Joseph A, Snellings R, Van den Heede P, Matthys S, De Belie N. The use of municipal solid  
321 waste incineration ash in various building materials: A Belgian point of view.  
322 *Materials* 2018; 11: 141.

323 Kirkelund GM, Magro C, Guedes P, Jensen PE, Ribeiro AB, Ottosen LM. Electrodialytic  
324 removal of heavy metals and chloride from municipal solid waste incineration fly ash  
325 and air pollution control residue in suspension – test of a new two compartment  
326 experimental cell. *Electrochimica Acta* 2015; 181: 73-81.

327 Latorre M, Cortés MP, Travisany D, Di Genova A, Budinich M, Reyes-Jara A, et al. The  
328 bioleaching potential of a bacterial consortium. *Bioresource Technology* 2016; 218:  
329 659-666.

330 Luo H, Cheng Y, He D, Yang E-H. Review of leaching behavior of municipal solid waste  
331 incineration (MSWI) ash. *Science of The Total Environment* 2019; 668: 90-103.

332 Mäkinen J, Salo M, Soini J, Kinnunen P. Laboratory scale investigations on heap (bio)  
333 leaching of municipal solid waste incineration bottom ash. *Minerals* 2019; 9: 290.

334 Mayes WM, Riley AL, Gomes HI, Brabham P, Hamlyn J, Pullin H, et al. Atmospheric CO<sub>2</sub>  
335 sequestration in iron and steel slag: Consett, Co. Durham, UK. *Environmental Science*  
336 *& Technology* 2018; 52: 7892-7900.

337 Meer I, Nazir R. Removal techniques for heavy metals from fly ash. *Journal of Material*  
338 *Cycles and Waste Management* 2018; 20: 703-722.

339 Miller MG. Environmental Metabolomics: A SWOT Analysis (Strengths, Weaknesses,  
340 Opportunities, and Threats). *Journal of Proteome Research* 2007; 6: 540-545.

341 Oberle B, Bringezu S, Hatfield-Dodds S, Hellweg S, Schandl H, Clement J, et al. *Global*  
342 *Resources Outlook 2019: Natural Resources for the Future We Want. A Report of the*  
343 *International Resource Panel. United Nations Environment Programme. Nairobi,*  
344 *Kenya 2019.*

345 Pollmann K, Kutschke S, Matys S, Raff J, Hlawacek G, Lederer FL. Bio-recycling of metals:  
346 Recycling of technical products using biological applications. *Biotechnology*  
347 *Advances* 2018; 36: 1048-1062.

348 Purnell P. On a voyage of recovery: a review of the UK's resource recovery from waste  
349 infrastructure. *Sustainable and Resilient Infrastructure* 2019; 4: 1-20.

350 Quina MJ, Bontempi E, Bogush A, Schlumberger S, Weibel G, Braga R, et al. Technologies  
351 for the management of MSW incineration ashes from gas cleaning: New perspectives

352 on recovery of secondary raw materials and circular economy. *Science of The Total*  
353 *Environment* 2018; 635: 526-542.

354 Ramanathan T, Ting Y-P. Alkaline bioleaching of municipal solid waste incineration fly ash  
355 by autochthonous extremophiles. *Chemosphere* 2016; 160: 54-61.

356 Sethurajan M, van Hullebusch ED, Nancharaiah YV. Biotechnology in the management and  
357 resource recovery from metal bearing solid wastes: Recent advances. *Journal of*  
358 *Environmental Management* 2018; 211: 138-153.

359 Simon F-G, Holm O. Resources from recycling and urban mining: Limits and prospects.  
360 2018.

361 Srichandan H, Mohapatra RK, Parhi PK, Mishra S. Bioleaching approach for extraction of  
362 metal values from secondary solid wastes: A critical review. *Hydrometallurgy* 2019;  
363 189: 105122.

364 Tang J, Su M, Wu Q, Wei L, Wang N, Xiao E, et al. Highly efficient recovery and clean-up  
365 of four heavy metals from MSWI fly ash by integrating leaching, selective extraction  
366 and adsorption. *Journal of Cleaner Production* 2019; 234: 139-149.

367 Tang J, Su M, Zhang H, Xiao T, Liu Y, Liu Y, et al. Assessment of copper and zinc recovery  
368 from MSWI fly ash in Guangzhou based on a hydrometallurgical process. *Waste*  
369 *Management* 2018; 76: 225-233.

370 Valix M. 18 - Bioleaching of Electronic Waste: Milestones and Challenges. In: Wong JWC,  
371 Tyagi RD, Pandey A, editors. *Current Developments in Biotechnology and*  
372 *Bioengineering*. Elsevier, 2017, pp. 407-442.

373 Velenturf APM, Archer SA, Gomes HI, Christgen B, Lag-Brotons AJ, Purnell P. Circular  
374 economy and the matter of integrated resources. *Science of The Total Environment*  
375 2019; 689: 963-969.

376 Xu T-J, Ramanathan T, Ting Y-P. Bioleaching of incineration fly ash by *Aspergillus niger* –  
377 precipitation of metallic salt crystals and morphological alteration of the fungus.  
378 *Biotechnology Reports* 2014; 3: 8-14.

379